



**CHEMICAL, BIOLOGICAL, AND PHYSICAL
CHARACTERIZATION
OF THE CHURCH CREEK AND
PAROLE PLAZA NPDES MONITORING STATIONS: 2016 -
2017**



FINAL REPORT

Prepared for

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February 8, 2018

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1 INTRODUCTION

In 1998, Anne Arundel County began implementing a long-term monitoring program that satisfies requirements for its Countywide National Pollutant Discharge Elimination System (NPDES) Municipal Separate Storm Sewer System Discharge Permit. Monitoring has continued to be required as part of the terms of each renewed permit. Currently, monitoring is required to satisfy conditions outlined in Section F: Assessment of Controls of the County's new permit issued in February 2014. The monitoring program includes chemical, biological, and physical monitoring in the Church Creek subwatershed located within the larger South River watershed. This document describes the monitoring effort undertaken from July 2016 through June 2017.

Biological and physical monitoring take place at monumented locations along the study reach, as described in more detail below. The chemical monitoring activities take place at two stations in the Church Creek subwatershed:

- Downstream of two high-imperviousness, commercial land use outfalls, called the Parole Plaza monitoring station
- An instream station downstream of the Route 2 culvert, called the Church Creek monitoring station

The basic permit requirements for storm event monitoring include sampling a target of 12 storms per year (three in each quarter) that are characterized by three representative (rising, peak, and falling limbs of the hydrograph) discrete samples per storm event, the collection of baseflow samples during extended dry periods, laboratory analysis of water quality parameters specified in the permit, biological and physical characterizations of the study reach, and continuous flow monitoring.

The County is interested in determining the extent to which the redevelopment of the Parole Plaza site (now known as the Annapolis Towne Centre at Parole) has affected the quality of the stormwater effluent from the site. Construction began in 2004, and the bulk of the site work was completed by late 2008. Stream restoration construction took place on a portion of Church Creek, as well as a tributary, from late 2015 into early 2016. This restoration included reengineering of stream channel and resulted in minor changes in the profile of the stream from 2015 to 2016.

During 2016, the South River Federation, in cooperation with Anne Arundel County, undertook restoration of a portion of Church Creek behind the Annapolis Harbor Center and nearby the County's existing biological and physical monitoring sites. This work consisted of 1,500 linear feet of stream restoration and implementation of step-pool storm conveyance, riffle weirs, and grade control structures to improve habitat and increase floodplain connectivity. The County's existing biological and physical monitoring locations downstream of this restoration will be useful in assessing the effects of this work.

2 METHODS

2.1 CHEMICAL MONITORING

During the 2017 sampling period, July 2016 through June 2017, twelve storm events were sampled and analyzed. This section describes the equipment and techniques used in this sampling program. It includes discussions of sample collection, sample analysis, flow data collection, and basin rainfall characterization. A summary of maintenance activities is also included. Data and quarterly data reports (Versar 2017a, 2017b, and 2017c) were used to prepare this annual summary report.

2.1.1 Monitoring Sites

The long-term chemical monitoring program is performed at one outfall station, Parole Plaza, and one instream station, Church Creek. The two stations are described below:

Parole Monitoring Station. This station is a restoration station located at the head of the Parole Tributary to Church Creek. There are two outfalls draining to the sampling station. The first is a 60” corrugated metal pipe (CMP) that has been the historical sampling location for the monitoring of this station. The second is a 54” reinforced concrete pipe (RCP) that was connected to the drainage network during the summer of 2007.

Church Creek Monitoring Station. This station is an instream station on the mainstem of Church Creek. It is located approximately 500 feet downstream of the confluence of the tributary that carries the runoff from the Parole Plaza monitoring station. The samples are collected in the 96” CMP culvert that carries Church Creek underneath Maryland State Highway 2 (Solomons Island Road). The bottom of this culvert is lined with concrete that extends 1.8 feet in height up the sides of the corrugated metal culvert.

Location information and land use data were taken from the *Annapolis Towne Centre @ Parole Stormwater Management Report* (Greenhorne & O’Mara 2005), and summarized for each site in Tables 2-1 and 2-2.

| Table 2-1. Drainage areas and site locations of monitoring stations in Church Creek watershed | | | |
|---|---------------------|---|--------------|
| Monitoring Station | Station Type | Location | Area (acres) |
| Parole Plaza | Restoration/Outfall | Southwest corner of Forest Drive and MD State Highway 2 | 60.41 |
| Church Creek | Instream | Downstream (east) of MD State Highway 2 | 279.09 |

Table 2-2. Land use summary for the monitoring stations in the Church Creek subwatershed

| Land Use | Land Use Area (acres) | | Percent of Total Acreage | |
|--------------|-----------------------|---------------|--------------------------|--------------|
| | Parole Plaza | Church Creek | Parole Plaza | Church Creek |
| Impervious | 52.81 | 191.37 | 87.4 | 68.6 |
| Open Space | 7.60 | 87.72 | 12.6 | 31.4 |
| TOTAL | 60.41 | 279.09 | 100 | 100 |

2.1.2 Water Sample Collection and Data Analysis

The sample period for this reporting cycle extended from July 2016 through June 2017. Samples are analyzed for the presence of the pollutants listed in Table 2-3.

Table 2-3. Analytes, detection limits, and analytical methods for the Church Creek and Parole Plaza Monitoring stations

| Parameter | Detection Limit (mg/L) | Analytical Method |
|-----------------------------------|------------------------|-------------------|
| Biochemical Oxygen Demand (5 Day) | 2.0 | SM 5210 B-01 |
| Total Kjeldahl Nitrogen | 0.5 | SM 4500-NH3 C97 |
| Nitrate + Nitrite | 0.05 | SM 4500-NO3 H00 |
| Total Phosphorus | 0.01 | SM 4500-P E99 |
| Total Suspended Solids | 1.0 | SM 2540 D-97 |
| Total Copper (µg/L) | 2.0 | EPA 200.8 |
| Total Lead (µg/L) | 2.0 | EPA 200.8 |
| Total Zinc (µg/L) | 20.0 | EPA 200.8 |
| Total Petroleum Hydrocarbons | 5.0 | EPA 1664 |
| <i>E. coli</i> (MPN/100 ml) | 10.0 | SM 9223 B |
| Hardness | 1.0 | SM 2340 C |

During the sampling period, twelve storm samples were collected; no baseflow samples were taken in lieu of storm samples.

Below is a discussion of the storm events that were sampled during the monitoring period. Additional discussion of each event can be found in Appendix A.

- July 28, 2016 Storm - The total rainfall for this event was 0.34” and lasted approximately 3.0 hours, based on data from the Church Creek rain gauge.
- September 19, 2016 Storm - The total rainfall for this event was 1.22” and lasted 3.0 hours, based on data from the Church Creek rain gauge.

- September 27, 2016 Storm - The total rainfall for this event was 1.79” and lasted 4.5 hours, based on data from the Church Creek rain gauge.
- November 9, 2016 Storm - The total rainfall for this event was 0.11” and lasted 7.0 hours, based on data from the Church Creek rain gauge.
- November 29, 2016 Storm - The total rainfall for this event was 0.08” and lasted approximately 7.0 hours, based on data from the Church Creek rain gauge. Because October and November were very dry months with a total rainfall of only 2.8 inches, this storm was accepted by Anne Arundel County though it did not meet the 0.1” minimum rainfall requirement.
- December 6, 2016 Storm - The total rainfall for this event was 0.78” and lasted 12 hours, based on data from the Church Creek rain gauge.
- January 23, 2017 Storm - The total rainfall for this event was 0.49” and lasted 12.0 hours, based on data from the Church Creek rain gauge.
- February 28, 2017 Storm - The total rainfall for this event was 0.33” and lasted approximately 9.0 hours, based on data from the Church Creek rain gauge.
- March 31, 2017 Storm - The total rainfall for this event was 1.12” and lasted 10 hours, based on data from the Church Creek rain gauge.
- April 25, 2017 - The total rainfall for this event was 0.39” and lasted approximately 5.0 hours, based on data from the Church Creek rain gauge.
- May 12, 2017 - The total rainfall for this event was 0.98” and lasted approximately 21.0 hours, based on data from the Church Creek rain gauge.
- June 19, 2017 - The total rainfall for this event was 0.36” and lasted approximately 4.0 hours, based on data from the Church Creek rain gauge.

Approximately 30.44 inches of precipitation was recorded at the Church Creek station during the 2017 reporting period. Rainfall was measured using a tipping bucket rain gage located at the Church Creek station. Due to branches hanging over the rain gauge, potentially interfering with accurate rainfall data collection and clogging the rain gauge, some rain data are missing from June 28, 2016 through July 13, 2016.

Table 2-4 lists the total rainfall for each sampled event. Hydrographs are provided in Appendix A. These data, along with stream level readings collected at 5 minute intervals from a permanently mounted pressure transducer, were logged into an ISCO 6712FR automated sampler.

| Table 2-4. Rainfall data for sampled storm events | |
|---|--------------------------|
| Date | Rainfall (inches) |
| 28 July 2016 | 0.34 |
| 19 September 2016 | 1.22 |
| 27 September 2016 | 1.79 |
| 9 November 2016 | 0.11 |
| 29 November 2016 | 0.08 |
| 6 December 2016 | 0.78 |
| 23 January 2017 | 0.49 |
| 28 February 2017 | 0.33 |
| 31 March 2017 | 1.12 |
| 25 April 2017 | 0.39 |
| 12 May 2017 | 0.98 |
| 19 June 2017 | 0.36 |

The ISCO sampler located at the Church Creek station is configured to hold 24 one-liter polyethylene bottles, and can be used to collect samples directly from the 96" CMP. However, this station is generally manned for the entire duration of each event. Therefore, samples are typically taken as grabs from the culvert outfall. Total petroleum hydrocarbon and *E. coli* samples are always collected as manual grab samples. The grab sample location is approximately six feet downstream of the intake for the automated sampler and therefore is considered effectively the same sampling location as for the other parameters using the automated sampler.

When the 54" RCP was put in service at Parole Plaza in the summer of 2007, portions of the drainage that had historically been passing through the 60" CMP began flowing through the new pipe. In order to maintain consistency in the characterization of the watershed, it was determined that samples were required from both pipes. Pressure transducers were permanently mounted in the 60" CMP and 54" RCP. These measured water depth at 5-minute intervals, and stored data for up to three months. Data were downloaded bi-weekly. Stage/discharge relationships were developed for each pipe, to determine the discharge from the pipes based on depth measurements from the pressure transducer. The relationships are based on a combination of field measurements and extrapolated values. The extrapolation is necessary to characterize major storm events where directly measured values are not currently available. The rating tables are included in Appendix A.

A spreadsheet was developed to allow the field sampling crews to input field-measured level data. The spreadsheet interpolated the corresponding flow from the rating curves developed as described above. The flows from the 60" CMP and the 54" RCP were totaled and the resulting combined hydrograph for each event was plotted real-time. This method allowed the field crews to determine when the rising, peak, and falling limbs for the combined hydrograph occurred. The spreadsheet also calculated the percentage of the combined flow that each pipe was contributing. Using volumetric containers, the sampling team prepared composite samples using these per-

centages, and distributed them to the sample containers. Because of a strong sewage smell during several storm events, the field crew has been sampling *E. coli* separately from each pipe at Parole Plaza. A Technical Memorandum describing the composite sampling procedures in detail was submitted to the Maryland Department of the Environment in May 2008, and is included in Appendix A.

Water quality instruments for measuring pH, temperature, and conductivity were used at both stations. At Parole, an In-Situ Troll 9500 unit mounted within each pipe was used to obtain measurements during storm events; providing measurements every 5 minutes. Measurements for these parameters were not available when personnel were not present due to the low flow conditions at this station. Permanently installed probes would likely dry out and need to be replaced often, thus these units are engaged only during storm events. At the Church Creek station, a YSI 600 XL multiparameter sonde was permanently mounted within the culvert and was connected directly to the ISCO automated sampler, providing measurements every 5 minutes. This unit operates continuously.

Samples were distributed into bottles provided by Martel Laboratories JDS, Inc. From the beginning of the sampling until the March 31, 2017 storm event *E. coli* samples were delivered to Chesapeake Environmental Lab for processing within six hours of being collected. Beginning with the April 25, 2017 storm event, *E. coli* samples were delivered to Water Testing Labs of Maryland. All other samples were delivered to Martel Laboratories within 48 hours.

Event Mean Concentrations (EMCs) for each parameter were calculated for each storm and applied to total stormflow discharges to calculate stormflow pollutant loads for each site. An EMC is a statistical parameter used to represent the flow-weighted average concentration of a given parameter during a storm event (USEPA 2002). The EMC for a storm event where discrete samples have been collected (i.e., samples collected during the rise, peak, and falling limb of a storm event), was calculated using the following formula:

$$EMC = \frac{\sum_{i=1}^n V_i C_i}{\sum_{i=1}^n V_i}$$

where,

- V: volume of flow during period *i*, which is determined from the interval associated with the samples collected during each limb
- C: analytical result associated with period *i*
- n: total number of limbs taken during event

The stormflow pollutant load for each parameter was calculated as:

$$Load = EMC_j V_j$$

where,

- V: total volume of flow during period *j* (entire storm event).

Average annual EMCs were calculated by taking the arithmetic average of EMCs calculated when non-detects were set to zero and when non-detects were set to the detection limit. Since the true concentration of non-detect samples falls somewhere between the detection limit and the null value, this calculation represents a more accurate estimate than using EMCs with non-detects set to either zero or the detection limit. Seasonal loads (also referred to as quarterly loads) for monitored events were calculated by summing all monitored event loads for a specific season. Total seasonal loads were calculated by multiplying the average seasonal EMC by the total volume for the season. Annual loads were calculated by summing all seasonal loads.

2.1.3 Monitoring Station Maintenance

Maintenance was conducted at each sampling station on a biweekly basis. Maintenance included calibration of all probes, inspection of the sampling equipment, intake lines, and programming, and an overall cleaning and organization of the stations. A few issues concerning the replacement of monitoring equipment and the loss of data occurred during the monitoring period; below is a summary of these issues:

- On July 7, Anne Arundel County notified Versar that branches were hanging over the rain gauge, potentially interfering with accurate rainfall data collection. During the next maintenance trip, field staff clipped the branches, and cleared the rain gauge of any debris. To avoid future clogging, on September 2, Versar field staff cut the branches that were hanging over the rain gauge. Before sampling the storm on July 28, field staff noticed the submerged probe was not accurately recording level at the Church Creek station, so it was adjusted for the storm. After the storm, the level continued to drift higher; therefore, on August 18, an Onset HOBO stage logger was installed in the stream to replace the malfunctioning Teledyne ISCO submerged probe. On September 13, the HOBO logger was removed, and a working Teledyne ISCO submerged probe was installed and reattached to the 720 module. Field staff verified that the ISCO was accurately recording level before leaving the site. Staff noted that the ISCO data had begun to drift on July 22, so Versar's data analyst corrected the data by factoring out the drift using linear regressions. The HOBO logger data were post-processed and substituted into the Church Creek Continuous Flow Data_07-09_16 Excel spreadsheet during the period of August 18, 2016 at 5 p.m. until September 13, 2016 at 1 p.m.
- On November 21, field staff replaced the YSI 600XL sonde at Church Creek with a spare provided by Anne Arundel County. All probes subsequently properly recorded water quality except for conductivity, which was inadvertently set by field staff to record specific conductivity. During the routine maintenance on November 28 at 3:00 p.m., the field team corrected the sonde settings to record conductivity. The specific conductivity readings are highlighted in the "Church Creek Continuous Flow Data_10-12_16" file. Versar will convert specific conductivity to conductivity and provide revised data in an addendum. Note that in the continuous flow data file, on December 30, higher ambient temperature readings caused ice and snow melt which

resulted in a period of higher flow and conductivity readings. At Parole Plaza after the storm on November 9, the Global Water loggers were adjusted to record data using a time stamp set to daylight savings time.

- During the routine maintenance at Church Creek on February 7, field staff noticed erroneous temperature readings (e.g., 101 degrees Fahrenheit) recorded by the YSI 600XL sonde. The problematic temperature data likely affected conductivity values. The temperature/conductivity probe was removed immediately and Anne Arundel County was informed that a new probe was needed. Also, during that visit, field staff removed leaves, trash, and debris from the sonde and level logger area of the stream. On February 23, staff installed the new temperature/conductivity probe and confirmed that all probes properly logged water quality. After reviewing the data, staff concluded that high conductivity readings did not result from salted road runoff, but from the malfunctioning probe. Erroneous conductivity measurements were removed from the data file where Versar determined the problem began. The incorrect data included the storm event on January 23, so there are no temperature and conductivity data for this storm on the EMC spreadsheet.
- The CMP Global Water logger was calibrated on May 24, 2017 because the level on the logger was not reading accurately with the actual measurement taken by the field staff. During the routine maintenance on June 8 at Parole Plaza both of the Global Water loggers were adjusted to record data using a time stamp set to daylight savings time.

2.2 BIOLOGICAL MONITORING

All biological assessment data were collected in accordance with the Anne Arundel County Biological Monitoring and Assessment Program: Quality Assurance Project Plan (Anne Arundel County 2010), which incorporates many elements of Maryland Department of Natural Resources' Maryland Biological Stream Survey (MBSS). Geomorphic assessment data were collected in accordance with the standard operating procedures (SOPs) approved for the County's NPDES Program. All methods are consistent with previous years' methods (with applicable updates) to ensure data comparability between years. Collection methods are summarized below. Field data were collected in 2017 by Versar, Inc., a consultant to Anne Arundel County.

2.2.1 Sampling Locations

The study area is located in the northern portion of the Church Creek subwatershed, within the larger South River watershed in Anne Arundel County, Maryland (Figure 2-1). A total of four 75-meter biological monitoring sites are positioned along the study reach and are monitored annually. Three sites were established and first monitored in 2006; one site is located on the Parole Plaza Tributary just below Forest Drive, and two sites are located along the Church

Creek mainstem, on either side of Solomons Island Road (Maryland State Highway 2). A fourth site, located just upstream of the confluence with the Parole Plaza Tributary, was added in 2007 to monitor the effects of runoff from the Festival at Riva shopping center.

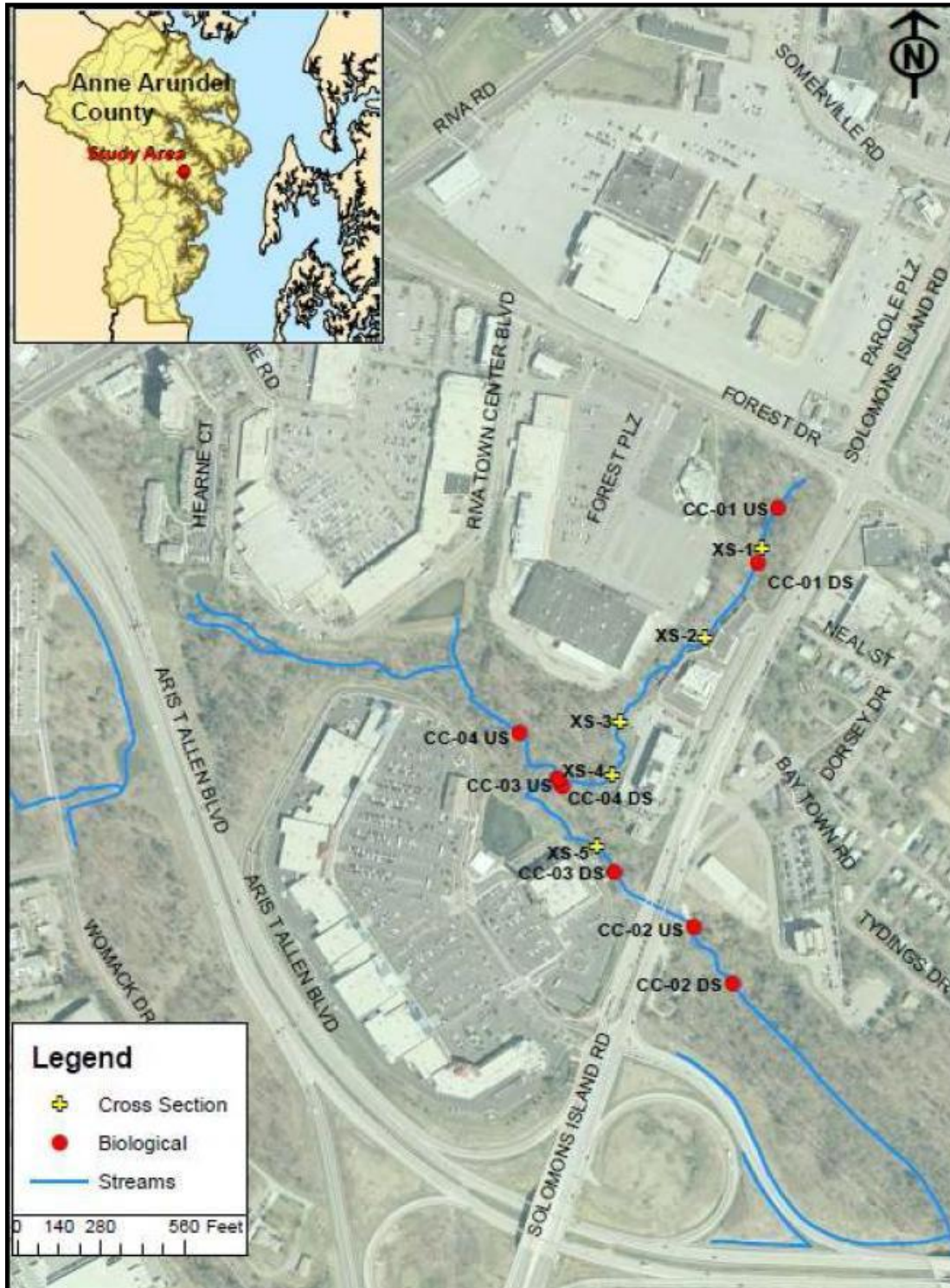


Figure 2-1. Church Creek study area and stream monitoring locations

2.2.2 Stream Habitat Evaluation

To support the biological monitoring, a visual assessment of physical habitat was completed at each monitoring site to evaluate the reach’s ability to support aquatic life. Both the MBSS Physical Habitat Index (PHI; Paul et al. 2003) and the U.S. Environmental Protection Agency (U.S. EPA) Rapid Bioassessment Protocol (RBP) habitat assessment for low gradient streams (Barbour et al. 1999) were used to visually assess the physical habitat at each site in conjunction with the spring benthic monitoring. Both habitat assessments consist of a review of biologically significant habitat parameters that evaluate a stream’s ability to support an acceptable level of biological health.

To calculate PHI at each site, six parameters were given a numerical score and a categorical rating: instream habitat, epibenthic substrate, remoteness, instream woody debris and rootwads, shading, and bank stability. The raw scores are then transformed into a scaled score (0-100 scale) as described in Paul et al. (2003), and the six scaled scores are averaged into an aggregate final PHI score. Narrative condition descriptions and scoring ranges for the PHI are displayed in Table 2-5.

| Score | Narrative |
|--------------|--------------------|
| 81-100 | Minimally Degraded |
| 66-80.9 | Partially Degraded |
| 51-65.9 | Degraded |
| 0-50.9 | Severely Degraded |

The Rapid Bioassessment Protocol (RBP) habitat assessment consists of a review of ten biologically significant habitat parameters that assess a stream’s ability to support an acceptable level of biological health: Epifaunal substrate/available cover, Embeddedness, Velocity/depth regime, Sediment deposition, Channel flow status, Channel alteration, Frequency of riffles/bends, Bank stability, Vegetative protection, and Riparian vegetative zone width. In the field, each parameter was given a numerical score from 0-20 (20=best, 0=worst), or 0-10 (10=best, 0=worst) for individual bank parameters, and a categorical rating of optimal, suboptimal, marginal or poor (Barbour et al. 1999). As overall habitat quality increases, the total score for each site typically increases. The individual RBP habitat parameters for each reach were summed to obtain an overall RBP assessment score. Because adequate reference conditions currently do not exist for Anne Arundel County, the percent comparability was calculated based on western coastal plain reference site conditions obtained from work done in Prince George’s County streams (Stribling et al. 1999). The percent of reference score, or percent comparability score, was then used to place each site into corresponding narrative rating categories. The ranges are shown in Table 2-6.

| Percent of Reference Score | Narrative |
|----------------------------|-------------------------|
| 90 - 100 | Comparable to Reference |
| 75.1 - 89.9 | Supporting |
| 60.1 - 75 | Partially Supporting |
| 0 - 60 | Non-Supporting |

2.2.3 Water Quality Measurement

In situ water quality was measured at each site with a YSI 6820 multiparameter water quality sonde. Turbidity was measured once at the upstream end of the site, all other parameters were measured from three locations within each sampling reach (upstream end, mid-point, and downstream end) and results were averaged to minimize variability and better represent water quality conditions throughout the entire sampling reach. Data were compared to the standards listed in the Code of Maryland Regulations (COMAR) 26.08.02.03-3 – Water Quality (MDE 2010) and shown in Table 2-7.

| Parameter | Standard |
|--|---|
| pH | 6.5 to 8.5 |
| Dissolved Oxygen (mg/L) | Minimum of 5 mg/L |
| Conductivity (µS/cm) | No existing standard |
| Turbidity (NTU) | Maximum of 150 NTU and maximum monthly average of 50 NTU |
| Temperature (°C) | Maximum of 32 °C (90 °F) or ambient temperature, whichever is greater |
| Source: Code of Maryland Regulations (COMAR) 26.08.02.03-3-Water Quality | |

2.2.4 Biological Sample Collection

Benthic macroinvertebrate samples were collected in April 2017 following the MBSS Spring index period protocols (MD DNR 2017) and as specified in *Anne Arundel County Biological Monitoring and Assessment Program: Quality Assurance Project Plan (QAPP; Anne Arundel County 2010)*. This methodology emphasizes the community composition and relative abundance of benthic macroinvertebrates inhabiting the most taxonomically diverse, or productive, instream habitats. In this sampling approach, a total of twenty jabs are distributed among the most productive habitats present within the 75-meter reach and sampled in proportion to their occurrence within the segment. The most productive stream habitats are riffles followed by rootwads, rootmats and woody debris and associated snag habitat; leaf packs; submerged macrophytes and associated substrate; and undercut banks. Other less preferred habitats include gravel, broken peat, clay lumps and detrital or sand areas in runs; however, of the aforementioned habitat types, those that are located within moving water are preferred over those in still water.

2.2.5 Biological Sample Processing and Identification

Benthic macroinvertebrate samples were processed and subsampled according to Maryland Biological Stream Survey methods described in the MBSS laboratory methods manual (Boward and Freidman, 2000) and as briefly summarized in the Anne Arundel County Biological Monitoring and Assessment Program: Quality Assurance Project Plan (Anne Arundel County 2010). Subsampling is conducted to standardize the sample size and reduce variation caused by field collection methods. In brief, the sample was washed of preservative in a 0.595 mm screen and spread evenly across a tray comprised of 100 numbered 5cm x 5cm grids. A random number between one and 100 was selected and the selected grid was picked entirely of macroinvertebrates under a bright light source. This process was repeated until a count of 120 was reached. The 120 organism target was used following MBSS methods to allow for specimens that are missing parts or are early instars, which cannot be properly identified.

The samples were taxonomically identified by Versar taxonomists certified by the Society for Freshwater Science (SFS) (formerly known as the North American Benthological Society, NABS). The taxonomic hierarchical level for most organisms was genus level when possible with the exception of Oligochaeta, which were identified to the family level. Early instars or damaged specimens were identified to the lowest possible level. Oligochaeta and Chironomidae specimens were permanently slide mounted for identification. Counts and identifications were recorded on a laboratory bench sheet and entered into a master database for analysis. A list of all taxa identified is provided in Appendix B: Master Taxa List.

2.2.6 Biological Data Analysis

Benthic macroinvertebrate data were analyzed using methods developed by MBSS as outlined in the *New Biological Indicators to Better Assess the Condition of Maryland Streams* (Southerland et al. 2005). The Benthic Index of Biotic Integrity (BIBI) approach involves statistical analysis using metrics that have a predictable response to water quality and/or habitat impairment. The metrics selected fall into five major groups including taxa richness, composition measures, tolerance to perturbation, trophic classification, and habit measures. Tolerance values were obtained from Bressler et al. (2005).

Raw values from each metric are given a score of 1, 3, or 5 based on ranges of values developed for each metric. The results are combined into a scaled BIBI score ranging from 1.0 to 5.0 and a corresponding narrative rating is assigned. Table 2-8 shows the thresholds for the determination of the metric scoring. Three sets of metric calculations have been developed for Maryland streams based on broad physiographic regions: Coastal Plain, Piedmont and Combined Highlands. The Coastal Plain and Piedmont regions are divided by the Fall Line. The current study area is located within the Coastal Plain region. The metrics calculated for Coastal Plain streams are as follows:

Total Number of Taxa – Equals the richness of the community in terms of the total number of genera at the genus level or higher. A large variety of genera typically indicate better overall water quality, habitat diversity and/or suitability, and community health.

Number of EPT Taxa – Equals the richness of genera within the Ephemeroptera (mayflies), Plecoptera (stoneflies), and Trichoptera (caddisflies). EPT taxa are generally considered pollution sensitive, thus higher levels of EPT taxa would be indicative of higher water quality.

Number of Ephemeroptera Taxa – Equals the total number Ephemeroptera Taxa in the sample. Ephemeroptera are generally considered pollution sensitive, thus communities dominated by Ephemeroptera usually indicate lower disturbances in water quality.

Percent Intolerant Urban – Percentage of sample considered intolerant to urbanization. Equals the percentage of individuals in the sample with a tolerance value of 0-3 out of 10. As impairment increases the percent of intolerant taxa decreases.

Percent Ephemeroptera – Equals the percent of Ephemeroptera individuals in the sample. Ephemeroptera are generally considered pollution sensitive, thus communities dominated by Ephemeroptera usually indicate lower disturbances in water quality.

Number Scraper Taxa – Equals the number of scraper taxa in the sample, those taxa that scrape food from the substrate. As the levels of stressors or pollution rise there is an expected decrease in the numbers of Scraper taxa.

Percent Climbers – Equals the percentage of the total number of individuals who are adapted to living on stem type surfaces. Higher percentages of climbers typically represent a decrease in stressors and overall better water quality.

All of the metric scores are summed and then averaged to obtain the final BIBI score. Table 2-9 shows the scores and narrative rankings of the MBSS BIBI. The biological assessment results are included in Appendix C. The QA/QC information is included in Appendix D.

| Metric | Score | | |
|------------------------------|-------|----------|-------|
| | 5 | 3 | 1 |
| Total Number of Taxa | ≥ 22 | 14-21 | < 14 |
| Number of EPT Taxa | ≥ 5 | 2-4 | < 2 |
| Number of Ephemeroptera Taxa | ≥ 2 | 1.9-1.0 | < 1.0 |
| Percent Intolerant Urban | ≥ 28 | 10-27 | < 10 |
| Percent Ephemeroptera | ≥ 11 | 0.8-10.9 | < 0.8 |
| Number of Scraper Taxa | ≥ 2 | 1.9-1.0 | < 1.0 |
| Percent Climbers | ≥ 8.0 | 0.9-7.9 | < 0.9 |

| BIBI Score | Narrative Ranking | Characteristics |
|-------------------|--------------------------|---|
| 4.0-5.0 | Good | Comparable to reference conditions, stream considered to be minimally impacted, biological metrics fall within upper 50th percentile of reference site conditions. |
| 3.0-3.9 | Fair | Comparable to reference conditions, but some aspects of biological integrity may not resemble the qualities of minimally impacted streams. |
| 2.0-2.9 | Poor | Significant deviation from reference conditions, indicating some degradation. On average, biological metrics fall below the 10th percentile of reference site values. |
| 1.0-1.9 | Very Poor | Strong deviation from reference conditions, with most aspects of biological integrity not resembling the qualities of minimally impacted streams, indicating severe degradation. On average, most or all metrics fall below the 10th percentile of reference site values. |

2.3 PHYSICAL MONITORING

2.3.1 Monitoring Sites

Five cross-sections (XS), four of which were established in 2003 and one which was established in 2007, have been measured annually through 2017. Four of these cross-sections are located along the Parole Plaza Tributary, and one cross-section is located on the Church Creek mainstem, just upstream of Solomon’s Island Road (Maryland State Highway 2; Figure 2-1). Cross-section monuments, placed on each bank, consist of capped steel reinforcement bars set within six inches of the ground surface. Field data collected by Versar, Inc. during 2017 were used to prepare this annual summary report.

2.3.2 Physical Data Collection and Analysis

Geomorphic assessments include a longitudinal profile survey, cross-section surveys, and representative pebble counts. A spreadsheet tool called *The Reference Reach Spreadsheet* version 4.3L (Mecklenburg 2006) was used to facilitate data entry and analyses. This spreadsheet was used to compile, manipulate, and plot field data and to analyze dimension, profile, and channel material characteristics of the Church Creek study area.

Data from geomorphic assessments were used to determine the stream type of each reach as categorized by the Rosgen Stream Classification methodology (Rosgen 1996). In this classification methodology, streams are categorized based on their measured field values of entrenchment ratio, width/depth ratio, sinuosity, water surface slope, and channel materials according to the table in Appendix E. As illustrated in Appendix E, the Rosgen Stream Classification categorizes streams into broad stream types, which are identified by the letters Aa, A, B, C, D,

DA, E, F, and G. Table 2-10 includes general descriptions of each Rosgen stream type. A summary of the stream types identified within this study is included in Appendix F.

| Table 2-10. Rosgen stream classification types | |
|---|---|
| Channel Type | General Description |
| Aa+ | Very steep, deeply entrenched, debris transport, torrent streams. |
| A | Steep, entrenched, confined, cascading, step/pool streams. High energy/debris transport associated with depositional soils. Very stable if bedrock or boulder dominated channel. |
| B | Moderately entrenched, moderate gradient, riffle dominated channel with infrequently spaced pools. Moderate width/depth ratio. Narrow, gently sloping valleys. Very stable plan and profile. Stable banks. |
| C | Low gradient, meandering, slightly entrenched, point-bar, riffle/pool, alluvial channels with broad, well-defined floodplains. |
| D | Braided channel with longitudinal and transverse bars. Very wide channel with eroding banks. Active lateral adjustment, high bedload and bank erosion. |
| DA | Anastomosing (multiple channels) narrow and deep with extensive, well-vegetated floodplains and associated wetlands. Very gentle relief with highly variable sinuosities and width/depth ratios. Very stable streambanks. |
| E | Low gradient, Highly sinuous, riffle/pool stream with low width/depth ratio and little deposition. Very efficient and stable. High meander/width ratio. |
| F | Entrenched, meandering riffle/pool channel on low gradients with high width/depth ratio and high bank erosion rates. |
| G | Entrenched “gully” step/pool and low width/depth ratio on moderate gradients. Narrow valleys. Unstable, with grade control problems and high bank erosion rates. |
| Source: Rosgen, D. 1996. Applied River Morphology. Wildland Hydrology. Pagosa Springs, Colorado | |

The cross-section surveys were performed at the five permanent cross-section locations, and photos were taken of upstream, downstream, left bank, and right bank views at each cross-section location. Cross-section surveys consisted of measuring the topographic variability of the associated stream bed, floodplains, and terraces, including:

- Monument elevations
- Changes in topography
- Top of each channel bank
- Elevations of bankfull indicators
- Edge of water during the time of survey
- Thalweg or deepest elevation along active channel
- Depositional and erosional features within the channel

During the cross-sectional survey, the following measurements and calculations of the bankfull channel, which are critical for determining the Rosgen stream type of each reach, also were collected:

- Bankfull Width (Wbkf): the width of the channel at the elevation of bankfull discharge or at the stage that defines the bankfull channel.
- Mean Depth (dbkf): the mean depth of the bankfull channel.
- Bankfull Cross-sectional Area (Abkf): the area of the bankfull channel, estimated as the product of bankfull width and mean depth.
- Width Depth Ratio (Wbkf/dbkf): the ratio of the bankfull width versus mean depth.
- Maximum Depth (dmbkf): the maximum depth of the bankfull channel, or the difference between the thalweg elevation and the bankfull discharge elevation.
- Width of Floodprone Area (Wfpa): the width of the channel at a stage of twice the maximum depth. If the width of the floodprone area was far outside of the channel, its value was visually estimated or paced off.
- Entrenchment Ratio (ER): the ratio of the width of the floodprone area versus bankfull width.
- Sinuosity (K): ratio of the stream length versus the valley length or the valley slope divided by the channel slope. Sinuosity was visually estimated or the valley length was paced off so that an estimate could be calculated.

To quantify the distribution of channel substrate particles sizes within the study area, a modified Wolman pebble count (Rosgen 1996) was performed at each cross-section location. Reach-wide proportional counts were used. Each pebble count consists of stratifying the reach based on the frequency of channel features in that reach (e.g., riffle, run, pool, glide) and measuring 100 particles across ten transects (i.e., 10 particles in each of 10 transects). The transects are allocated across all feature types in the proportion at which they occur within the reach. The intermediate axis of each measured pebble is recorded. The goal of the pebble count is to measure 100 particles across the bankfull width of the channel and calculate the median substrate particle size (i.e., D50) of the reach. This value is used for categorizing the sites into the Rosgen Stream Classification (Rosgen 1996). If a channel was clearly a sand or silt bed channel with no distinct variation in material size, the pebble count was not performed, and the D50 was visually estimated. However, if the channel did have variation in bed material size from feature to feature, a full pebble count was performed.

2.4 LAND USE AND STORMWATER MANAGEMENT ASSESSMENT

2.4.1 Church Creek Watershed Land Use

A previous report (Versar 2013) provided information on land use, based on field reconnaissance conducted during 2013. As seen in an aerial photograph and stormwater best management practice (BMP) facilities map (Figure 2-2), the watershed is predominantly

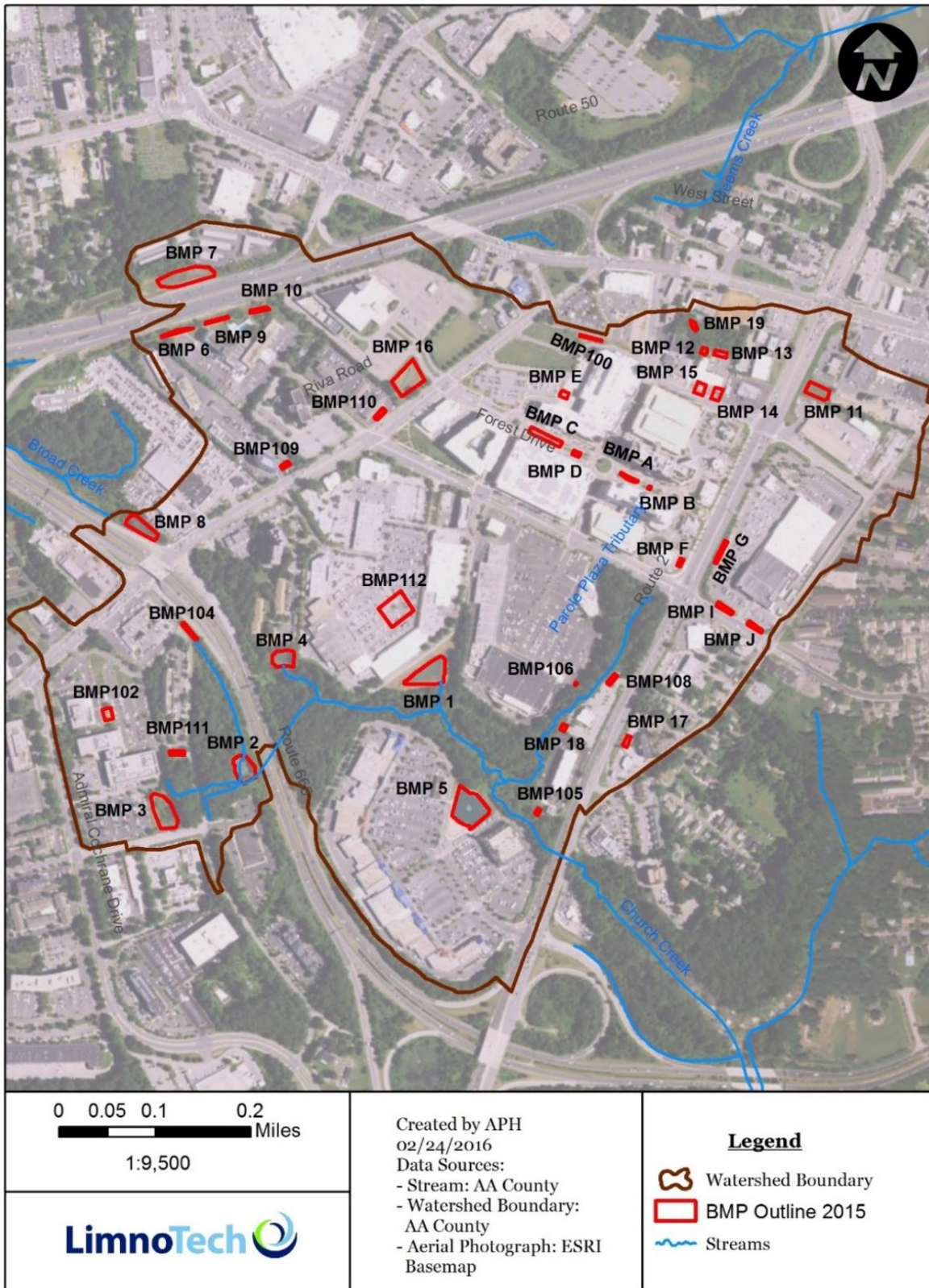


Figure 2-2. Church Creek BMPs

commercial with open space area adjacent to the stream channels. There is little available area for further development in the watershed except for areas that are being redeveloped. Anecdotal information indicates there has been no change in land use in this watershed since the 2013 land use evaluation.

2.4.2 Church Creek Watershed BMP Inspections

The Church Creek watershed contains 40 BMPs as shown in Figure 2-2 and Table 2-11. Inspections were conducted on December 9 and 12, 2015 and January 13, 2016. Inspection and maintenance information is currently being updated in the County's BMP database. Inspection reports are maintained at the County's offices. Appendices H and I contain BMP codes and acronyms.

| Table 2-11. Church Creek BMP overview | | | | | | | | |
|---------------------------------------|--|---|------------------------------|--------------------------------------|---|--|----------------|--|
| Church Creek BMP | AA County Urban BMP Database ID ^(a) | Current BMP Type in Database ^(b) | Recommended Updated BMP Type | Drainage Area (acres) ^(c) | Location | Address | Presumed Owner | County Follow-up Completed |
| 1 | AA001128 | PWED | PWED | 26.95 | Festival at Riva Shopping Center | Riva Road and Forest Drive | County | --- |
| 2 | Not found in 2015 database | N/A | DP | - | Forested Area | Between Aris T. Allen Boulevard and Womack Drive | SHA | --- |
| 3 | AA001962 | XDPD | XDPD | 1.15 | ARINC Parking Lot | Spruill Road and Admiral Cochrane Drive | County | --- |
| 4 | AA000074 | XDPD | XDPD | 4.47 | Forest Garden Apartments | 130 Hearne Court | County | --- |
| 5 | AA001042 | PWED | PWED | 18.8 | Annapolis Harbour Center | Solomons Island Road | County | --- |
| 6 | Not found in 2015 database | N/A | IBAS | - | Double Tree by Hilton Hotel Parking Lot | Route 50 | SHA | --- |
| 7 | AA001069 | XDPD | WP | 9.7 | Annapolis Self Storage | Route 50 near East Classic Court | SHA | No action taken – current location of feature within BMP inventory database is outside of Church Creek drainage area |
| 8 | Not found in 2015 database | N/A | IBAS | - | Sheehy Nissan of Annapolis | Aris T. Allen Boulevard and Riva Road | SHA | --- |

^(a) Numbering system carried over from the 2013 BMP inspection report.

^(b) The 2015 Anne Arundel County Urban BMP database was used to identify the BMP type data for the 2015 inspections (See List of Acronyms in Appendix A). Thus BMP type data may be different from the 2013 BMP type data for the same BMP. In addition, LimnoTech has recommended that some of these BMP types be changed based on what was observed in the field.

^(c) The 2015 Anne Arundel County Urban BMP database was used to update drainage areas for the 2015 inspections. Some drainage areas are missing in the Urban BMP database.

| Table 2-11. (Continued) | | | | | | | | |
|-------------------------|--|---|------------------------------|--------------------------------------|---|---------------------------|-------------------|--|
| Church Creek BMP | AA County Urban BMP Database ID ^(a) | Current BMP Type in Database ^(b) | Recommended Updated BMP Type | Drainage Area (acres) ^(c) | Location | Address | Presumed Owner | County Follow-up Completed |
| 9 | Not found in 2015 database | N/A | IBAS | - | Double Tree by Hilton Hotel Parking Lot | Route 50 | SHA | --- |
| 10 | Not found in 2015 database | N/A | IBAS | - | Double Tree by Hilton Hotel Parking Lot | Route 50 | SHA | --- |
| 11 | AA001446 | PWED | XDPD | 1.13 | Second National Federal Savings Bank | 2045 West Street | City of Annapolis | --- |
| 12 | AA012015 | MSGW | MSGW | 1.94 | AAA Mid Atlantic Car Care | 2054 Somerville Road | County | --- |
| 13 | AA012014 | APRP | APRP | 0.03 | AAA Mid Atlantic Car Care | 2054 Somerville Road | County | --- |
| 14 | AA012013 | SPSC | MMBR | 0.26 | AAA Mid Atlantic Car Care | 2054 Somerville Road | County | --- |
| 15 | AA012012 | SPSC | MMBR | 0.21 | AAA Mid Atlantic Car Care | 2054 Somerville Road | County | --- |
| 16 | AA000071 | XDED | XDED | 3.71 | Nationwide Insurance | 2453-2499 Riva Road | County | --- |
| 17 | AA006493 | ITRN | ITRN | 1.00 | Annapolis Station | 2431 Solomons Island Road | County | |
| 18 | AA001872 | ITRN | ITRN | - | Two Restaurant Sites | 2436 Solomons Island Road | County | County removed feature from its BMP dataset during completion of the BMP inspection report |

| Table 2-11. (Continued) | | | | | | | | |
|-------------------------|--|---|------------------------------|--------------------------------------|----------------------------------|---------------------------|----------------|--|
| Church Creek BMP | AA County Urban BMP Database ID ^(a) | Current BMP Type in Database ^(b) | Recommended Updated BMP Type | Drainage Area (acres) ^(c) | Location | Address | Presumed Owner | County Follow-up Completed |
| 19 | Not found in 2015 database | N/A | DP | - | Capitol One Bank | 2200 Somerville Road | County | |
| A | AA008471 | FSND | FBIO | 0.3 | Annapolis Towne Center at Parole | Towne Center Boulevard | County | Feature type corrected and relocated according to grading plan |
| B | AA008472 | XOGS | XOGS | 0.53 | Annapolis Towne Center at Parole | Towne Center Boulevard | County | Feature relocated according to grading plan |
| C | AA008475 | FBIO | FBIO | 0.4 | Annapolis Towne Center at Parole | Towne Center Boulevard | County | Feature relocated according to grading plan |
| D | AA008473 | SF/SFND | XOGS | 1.88 | Annapolis Towne Center at Parole | Towne Center Boulevard | County | Feature type corrected and relocated according to grading plan |
| E | AA008474 | SF/SFND | XOGS | 5.63 | Annapolis Towne Center at Parole | Towne Center Boulevard | County | Feature type corrected and feature relocated according to grading plan |
| F | AA008470 | FSND | FSND | 4.78 | Annapolis Towne Center at Parole | 2398 Solomons Island Road | County | Feature relocated according to grading plan |
| G | Not found in 2015 database | N/A | FBIO | - | Shoppers Food Warehouse | 2371 Solomons Island Road | County | --- |
| H | <i>There is no BMP H, please refer to the 2012 BMP inspection report for further explanation</i> | | | | | | | |
| I | Not found in 2015 database | N/A | FBIO | | Shoppers Food Warehouse | 2371 Solomons Island Road | County | --- |
| J | Not found in 2015 database | N/A | FBIO | | Shoppers Food Warehouse | 2371 Solomons Island Road | County | --- |
| 100 | AA006819 | NDRR | NDRR | 0.2 | Holiday Inn Express & Suites | 2451 Riva Road | Private | --- |
| 101 | AA008115 | IMPP | IMPP | - | Annapolis Towne Center | 2348 Forest Drive | Private | Not found on grading plan, BMP removed from dataset |

| Table 2-11. (Continued) | | | | | | | | |
|-------------------------|--|---|------------------------------|--------------------------------------|--|---------------------------|----------------|---|
| Church Creek BMP | AA County Urban BMP Database ID ^(a) | Current BMP Type in Database ^(b) | Recommended Updated BMP Type | Drainage Area (acres) ^(c) | Location | Address | Presumed Owner | County Follow-up Completed |
| 102 | AA031422 | SPSC | FBIO | 0.3 | ARINC | 2551 Riva | Private | --- |
| 103 | Not found in 2015 database | N/A | ITRN | - | Spruill Rd and Womack Drive Intersection | Womack Drive | Unknown | --- |
| 104 | AA010516 | ODSW | ODSW | 0.14 | Highway 665 On Ramp | 2525 Riva Road | SHA | --- |
| 105 | Not found in 2015 database | N/A | ITRN | - | Wainwright Avenue | 2441 Solomons Island Road | Private | --- |
| 106 | AA005932 | ITRN | ITRN | 6.73 | South Annapolis Forest Drive Home Depot | Forest Drive | Private | --- |
| 108 | AA001180 | ITRN | ITRN | 0.53 | Two Restaurant Site | 2436 Solomons Island Road | Private | --- |
| 109 | AA000335 | ITRN | ITRN | 9.16 | Sovran Building | - | Private | --- |
| 110 | AA000058 | XDPD | XDED | 2.07 | Parole Professional Building | 132 Holiday Court | Private | --- |
| 111 | AA002384 | ITRN | ITRN | 2.23 | Hampton Inn & Suites | 124 Womack Drive | Private | --- |
| 112 | AA002634, AA003322, AA003350, AA003388 | ITRN | ITRN | 0.5 (all) | Festival at Riva | Riva Road & Forest Drive | County | All four trenches relocated according to grading plan |

3 RESULTS

3.1 POLLUTANT CONCENTRATIONS

During this sampling period, 72 water chemistry samples were analyzed. In a few instances, analyte concentrations fell below the specified detection limits. Table 3-1 shows the percentage of samples that were below the detection limit.

| Table 3-1. The percentage of non-detects by parameter | | |
|---|------------------|-------------|
| Parameter | Detection Limit | Wet Weather |
| BOD ₅ (mg/L) | 2.0/4.0 | 31 |
| TKN (mg/L) | 0.5 | 12 |
| Nitrate + Nitrite (mg/L) | 0.05 | 0 |
| Total Phosphorus (mg/L) | 0.01 | 0 |
| TSS (mg/L) | 1.0 | 0 |
| Total Copper (µg/L) | 2.0 | 0 |
| Total Lead (µg/L) | 2.0 | 23 |
| Total Zinc (µg/L) | 20 | 0 |
| TPH (mg/L) | 5.0 | 56 |
| <i>E. coli</i> (MPN/100 ml) | 1.0, 10.0, 100.0 | 6 |
| Hardness (mg/L) | 1.0 | 0 |

Table 3-2 shows the maximum values observed for wet weather samples for both stations. The maximum value for each parameter during wet weather monitoring, station of occurrence, and storm date of observation are listed in Table 3-3. Church Creek had the highest values for seven of the thirteen parameters measured during wet weather sampling in 2017. Most of the maximum wet weather values for the parameters were measured during the June 19 storm event. The maximum *E. coli* concentration at Church Creek was 91,000 MPN/100 ml and was observed during the September 27 storm. Chemical monitoring summaries can be found in Appendix G.

| Parameter | Church Creek | Parole Plaza |
|-----------------------------|--------------|--------------|
| Water Temperature (°F) | 80.4 | 81.4 |
| pH | 7.5 | 8.92 |
| BOD ₅ (mg/L) | 23 | 25 |
| TKN (mg/L) | 4.4 | 3.0 |
| Nitrate + Nitrite (mg/L) | 1.70 | 3.0 |
| Total Phosphorus (mg/L) | 2.60 | 0.49 |
| TSS (mg/L) | 400 | 530 |
| Total Copper (µg/L) | 408 | 200 |
| Total Lead (µg/L) | 401 | 55 |
| Total Zinc (µg/L) | 3600 | 1810 |
| TPH (mg/L) | 12 | 10 |
| <i>E. coli</i> (MPN/100 ml) | 91,000 | 65,044 |
| Hardness (mg/L) | 230 | 400 |
| BDL: Below Detection Limit | | |

| Parameter | Date of Storm | Site | Maximum Value |
|-----------------------------|---------------|--------------|---------------|
| Water Temperature (°F) | 6/19/17 | Parole Plaza | 81.4 |
| pH | 12/6/16 | Parole Plaza | 8.92 |
| BOD ₅ (mg/L) | 4/25/17 | Parole Plaza | 25 |
| TKN (mg/L) | 9/27/16 | Church Creek | 4.4 |
| Nitrate + Nitrite (mg/L) | 9/27/16 | Parole Plaza | 3.0 |
| Total Phosphorus (mg/L) | 11/29/16 | Church Creek | 2.6 |
| TSS (mg/L) | 6/19/17 | Parole Plaza | 530 |
| Total Copper (µg/L) | 6/19/17 | Church Creek | 408 |
| Total Lead (µg/L) | 6/19/17 | Church Creek | 401 |
| Total Zinc (µg/L) | 6/19/17 | Church Creek | 3600 |
| TPH (mg/L) | 6/19/17 | Church Creek | 12 |
| <i>E. coli</i> (MPN/100 ml) | 9/27/16 | Church Creek | 91,000 |
| Hardness (mg/L) | 6/19/17 | Parole Plaza | 400 |

3.2 EVENT MEAN CONCENTRATIONS AND POLLUTANT LOADINGS

Flow-weighted stormflow event mean concentrations (EMCs) values are presented in Table 3-4. EMCs for BOD, TKN, total phosphorus, lead, and TPH were higher at Church Creek than at Parole Plaza.

Table 3-4. Average EMCs observed during July 2016 to June 2017

| Parameter | Church Creek | Parole Plaza |
|-----------------------------|---------------------|---------------------|
| Water Temperature (°F) | 54.3 | 57.3 |
| pH | 7.04 | 7.73 |
| BOD ₅ (mg/L) | 3 | 2 |
| TKN (mg/L) | 0.69 | 0.68 |
| Nitrate + Nitrite (mg/L) | 0.33 | 0.37 |
| Total Phosphorus (mg/L) | 0.14 | 0.12 |
| TSS (mg/L) | 38 | 47 |
| Total Copper (µg/L) | 13.1 | 19.7 |
| Total Lead (µg/L) | 5.1 | 2.2 |
| Total Zinc (µg/L) | 83 | 137 |
| TPH (mg/L) | 0.14 | 0.46 |
| <i>E. coli</i> (MPN/100 ml) | 5,597 | 7,366 |
| Hardness (mg/L) | 43 | 44 |

Summed loads for the sampled events monitored during the July 2016 to June 2017 sampling period are shown in Table 3-5. Church Creek per-acre loading rates for monitored events were higher than Parole Plaza for all parameters.

Table 3-5. Estimated pollutant loadings for observed events, in pounds, for the July 2016 to June 2017 sampling period

| Parameter | Church Creek | | Parole Plaza | |
|-------------------|---------------------|-----------------|---------------------|-----------------|
| | Total | Per Acre | Total | Per Acre |
| BOD ₅ | 1,358 | 4.87 | 111.30 | 1.84 |
| TKN | 268.71 | 1.03 | 30.86 | 0.51 |
| Nitrate + Nitrite | 124.01 | 0.44 | 15.46 | 0.26 |
| Total Phosphorus | 52.91 | 0.19 | 4.93 | 0.08 |
| TSS | 14,267.97 | 51.12 | 1,934.02 | 32.01 |
| Total Copper | 4.96 | 0.018 | 0.82 | 0.014 |
| Total Lead | 2.20 | 0.008 | 0.12 | 0.002 |
| Total Zinc | 31.35 | 0.112 | 5.71 | 0.095 |
| TPH | 979.61 | 3.51 | 108.48 | 1.80 |
| Hardness | 16,116.07 | 57.75 | 1,844.82 | 30.54 |

3.3 BIOLOGICAL ASSESSMENT

Biological and physical habitat assessments were completed on April 12 and 17, 2017. Presented below are the summary results for each assessment site. For full bioassessment data and results, refer to Appendix C. A complete taxonomic list can be found in Appendix B. QA/QC information is in Appendix D. As introduced in Section 1, the South River Federation, in cooperation with the County, undertook restoration of Church Creek in the vicinity of the existing biological and physical monitoring sites beginning in late January, 2016. This work consisted of 1,500 linear feet of stream restoration and implementation of step-pool storm conveyance, riffle weirs, and grade control structures to improve habitat and increase floodplain connectivity. All of the CC-04 and part of the CC-03 biological monitoring sites were within the restored reach of stream.

Physical habitat quality was evaluated using the MBSS PHI, and rated “Partially Degraded” for three sites and “Degraded” for one site (Table 3-6). Index scores ranged from a low of 70.0 at CC-02 to a high of 67.9 at CC-04. All sites received very low scores for remoteness due to the proximity of the stream channel to roads and development. Generally, instream woody debris scored high for all the sites. Individual parameter results are listed in Appendix C. Overall, PHI scores throughout the study area indicate habitat conditions may be limiting the potential for healthy biological communities.

The RBP was also used to evaluate the physical habitat quality and rated “Partially Supporting” for three sites and “Supporting” for one site (Table 3-6). Scores ranged from 61 at CC-02 to 78 at CC-04. Generally, epifaunal substrate/cover, velocity/depth regime, sediment deposition, and vegetative protection scored low for all the sites. Overall, RBP scores throughout the study area indicate that physical habitat conditions could limit the potential for healthy, stable biological communities, similar to what was found using the PHI.

| Site | PHI Score | PHI Narrative Rating | RBP Score | RBP Narrative Rating |
|-------|-----------|----------------------|-----------|----------------------|
| CC-01 | 67.4 | Partially Degraded | 74 | Partially Supporting |
| CC-02 | 61.0 | Degraded | 61 | Partially Supporting |
| CC-03 | 71.7 | Partially Degraded | 70 | Partially Supporting |
| CC-04 | 67.9 | Partially Degraded | 78 | Supporting |

For biological conditions, three stations received a rating of “Poor” and one station was rated as “Very Poor”, indicating a highly impaired benthic macroinvertebrate community. Low BIBI scores are driven by low metric scores at all sites for Number of EPT taxa, Number of Ephemeroptera, Percent Ephemeroptera, and Percent Intolerant Urban. The Percent Climbers metric and Number of Scrapers both received average to high scores for all sites while the Percent Intolerant to Urban metric received low scores (< 5% per subsample) at all sites. Poor

habitat conditions and marginal water quality parameters may contribute to low BIBI scores at the Church Creek sites. BIBI scores and ratings are summarized in Table 3-7.

| Site | BIBI Score | Narrative Rating |
|-------|------------|------------------|
| CC-01 | 2.14 | Poor |
| CC-02 | 2.14 | Poor |
| CC-03 | 2.43 | Poor |
| CC-04 | 1.86 | Very Poor |

To supplement the biological assessment data, *in situ* water quality parameters were measured at each biological monitoring site prior to sample collection. Table 3-8 shows the water quality data for each site. pH at two sites, CC-01 and CC-04, were also slightly less than the minimum pH of 6.5 specified for Use class I streams. Church Creek site conductivity values were elevated, particularly at CC-01, compared to most coastal plain streams, and exceeded the 75th percentile of values (i.e., 307 $\mu\text{S}/\text{cm}$) measured during Round One (2004-2008) of the Countywide Biological Monitoring and Assessment Program (Hill and Pieper, 2011), as well as higher than the range of those found in other urban, or highly impervious, drainage areas in Maryland (MD DNR, 2001, 2003, 2005; KCI, 2009; Hill and Crunkleton, 2009). Stream conductivity is affected by inorganic dissolved solids such as chloride, nitrate, sulfate, and phosphate anions or sodium, magnesium, calcium, iron, and aluminum cations, many of which are generally found at elevated concentrations in urban streams (Paul and Meyer, 2001). Increased stream ion concentrations (measured as conductivity) in urban systems are typically a result of runoff over impervious surfaces, passage through pipes, and exposure to other infrastructure (Cushman 2006). Seasonal use of road salt has most likely caused conductivity values to be high.

| Site | pH | Temperature | Dissolved Oxygen | Turbidity | Conductivity |
|-------|------|--------------------|------------------|-----------|-------------------------|
| | SU | $^{\circ}\text{C}$ | mg/L | NTU | $\mu\text{S}/\text{cm}$ |
| CC-01 | 6.29 | 17.6 | 13.86 | 17.5 | 2319 |
| CC-02 | 7.04 | 15.4 | 7.86 | 16.2 | 468 |
| CC-03 | 6.67 | 17.0 | 6.08 | 15.7 | 479 |
| CC-04 | 6.49 | 19.3 | 5.58 | 13.2 | 392 |

3.4 GEOMORPHIC ASSESSMENT

Due to the highly altered conditions of the drainage area (i.e., high imperviousness, altered flow regime, numerous stormwater outfalls) and stream channel (i.e., channelization, stabilization) in the study area, reliable bankfull indicators were often difficult to locate in the field. In the absence of reliable bankfull indicators, bankfull elevations were adjusted to match the predicted values for bankfull area provided by the bankfull channel geometry relationship for urban streams developed specifically for Anne Arundel County (AADPW 2002). Furthermore, categorization of segments into the Rosgen Classification scheme for natural rivers required a fair amount of professional judgment to interpret the data. When assigning the stream classification types, values for some parameters would often fit into the prescribed ranges according to the Rosgen Classification while others would not. Many of the features at the existing cross-section locations have shifted from riffle features to pool features, which can skew the channel dimensions since classifications are based on riffle dimensions. Consequently, it was often necessary to apply best professional judgment and incorporate supplemental information (e.g., presence of depositional features) in order to assign the most appropriate stream classifications. The Rosgen classification system is summarized in Appendix E and 2017 data for Church Creek sites are in Appendix F. Also noteworthy, prior to the 2016 geomorphic survey, stream restoration occurred downstream of XS-4, on an unnamed tributary, and upstream of XS-5 on the mainstem Church Creek in the vicinity of the Annapolis Harbor Center. Thus, longitudinal profile length has shortened between the 2015 and 2016 surveying as a result of the stream restoration construction and channel reengineering. The 2017 geomorphic survey gives a first look at any changes a year after the restoration was completed between XS-4 and XS-5.

The most upstream reach on the Parole Plaza Tributary, XS-1, has been undergoing a transition from a Rosgen C4/5 channel to a F4 channel, as evidenced by changes in the width/depth and entrenchment ratios. Previous monitoring in 2010 suggested that this reach was shifting from an E to a C channel as evidenced by channel degradation along the right bank and a notable increase in sediment deposition and point bar formation along the right bank just downstream. Additional degradation between 2010 and 2012 suggest that the channel had lost connectivity to the floodplain and had likely shifted to an F stream type. Mid-channel degradation continued between 2014 and 2017 showing approximately a 0.8 feet difference. In 2017, geomorphic assessment parameters continue to support the classification of this reach as an F4 channel. The channel evolution is supported by a 61.3% increase in channel cross-sectional area since 2003 and considerable widening and mid-channel bar formation immediately downstream, which is indicative of a channel that it not stable and is undergoing a widening and degradation phase. From 2014 to 2017, the cross-sectional area continues a steady increase. Left bank widening was also apparent between 2013 and 2014 monitoring years and remained consistent during 2015 and through 2017. However, it is also important to acknowledge that this cross-section is no longer located in a riffle feature and is now in a pool feature, which affects the channel dimensions and complicates classification using the Rosgen system.

The next site downstream on the Parole Plaza Tributary, XS-2, was classified as a Rosgen G4c channel based on its continued low width/depth ratio, low slope, and gravel substrate. From 2016 to the 2017 survey, the substrate size decreased slightly. Since 2012 its entrenchment ratio

was slightly higher than those typical of G streams, but in 2017 the ratio of entrenchment was shown to decrease. This reach was previously classified as an E type channel; however, it was noted that the reach was actively degrading and widening. While E streams are typically more sinuous, this segment has been noticeably straightened and stabilized by a retaining wall and rubble/fill along the left bank (facing downstream). The lack of sinuosity in the channel has likely resulted in instability, and consequently resulted in a shift to a less-stable form.

Site XS-3, located along the restored segment of Parole Plaza Tributary, was not classified until 2013, after allowing 3 years of stabilization after restoration. In 2013 and 2014 it was classified as a Rosgen G4c channel based on its low entrenchment ratio, low width/depth ratio, and low slope. In 2015 XS-3 remained a G type channel; however, the substrate had become coarser resulting in a G4/3c classification. Variable coarseness caused XS-3 to return to a G4c during the 2016 survey and it has maintained that classification in 2017. Before restoration, this cross-section was classified as a Rosgen G5c channel; however, since the Rosgen scheme was developed to classify natural channels, or those that are shaped naturally by fluvial processes, it was deemed inappropriate to classify immediately after construction. This section is still heavily armored and reliable bankfull indicators are not easily identified. Not much change has been documented at XS-3 but during the 2017 survey the right bank has begun to erode behind the armored bank.

The most downstream site on the Parole Plaza Tributary, XS-4, has transitioned from a Rosgen E5 channel to a C5/4 back to an E5/4 channel and now an E4/5 channel due to increased substrate size and fluctuations in width/depth ratio. A large woody debris jam located just downstream of the cross-section location has resulted in a considerable accumulation of fine sediment and debris across the channel and, consequently, has led to aggradation and a reduction in the cross-sectional area up until 2016. In 2016, before the cross-section survey was performed, restoration on the reach began and was completed just downstream of XS-4. Construction activities included the removal of the woody debris jam. A year after the construction it is likely the fine sediment that was behind the debris jam has cleared resulting in the increase substrate size. Between 2011 and 2015 cross-sectional area has consistently been lower than baseline monitoring in 2003. Restoration in 2016 has caused cross-sectional area to increase by 9.8% from 2003 monitoring. Subsequently, in 2017 the cross-sectional area decreased from 2016 by 15.3% and has decreased by 7.0 % since the 2003 monitoring.

Located on the mainstem of Church Creek, upstream of the MD Rt. 2 culvert, XS-5 has transformed from a Rosgen C3/5 channel into a F4 channel due to a significantly decreased entrenchment ratio from 4.0 to 1.7 between 2012 and 2017. Between 2015 and 2016 this portion of the reach has become slightly less coarse from a D50 of 61 mm to 24 mm. In 2017, this reach maintained the same D50 particle size. This segment shows evidence of previous alteration in the form of cobble-sized riprap armoring along the bed and lower banks to protect a sewer line crossing and obvious channel straightening, which explains the lack of sinuosity typical of F type streams. The substantial amount of cobble-sized rip-rap in the stream channel has resulted in a bi-modal distribution of substrate particles within this reach, with a predominance of gravel in the pools and artificial cobbles in the riffles. In 2017, the cross-sectional area and the width/depth ratio decreased slightly.

4 DISCUSSION

Results from the July 2016-June 2017 study period are discussed in the following section. Water quality, biological, and geomorphological data are interpreted, presented and compared to previous data. A discussion of the characteristics of the watershed is also included.

4.1 WATER CHEMISTRY

Water quality criteria are presented in Table 4-1. The measured data are compared, where possible, to these criteria to assess the extent of the pollution in this tributary.

| Table 4-1. State and Federal water quality criteria available for parameters sampled at Church Creek | | | |
|--|---------|-------|----------------------|
| Parameter (mg/L, except as noted) | Chronic | Acute | Reference |
| Lead (µg/L) | 2.5 | 65 | COMAR 26.08.02.03-2 |
| Copper (µg/L) | 9 | 13 | COMAR 26.08.02.03-2 |
| Zinc (µg/L) | 120 | 120 | COMAR 26.08.02.03-2 |
| Total P | 0.0225 | | USEPA 2000 |
| BOD ₅ | 7 | | USEPA 1986 |
| Nitrate + Nitrite | 0.095 | | USEPA 2000 |
| TSS | 500 | | USEPA 1974 |
| TKN | None | | |
| TPH | None | | |
| <i>E. coli</i> * (MPN/100ml) | 235 | | COMAR 26.08.02.03-3. |
| Hardness | None | | |

* Used most restrictive standard for *E. coli* as a conservative approach: frequent full body contact recreation criterion.

Criteria are used to protect against both short-term and long-term effects. Numeric criteria are important where the cause of toxicity is known or for protection against pollutants with potential human health impacts or bioaccumulation potential. Narrative criteria can be the basis for limiting toxicity in discharges where a specific pollutant can be identified as contributing to the toxicity. Biological criteria can be used to complement traditional, chemical-specific criteria as indicators of aquatic health and impacts to the aquatic ecosystem.

Table 4-2 compares storm event results to the Federal and State acute criteria. Comparison and interpretation of Church Creek pollutant concentrations to Federal and State water quality criteria, and relating these conditions to ultimate ecological outcomes in the system, however, are difficult. Criteria do not exist for all parameters measured at the monitoring stations. In addition, a clear cause and effect relationship between water quality and

ecological condition is difficult to determine. However, these comparisons can be used as general indicators of water quality impairment. Both State and Federal criteria are based on ambient stream conditions. Chronic criteria consider the maximum levels at which aquatic life can survive if continuously subjected to a pollutant concentration. Acute criteria reflect the maximum level at which an aquatic organism can survive if periodically subjected to a pollutant concentration. Since storm events represent a periodic condition, wet-weather samples are compared only to acute criteria.

| Table 4-2. Maximum concentrations observed for wet weather samples compared to appropriate criteria | | | |
|---|--------|--------------|--------------|
| Parameter (mg/L, except as noted) | Acute | Church Creek | Parole Plaza |
| Lead (µg/L) | 65 | 401* | 55 |
| Copper (µg/L) | 13 | 408* | 200* |
| Zinc (µg/L) | 120 | 3600* | 1810* |
| Total P | 0.0225 | 2.6* | 0.49* |
| BOD ₅ | 7 | 23* | 25* |
| Nitrate + Nitrite | 0.095 | 1.70* | 3* |
| TSS | 500 | 400 | 530* |
| TKN | None | 4.4 | 3 |
| TPH | None | 12 | 10 |
| <i>E. coli</i> ** (MPN/100ml) | 235 | 91,000* | 65,044* |
| Hardness | None | 230 | 400 |
| * Criterion exceeded | | | |
| ** Used most restrictive standard for <i>E. coli</i> as a conservative approach: frequent full body contact recreation criterion. | | | |

As in prior years, comparisons to water quality criteria continue to indicate elevated pollutant concentrations in the Church Creek watershed. In particular, copper, zinc, total phosphorous, BOD₅, nitrate-nitrite, and *E. coli* frequently exceeded criteria at both sampling stations. During the June 19, 2017 storm event, at Church Creek, lead exceeded its corresponding criterion along with TSS at Parole Plaza. Table 4-2 (above) shows the maximum concentrations for each sampling site, and compares these to the criteria.

Table 4-3 shows the percentage of wet weather samples for which criteria were exceeded. Water quality criteria for the pollutants listed above were more frequently exceeded at the Parole Plaza station than at the Church Creek station for all contaminants except lead. *E. coli* concentrations also remained high at both stations throughout the 2017 monitoring period, exceeding water quality criteria 72 percent of the time at Church Creek and 81 percent of the time at Parole Plaza. Note that prior to site stabilization, total suspended solids concentrations had been particularly high due to construction activity at Annapolis Towne Centre. Following stabilization of the site in Fall 2008, the event mean concentrations for total suspended solids

have dropped significantly. However, during the June 19 storm event, total suspended solids concentrations exceeded the water quality criterion during the rising limb at Parole Plaza, but the event mean concentration for the storm remained low. At Church Creek TSS did not exceed the criterion, but there was a higher concentration (400 mg/L) during the September 27, 2017 storm event.

| Table 4-3. Percentage of all wet weather samples that exceed appropriate criteria | | | |
|---|-------------------------------------|---------------------|---------------------|
| Parameter | Criteria (mg/L, except as noted) | Church Creek (%) | Parole Plaza (%) |
| Lead (µg/L) | 65 | 3 | 0 |
| Copper (µg/L) | 13 | 36 | 81 |
| Zinc (µg/L) | 120 | 22 | 69 |
| Total P | 0.0225 | 100 | 100 |
| BOD ₅ | 7 | 31 | 42 |
| Nitrate + Nitrite | 0.095 | 100 | 100 |
| TSS | 500 | 0 | 3 |
| TKN | None | NA | NA |
| TPH | None | NA | NA |
| <i>E. coli</i> * (MPN/100ml) | 235 | 72 | 81 |
| Hardness | None | NA | NA |

* Used most restrictive standard for *E. coli* as a conservative approach: frequent full body contact recreation criterion.

Table 4-4 shows the annual average event mean concentrations that exceeded water quality criteria. As can be seen from the table, copper, total phosphorous, nitrate-nitrite, and *E. coli* consistently exceeded their corresponding criteria at both stations. At both sites, the annual average event mean concentrations for copper exceeded both the chronic and acute criteria.

High levels of pollutants observed in the watershed are typical for commercial and retail land uses that are coupled with high levels of automobile traffic and impervious surface area (U.S. EPA 1983). As shown in Table 2-2, 87% of the watershed to the Parole monitoring station and 69% of the watershed to the Church Creek station is impervious.

In 2007, loading rates (Tables 4-5 and 4-6) increased sharply at both stations. Loading rates in 2008 were still high, when compared to historical values, but dropped dramatically from the 2007 levels. During the 2009 reporting year, loading rates dropped further, and aligned more closely with historical values. High loading rates in 2007 likely resulted from construction activity that was underway immediately upstream of the Parole Plaza station. Since the majority of the site was stabilized by the end of 2008, the cessation of construction likely caused pollutant loads to decrease.

| Table 4-4. Annual average event mean concentrations and criteria (parameters that exceeded appropriate criteria are indicated) | | | | |
|--|------------------|----------------|-----------------------|-----------------------|
| Parameter (mg/L, except as noted) | Chronic Criteria | Acute Criteria | Church Creek | Parole Plaza |
| Lead (µg/L) | 2.5 | 65 | 5.1 ^(a) | 2.2 |
| Copper (µg/L) | 9 | 13 | 13.1 ^(a,b) | 19.7 ^(a,b) |
| Zinc (µg/L) | 120 | 120 | 83 | 137 ^(a,b) |
| Total P | 0.0225 | | 0.14 ^(a) | 0.12 ^(a) |
| BOD ₅ | 7 | | 3 | 2 |
| Nitrate + Nitrite | 0.095 | | 0.33 ^(a) | 0.37 ^(a) |
| TSS | 500 | | 38 | 47 |
| TKN | None | | 0.69 | 0.68 |
| TPH | None | | 0.14 | 0.46 |
| <i>E. coli</i> * (MPN/100ml) | 235 | | 5,597 ^(a) | 7,366 ^(a) |
| Hardness | None | | 43 | 44 |

^(a) Chronic or general criterion exceeded
^(b) Acute criterion exceeded
* Used most restrictive standard for *E. coli* as a conservative approach: frequent full body contact recreation criterion.

| Table 4-5. Total annual loading rates, in pounds, observed at the Parole Plaza Sampling Station from 2002 to 2017 | | | | | | | | | | |
|---|--------------|---------------|------------|------------|----------------------------------|-----------|-----------|-----------|---------------|-------------------------------|
| Year | BOD | TSS | TP | TKN | NO ₃ +NO ₂ | Zinc | Lead | Copper | Hardness | Fecal Coliform ^(a) |
| 2002 | 2,912 | 26,585 | 1,178 | 388 | 323 | 58 | 14 | 1 | NA | 1,152,001 |
| 2003 | 21,665 | 86,385 | 372 | 1,477 | 714 | 176 | 69 | 15 | NA | 5,350,164 |
| 2004 | 8,025 | 57,447 | 293 | 655 | 391 | 57 | 7 | 8 | NA | 402,127 |
| 2005 | 4,573 | 33,015 | 184 | 483 | 350 | 50 | 12 | 8 | NA | 665,232 |
| 2006 | 13,562 | 94,306 | 650 | 1,867 | 410 | 177 | 13 | 25 | NA | 3,360,952 |
| | | | | | | | | | | <i>E. coli</i> ^(a) |
| 2007 | 40,009 | 848,116 | 1,649 | 2,328 | 1,401 | 349 | 26 | 162 | NA | 11,017 |
| 2008 ^(b) | NA | NA | NA | NA | NA | NA | NA | NA | NA | NA |
| 2009 | 2,175 | 11,787 | 59 | 490 | 117 | 56 | 0.8 | 6.5 | NA | 2,115 |
| 2010 | 2,209 | 17,609 | 89 | 309 | 120 | 40 | 1.2 | 4.1 | NA | 1,740 |
| 2011 | 2,114 | 13,894 | 42 | 371 | 131 | 58 | 1.1 | 6.3 | 6,987 | 2,682 |
| 2012 | 3,660 | 15,335 | 62 | 284 | 214 | 57 | 1.0 | 6.6 | 14,578 | 10,209 |
| 2013 | 1,481 | 6,079 | 34 | 155 | 108 | 34 | 0.5 | 4.9 | 8,586 | 16,041 |
| 2014 | 2,040 | 18,953 | 54 | 536 | 497 | 50 | 1.0 | 8.1 | 36,945 | 12,716 |
| 2015 | 940 | 14,606 | 45 | 232 | 162 | 38 | 1.1 | 5.3 | 29,023 | 3,333 |
| 2016 | 1,308 | 10,887 | 29 | 218 | 103 | 36 | 1.0 | 9.3 | 14,779 | 18,268 |
| 2017 | 1,120 | 19,913 | 50 | 318 | 161 | 57 | 1.2 | 8.3 | 18,876 | 7,366 |
| 2002-2006 Mean | 8,544 | 59,548 | 535 | 974 | 438 | 104 | 23 | 11 | NA | 2,186,095 |
| 2009-2017 Mean | 1,894 | 14,340 | 52 | 324 | 179 | 47 | 1 | 7 | 18,539 | 8,274 |
| 2002-2017 Mean | 6,652 | 84,994 | 319 | 674 | 347 | 86 | 10 | 19 | 18,539 | 8,549 |

^(a) Units of Fecal Coliform and *E. coli* are MPN/100 mL.
^(b) In 2008, monitoring was conducted for both outfalls at Parole Plaza, but continuous level monitoring was not available for the 54" RCP; therefore, loads could not be calculated.
^(c) Mean *E. coli* value, does not include pre-2007 Fecal Coliform data.

| Table 4-6. Total annual loading rates, in pounds, observed at the Church Creek Sampling Station from 2002 to 2017 | | | | | | | | | | |
|---|---------------|----------------|--------------|--------------|----------------------------------|------------|-----------|-----------|----------------|------------------------|
| Year | BOD | TSS | TP | TKN | NO ₃ +NO ₂ | Zinc | Lead | Copper | Hardness | Fecal Coliform* |
| 2002 | 6,408 | 58,501 | 2,593 | 854 | 711 | 127 | 32 | 3 | NA | 2,534,970 |
| 2003 | 47,673 | 190,090 | 818 | 3,250 | 1,571 | 387 | 151 | 32 | NA | 11,773,001 |
| 2004 | 17,660 | 126,411 | 645 | 1,441 | 860 | 126 | 19 | 18 | NA | 884,887 |
| 2005 | 10,062 | 72,648 | 405 | 1,062 | 771 | 109 | 27 | 16 | NA | 1,463,839 |
| 2006 | 29,844 | 207,520 | 1,431 | 4,109 | 902 | 390 | 29 | 54 | NA | 7,395,753 |
| | | | | | | | | | | <i>E. coli</i>* |
| 2007 | 265,499 | 3,312,794 | 8,381 | 20,330 | 436,206 | 3,663 | 277 | 652 | NA | 1,755 |
| 2008 | 60,843 | 458,185 | 3,037 | 12,468 | 4,444 | 693 | 37 | 36 | NA | 3,857 |
| 2009 | 35,521 | 206,184 | 1,296 | 9,377 | 2,505 | 531 | 30 | 57 | NA | 3,912 |
| 2010 | 49,256 | 341,877 | 2,066 | 9,561 | 2,912 | 739 | 39 | 77 | NA | 3,358 |
| 2011 | 42,883 | 214,820 | 1,340 | 7,410 | 3,606 | 704 | 30 | 41 | 259,076 | 3,995 |
| 2012 | 40,145 | 150,490 | 1,103 | 3,714 | 3,018 | 551 | 20 | 31 | 250,747 | 5,549 |
| 2013 | 43,980 | 180,946 | 899 | 3,326 | 2,782 | 558 | 27 | 57 | 314,179 | 2,399 |
| 2014 | 31,969 | 299,830 | 1,065 | 12,177 | 6,019 | 551 | 27 | 78 | 646,801 | 8,638 |
| 2015 | 19,643 | 344,419 | 1,057 | 5,743 | 3,148 | 665 | 35 | 99 | 455,627 | 2,100 |
| 2016 | 46,587 | 335,422 | 1,026 | 6,648 | 3,081 | 818 | 41 | 92 | 344,729 | 8,049 |
| 2017 | 23,557 | 230,599 | 855 | 4,699 | 2,044 | 468 | 34 | 71 | 257,816 | 5,597 |
| 2002-2006 Mean | 22,329 | 131,034 | 1,178 | 2,143 | 963 | 228 | 52 | 25 | NA | 4,810,490 |
| 2009-2017 Mean | 37,060 | 256,065 | 1,190 | 6,962 | 3,235 | 621 | 31 | 67 | 361,282 | 4,844 |
| 2002-2017 Mean | 48,221 | 420,671 | 1,751 | 6,636 | 29,661 | 693 | 53 | 88 | 361,282 | **4,474 |

* Units of Fecal Coliform and *E. coli* are MPN/100 mL.
 ** Mean *E. coli* value, does not include pre-2007 Fecal Coliform data.

When compared to the 2016 reporting year, 2017 loading rates increased for all sampled parameters at the Parole Plaza Station with the exception of BOD₅, copper, and *E. coli*. At the Church Creek Station, 2017 reporting year loading rates decreased for all sampled parameters when compared to 2016. The reduction in loads at the Church Creek Station may indicate that the stream restoration provided immediate benefit. The change in loading may also be due to natural variability in stream pollutant concentrations coupled with a 17% decline in annual discharge because of lower total rainfall in 2017 compared to 2016.

During the post-construction period (2009 to 2017), average loading rates at Parole Plaza have been lower than the levels existing prior to the redevelopment of the Towne Centre. However, at the Church Creek station, all average post-construction parameters except for lead and *E. coli* have exceeded average pre-construction (2002-2006) monitoring levels, and continued to do so in 2017.

Seasonal pollutant loads in 2017 are provided in Table 4-7. At Church Creek, the highest pollutant loads were recorded in the summer, except for copper and zinc which were highest in the winter. The highest loads of all parameters at Parole Plaza were recorded in the winter, except for *E. coli* (spring). At Parole Plaza, hardness was much higher in the winter, likely due to the use of salt to deice local roads and sidewalks associated with the extensive residential and shopping areas in this drainage area. TSS was distinctly higher, probably due to scraping action of plows on driving surfaces and the breakup of asphalt after several freeze-thaw cycles caused

by the swings in temperature observed during the 2017 winter. At Parole Plaza in the winter, metal and total phosphorus loads increased and were likely associated with high seasonal TSS load.

Table 4-7. Seasonal loading rates, in pounds, observed at the Church Creek and Parole Plaza sampling stations in 2017

| Season | BOD | TSS | TP | TKN | NO ₃ +NO ₂ | Zinc | Lead | Copper | Hardness | <i>E. coli</i> * |
|---------------------|--------|---------|-----|-------|----------------------------------|------|------|--------|----------|------------------|
| Church Creek | | | | | | | | | | |
| Summer | 10,878 | 103,527 | 295 | 1,641 | 691 | 133 | 14 | 17 | 76,274 | 10,021 |
| Fall | 4,912 | 25,049 | 181 | 1,142 | 514 | 65 | 2 | 8 | 61,669 | 2,222 |
| Winter | 3,548 | 55,872 | 207 | 1,200 | 447 | 167 | 11 | 28 | 63,794 | 2,282 |
| Spring | 4,219 | 46,150 | 172 | 716 | 392 | 105 | 7 | 19 | 56,080 | 7,489 |
| Parole Plaza | | | | | | | | | | |
| Summer | 217 | 1,347 | 9 | 61 | 43 | 8 | 0.1 | 1 | 1,607 | 4,895 |
| Fall | 191 | 1,362 | 8 | 59 | 27 | 8 | 0.1 | 1 | 2,487 | 2,504 |
| Winter | 462 | 15,354 | 25 | 146 | 67 | 30 | 0.8 | 5 | 12,060 | 3,971 |
| Spring | 251 | 1,850 | 9 | 52 | 24 | 12 | 0 | 1 | 2,722 | 16,530 |

* Units of *E. coli* are MPN/100 mL.

Annual average EMCs were plotted for each monitoring year. Plots were constructed to illustrate the impact that construction activity and redevelopment of the Annapolis Towne Centre site has had on water quality within the study reach. Figures 4-1 through 4-5 show how EMCs have changed from 2004 to 2017 at the Parole Monitoring Station. Nearly every concentration rose substantially between 2006 and 2007 when the majority of the site work took place at the Towne Centre. These concentrations fell notably in 2008, as the site stabilized. This downward trend continued in 2009. The reduction in pollutant concentrations stabilized in 2010 and 2011 possibly indicating that the stream has reached a post-construction baseline. The 2013 rise in TPH was due to an increase in the detection limit, and may not be associated with an actual increase in concentration, as greater than 95% of TPH concentrations fell below the detection limit. It is important to note that the 2013 data included in these plots do not include summer season data, which is often the season that produces the highest EMCs for many of the parameters. At Parole Plaza in 2017, most parameters were highest during the winter compared to other seasons, except *E. coli* which was highest during the spring. Annual pollutant concentrations in 2017 slightly decreased for most parameters, and substantially decreased for *E. coli*, compared to those from 2016. Total zinc, TKN, and TSS slightly increased in 2017 at Parole Plaza.

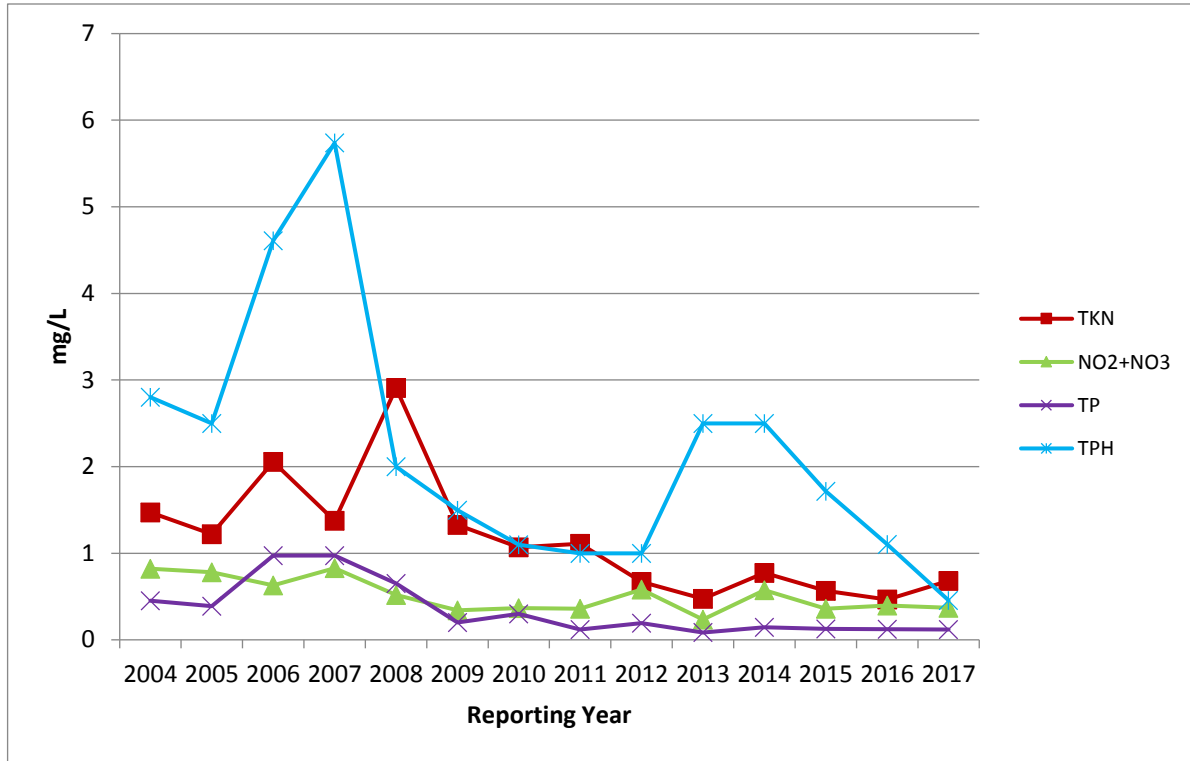


Figure 4-1. Parole Plaza station long-term monitoring: annual EMCs (TKN, NO₂+NO₃, TP, TPH; mg/L)

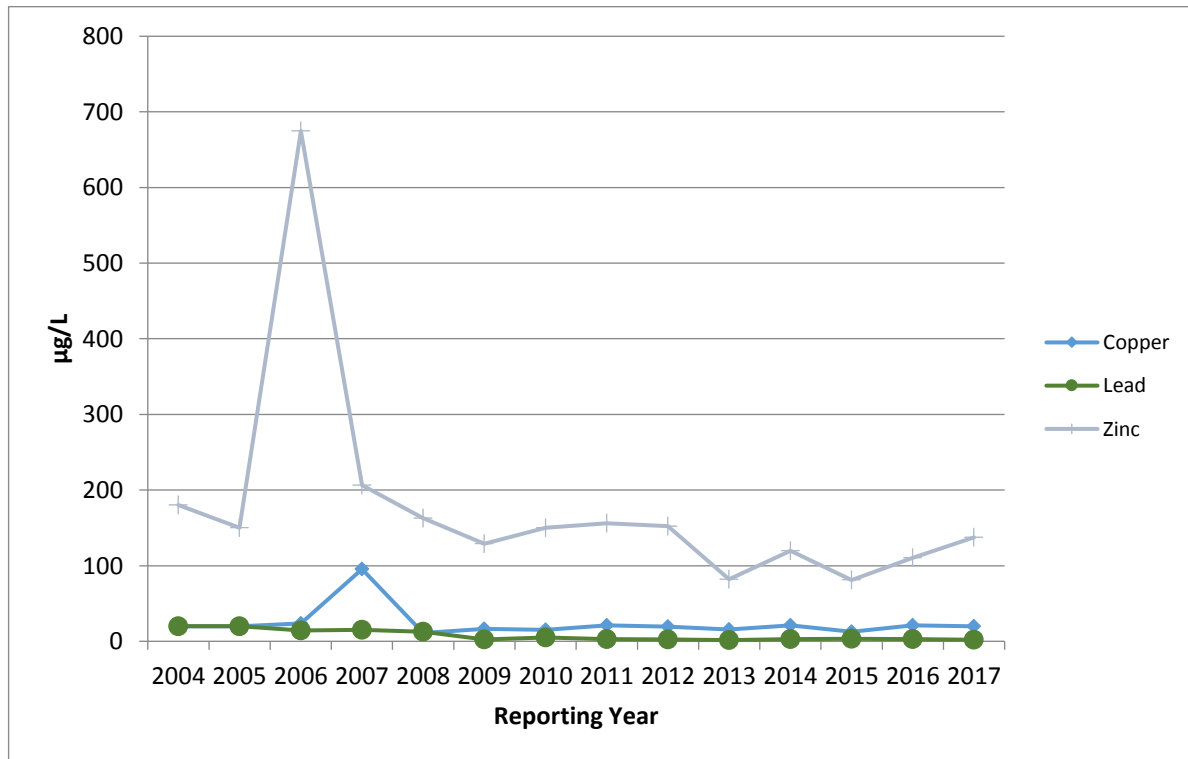


Figure 4-2. Parole Plaza station long-term monitoring: annual EMCs (Cu, Pb, Zn; µg/L)

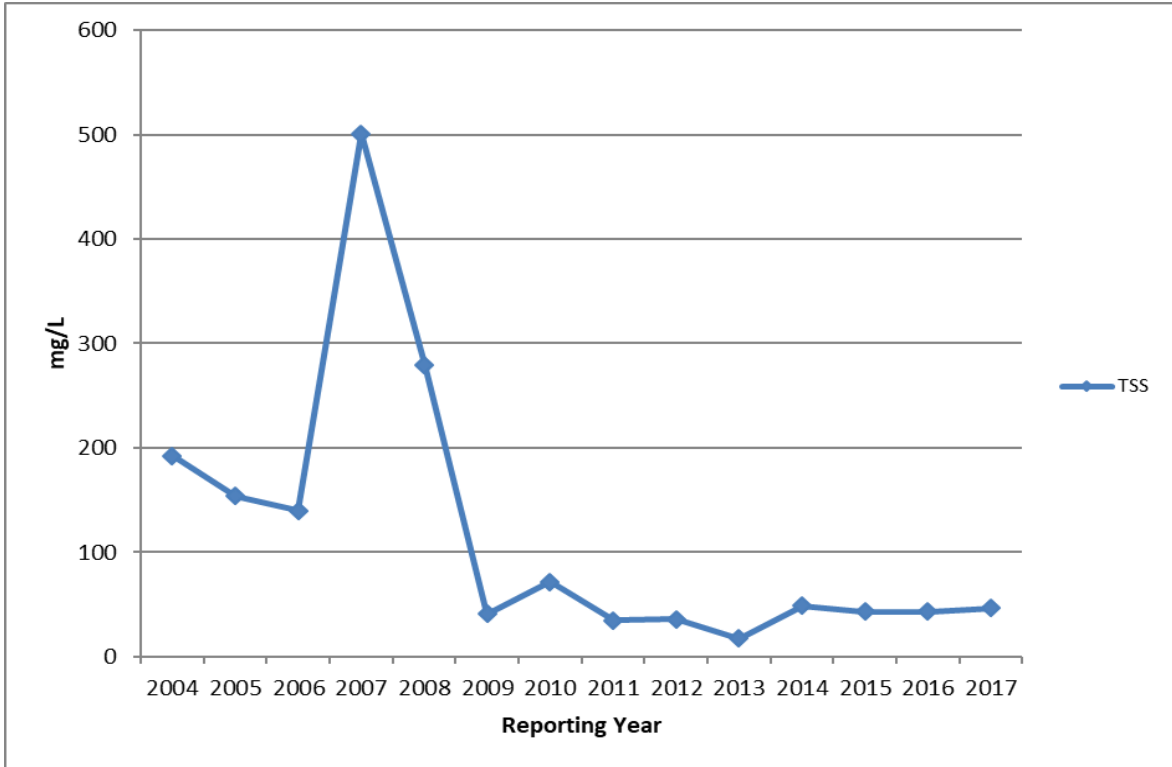


Figure 4-3. Parole Plaza station long-term monitoring: annual EMCs (TSS; mg/L)

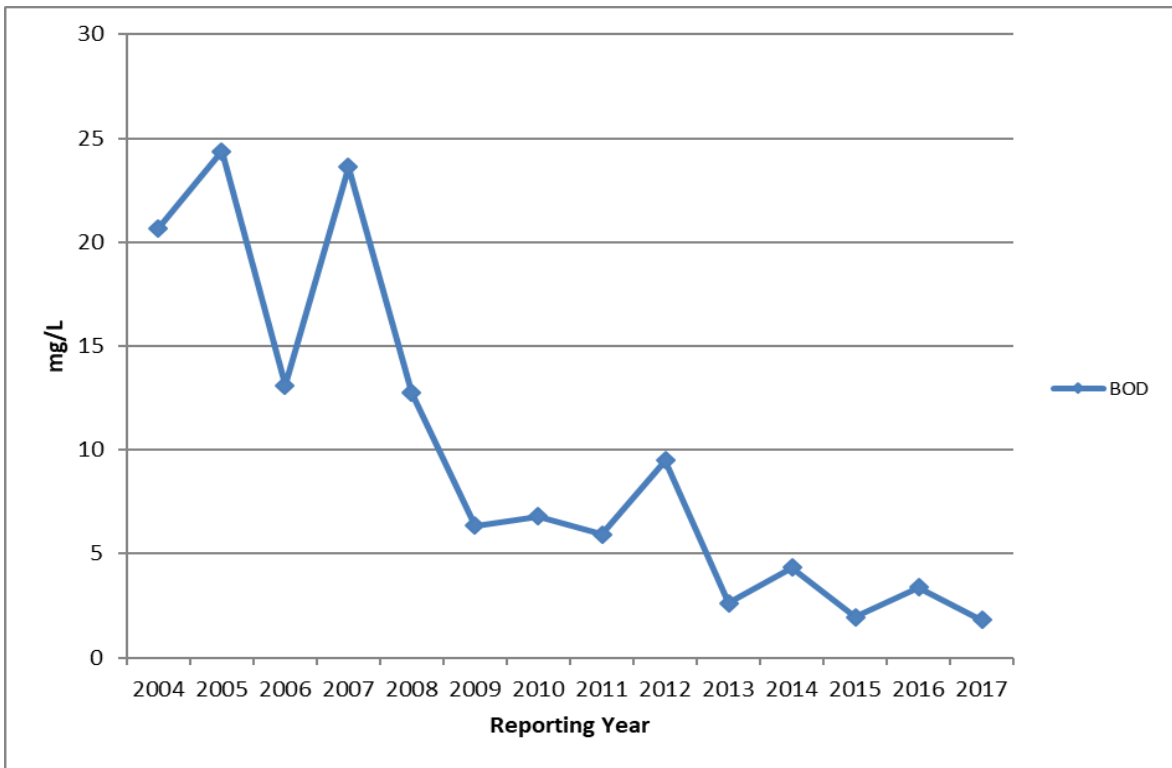


Figure 4-4. Parole Plaza station long-term monitoring: annual EMCs (BOD₅; mg/L)

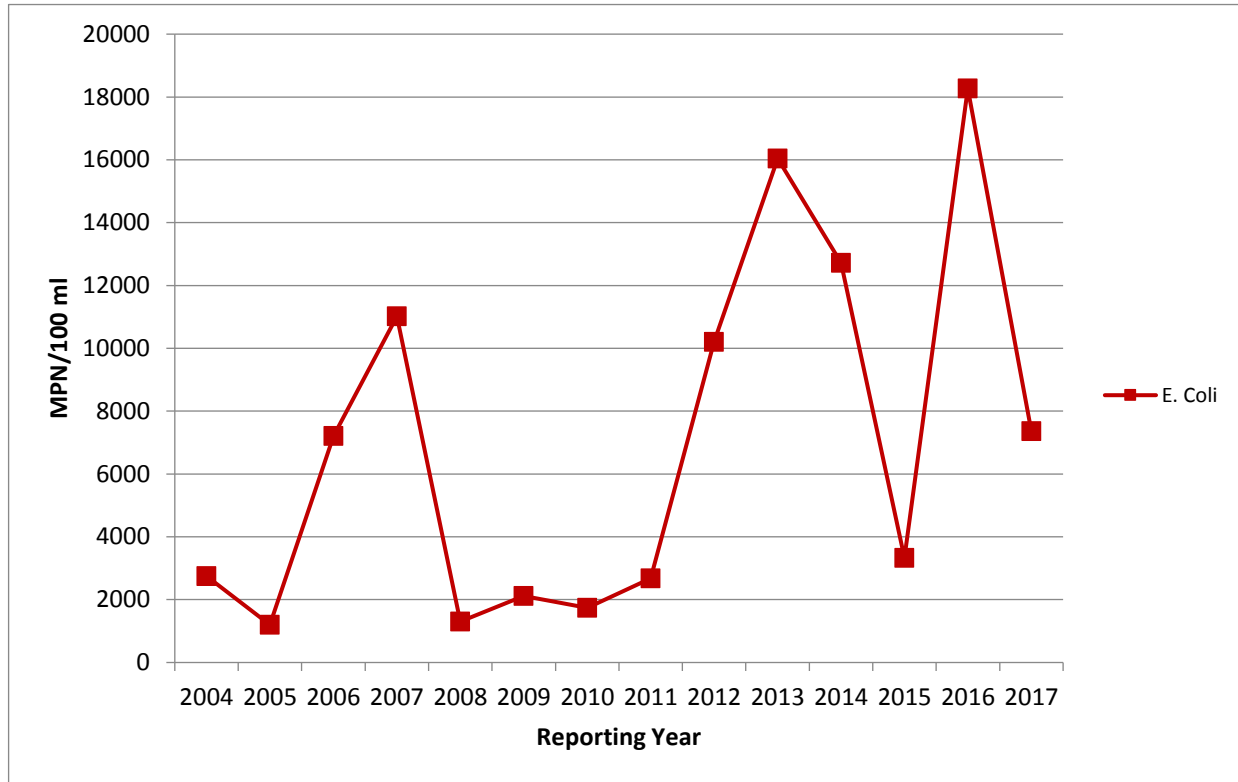


Figure 4-5. Parole Plaza station long-term monitoring: annual EMCs (*E. coli*; MPN/100 mL)

Figures 4-6 through 4-10 show a slight change in the trend in EMCs for the Church Creek monitoring station. A decrease in pollutant concentrations at Church Creek was observed in 2017 for most parameters when compared to 2016 EMCs except for total phosphorus and total copper which slightly increased. Note that the apparent rise in TPH at Church Creek in 2013 was due to an increase in the detection limit. Also, summer season concentrations were not included with the 2013 EMC data. In 2017, at Church Creek total zinc and copper were highest during the winter but all other parameters were highest during the summer. The restoration work conducted during the prior monitoring period (FY2016) could not be directly connected to the observed pollutant concentrations at Church Creek in the 2017 monitoring period. Changes in annual EMCs between the prior and current monitoring period appear to be within the normal variability of historical values or continuations of already decreasing trends (e.g., BOD₅ and TPH), although the influence of the stream restoration work cannot be ruled out.

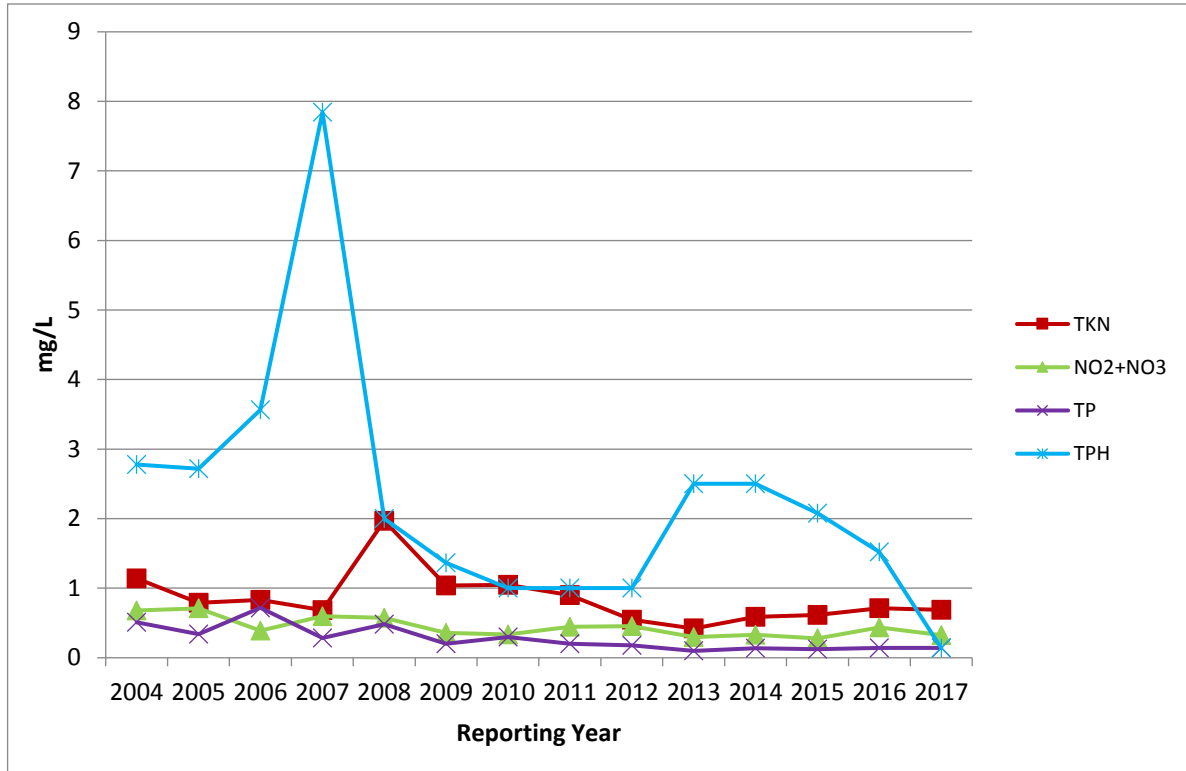


Figure 4-6. Church Creek station long-term monitoring: annual EMCs (TKN, NO₂+NO₃, TP, TPH; mg/L)

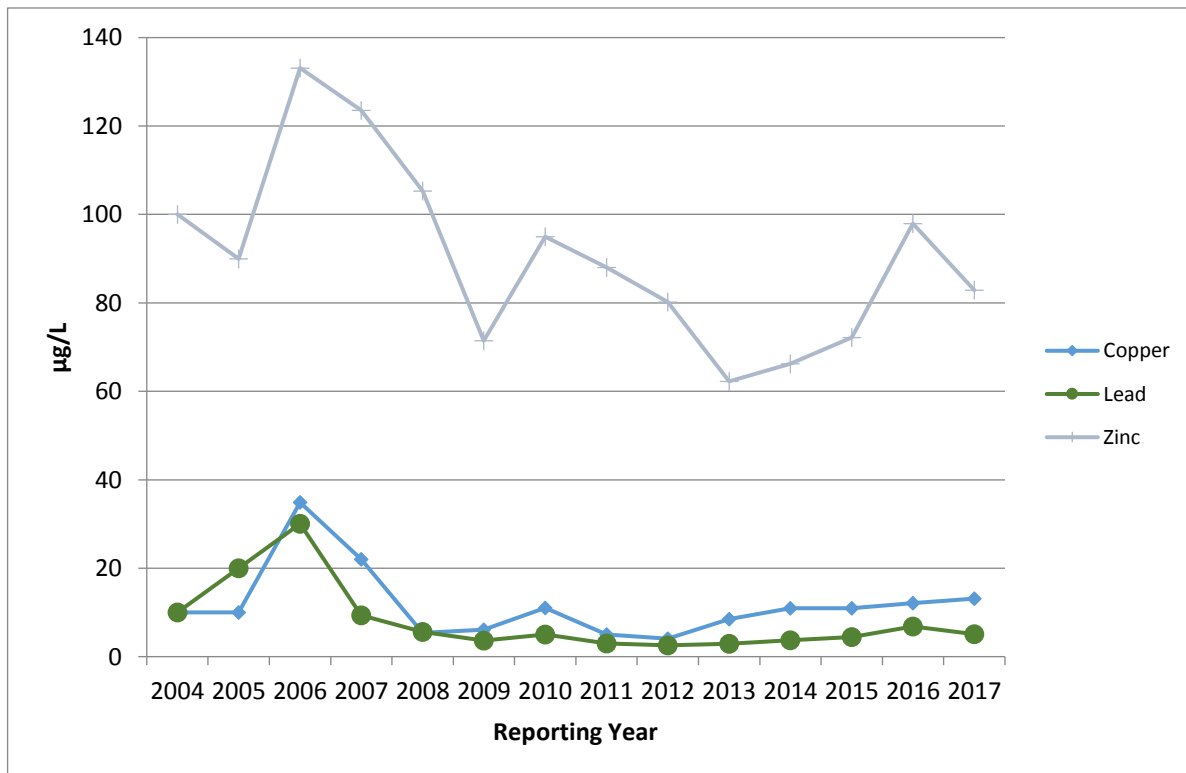


Figure 4-7. Church Creek station long-term monitoring: annual EMCs (Cu, Pb, Zn; mg/L)

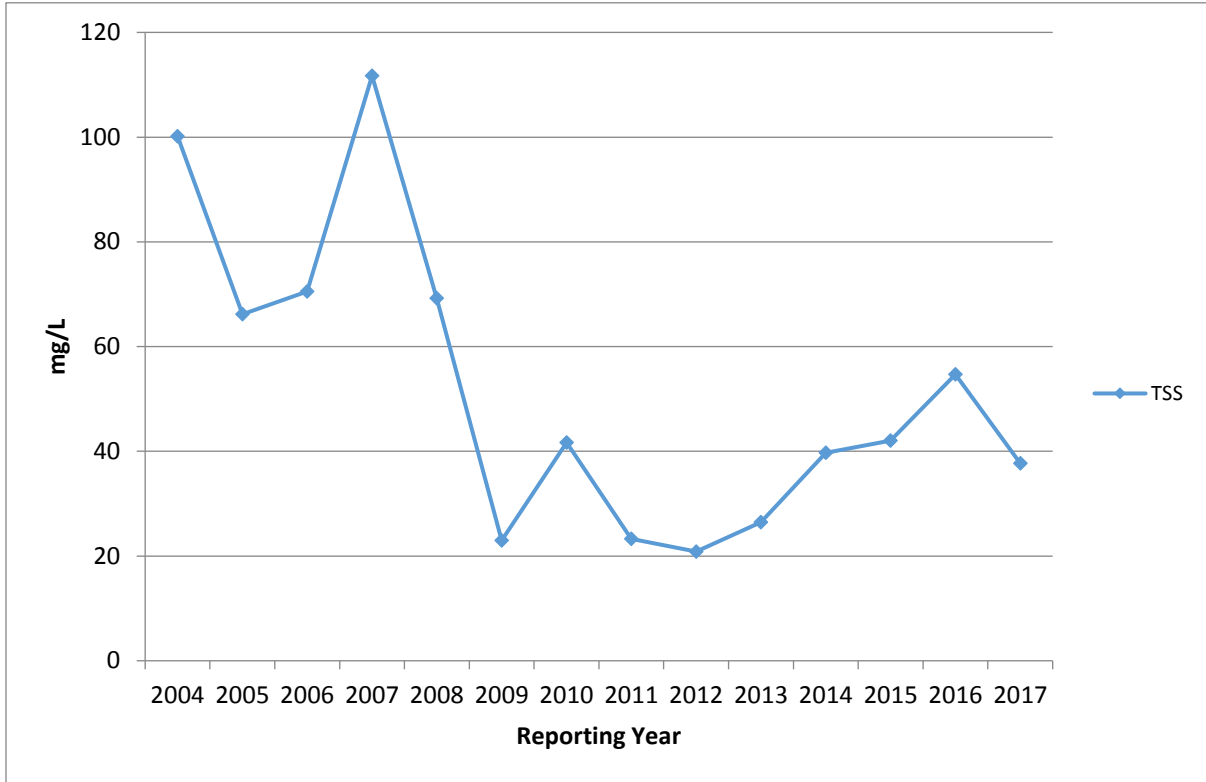


Figure 4-8. Church Creek station long-term monitoring: annual EMCs (TSS; mg/L)

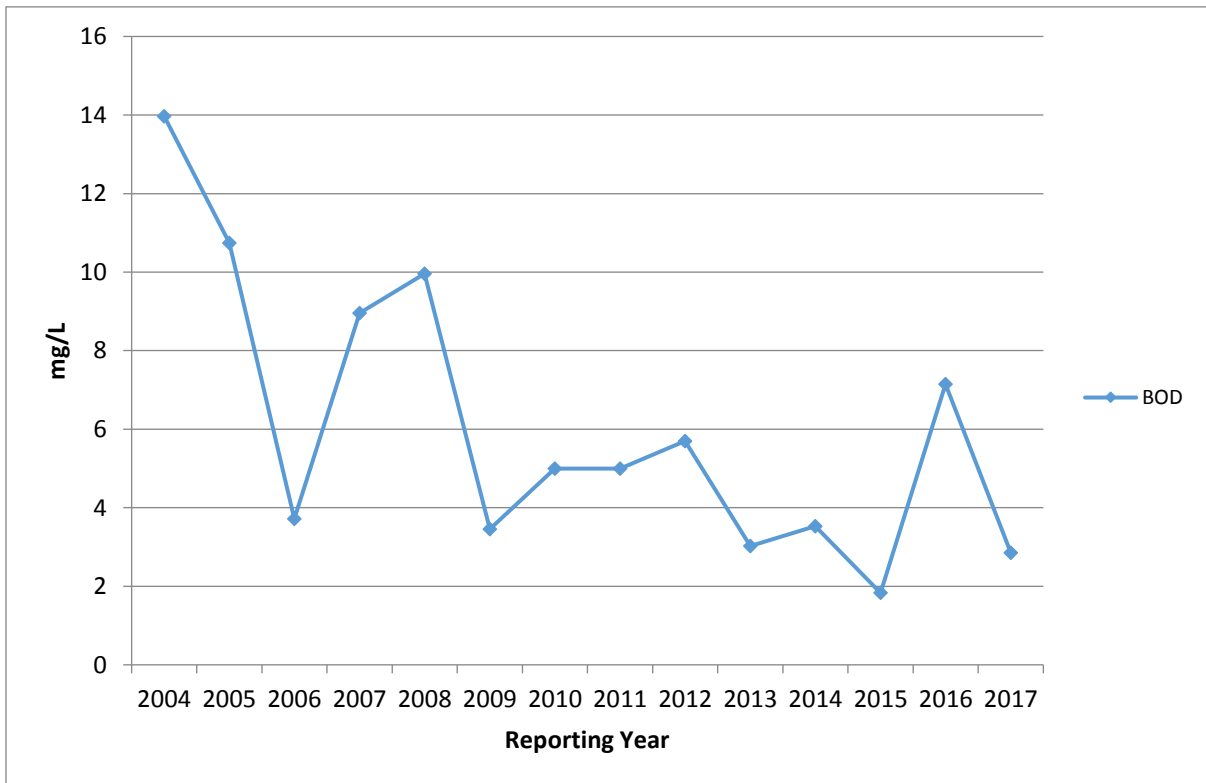


Figure 4-9. Church Creek station long-term monitoring: annual EMCs (BOD₅; mg/L)

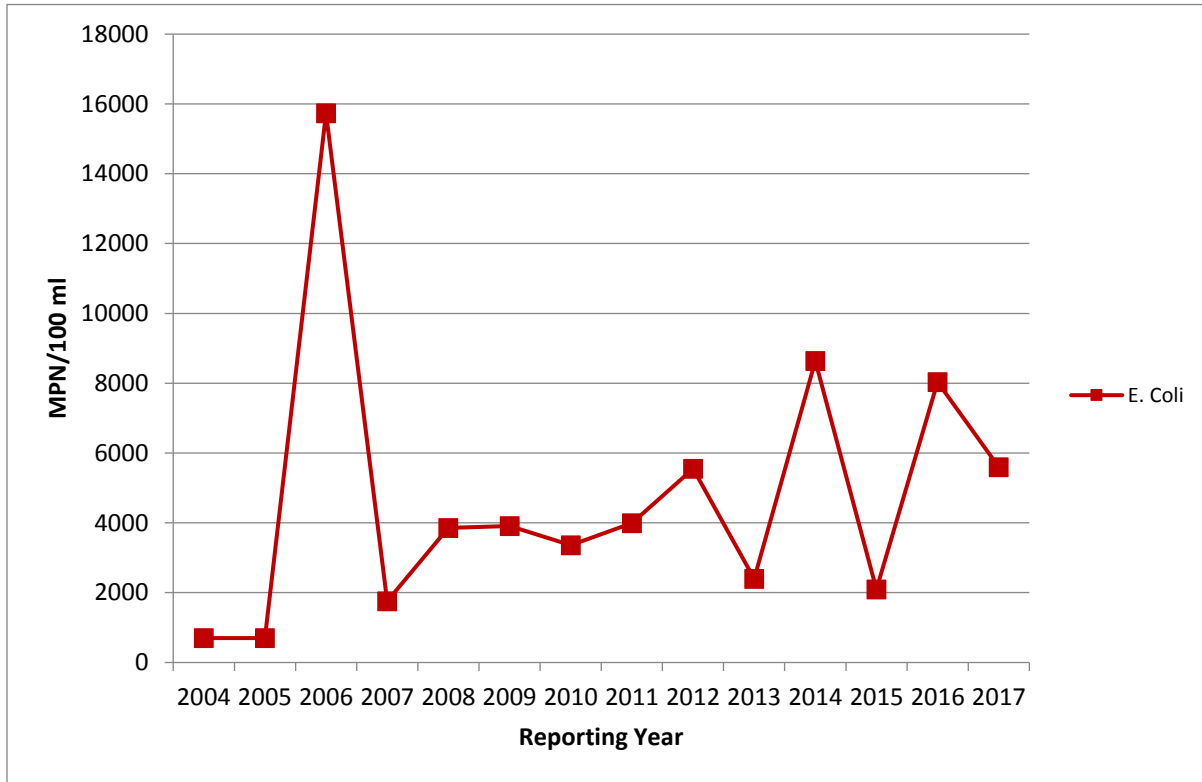


Figure 4-10. Church Creek station long-term monitoring: annual EMCs (*E. coli*; MPN/100 mL)

4.2 PHYSICAL HABITAT AND BIOLOGICAL CONDITIONS

Physical habitat and biological conditions within the Church Creek study area continue to be impaired by urbanization within the surrounding watershed. Stream physical habitat remains degraded throughout the entire study reach (Table 4-8, Figure 4-11). While scores at all sites increased in 2017, the change in score was sufficient to shift the associated narrative rating for CC-01 and CC-04 into a higher category, ‘Partially Degraded’, than that observed in 2016, ‘Degraded’. Nonetheless, urban stressors such as hydrologic alteration (i.e., increased runoff, increased frequency of peak flows, reduced infiltration) within the watershed have resulted in a reduction of stable instream habitat as well as increased channel erosion and sedimentation. A general lack of a stable epifaunal substrate further limits the capacity of the stream to support a diverse and healthy macroinvertebrate community. In addition, elevated conductivity levels reflect high levels of dissolved solids during baseflow conditions, which typically indicate the presence of water quality stressors.

| Table 4-8. PHI scores from 2006 to 2017 | | | | | |
|---|-----------|--------------------|----------|--------------------|--------------------|
| Site | | CC-01 | CC-02 | CC-03 | CC-04 |
| 2006 | PHI Score | 51.1 | 55.4 | 56.8 | No Data |
| | Rating | Degraded | Degraded | Degraded | Collected |
| 2007 | PHI Score | 61.2 | 59.1 | 65.7 | 60.8 |
| | Rating | Degraded | Degraded | Degraded | Degraded |
| 2008 | PHI Score | 57.1 | 56.8 | 66.6 | 62.6 |
| | Rating | Degraded | Degraded | Partially Degraded | Degraded |
| 2009 | PHI Score | 73.2 | 59.6 | 69.2 | 65.2 |
| | Rating | Partially Degraded | Degraded | Partially Degraded | Degraded |
| 2010 | PHI Score | 64.3 | 53.9 | 65.0 | 62.3 |
| | Rating | Degraded | Degraded | Degraded | Degraded |
| 2011 | PHI Score | 67.4 | 55.3 | 66.9 | 61.5 |
| | Rating | Partially Degraded | Degraded | Partially Degraded | Degraded |
| 2012 | PHI Score | 69.2 | 51.5 | 62.5 | 58.3 |
| | Rating | Partially Degraded | Degraded | Degraded | Degraded |
| 2013 | PHI Score | 63.0 | 53.5 | 66.6 | 57.5 |
| | Rating | Degraded | Degraded | Partially Degraded | Degraded |
| 2014 | PHI Score | 65.85 | 56.16 | 70.79 | 61.01 |
| | Rating | Degraded | Degraded | Partially Degraded | Degraded |
| 2015 | PHI Score | 66.35 | 52.93 | 66.68 | 62.70 |
| | Rating | Partially Degraded | Degraded | Partially Degraded | Degraded |
| 2016 | PHI Score | 64.80 | 58.47 | 68.64 | 62.70 |
| | Rating | Degraded | Degraded | Partially Degraded | Degraded |
| 2017 | PHI Score | 67.41 | 60.97 | 71.72 | 67.92 |
| | Rating | Partially Degraded | Degraded | Partially Degraded | Partially Degraded |

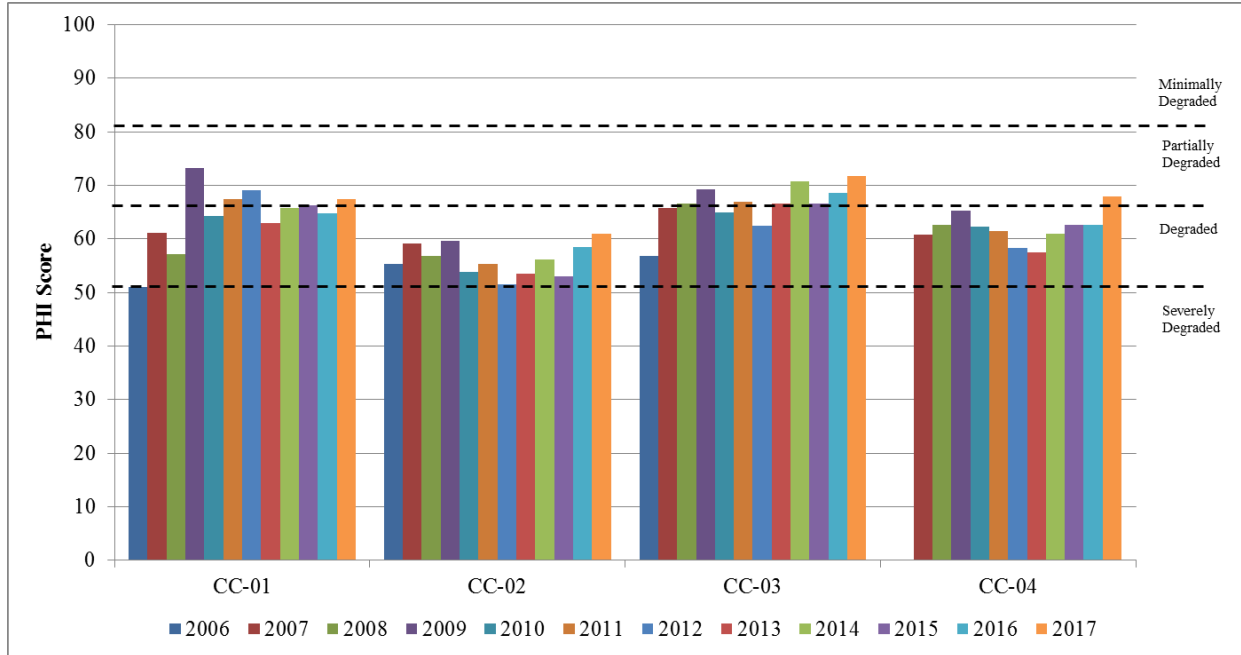


Figure 4-11. Comparison of PHH scores from 2006 to 2017

Beginning in 2013 and 2014, the updated MBSS PHH methods (Paul et al. 2003) were used to calculate PHH instead of the original MBSS methods (Hall et al. 2002) which had been used in the Church Creek watershed reports from prior years. Scores for 2006-2012 shown in Table 4-8 and Figure 4-11 were calculated using the original method, while scores for 2013-2017 were calculated using the updated method.

Biological impairment is evident within this watershed as reflected by the macroinvertebrate communities found throughout the study reach. A comparison of BIBI scores from 2006 through 2017 (Table 4-9) shows no substantial change in biological conditions throughout the study reach. While BIBI scores tend to fluctuate from year to year, overall classifications have changed very little with sites consistently rating either “Poor” or “Very Poor,” no clear trends have been established (Figure 4-12). It appears that the biological community continues to be limited by the presence of urban stressors and degraded physical condition of the stream, and annual shifts in BIBI scores are likely related to random and systematic variability inherent in the assessment process.

| Table 4-9. BIBI scores from 2006 to 2017 | | | | | |
|--|------------|-----------|-----------|-----------|-----------|
| | Site | CC-01 | CC-02 | CC-03 | CC-04 |
| 2006 | BIBI Score | 1.86 | 2.43 | 1.86 | No Data |
| | Rating | Very Poor | Poor | Very Poor | Collected |
| 2007 | BIBI Score | 1.00 | 1.86 | 2.71 | 2.71 |
| | Rating | Very Poor | Very Poor | Poor | Poor |
| 2008 | BIBI Score | 2.43 | 2.43 | 2.43 | 2.14 |
| | Rating | Poor | Poor | Poor | Poor |
| 2009 | BIBI Score | 1.86 | 1.86 | 2.14 | 2.43 |
| | Rating | Very Poor | Very Poor | Poor | Poor |
| 2010 | BIBI Score | 1.29 | 1.86 | 1.57 | 2.14 |
| | Rating | Very Poor | Very Poor | Very Poor | Poor |
| 2011 | BIBI Score | 1.57 | 1.86 | 1.57 | 2.14 |
| | Rating | Very Poor | Very Poor | Very Poor | Poor |
| 2012 | BIBI Score | 1.86 | 2.43 | 1.57 | 2.43 |
| | Rating | Very Poor | Poor | Very Poor | Poor |
| 2013 | BIBI Score | 1.57 | 2.43 | 1.86 | 1.29 |
| | Rating | Very Poor | Poor | Very Poor | Very Poor |
| 2014 | BIBI Score | 1.57 | 1.86 | 1.29 | 1.57 |
| | Rating | Very Poor | Very Poor | Very Poor | Very Poor |
| 2015 | BIBI Score | 1.57 | 1.57 | 2.14 | 1.86 |
| | Rating | Very Poor | Very Poor | Poor | Very Poor |
| 2016 | BIBI Score | 1.86 | 1.57 | 2.14 | 2.71 |
| | Rating | Very Poor | Very Poor | Poor | Poor |
| 2017 | BIBI Score | 2.14 | 2.14 | 2.43 | 1.86 |
| | Rating | Poor | Poor | Poor | Very Poor |

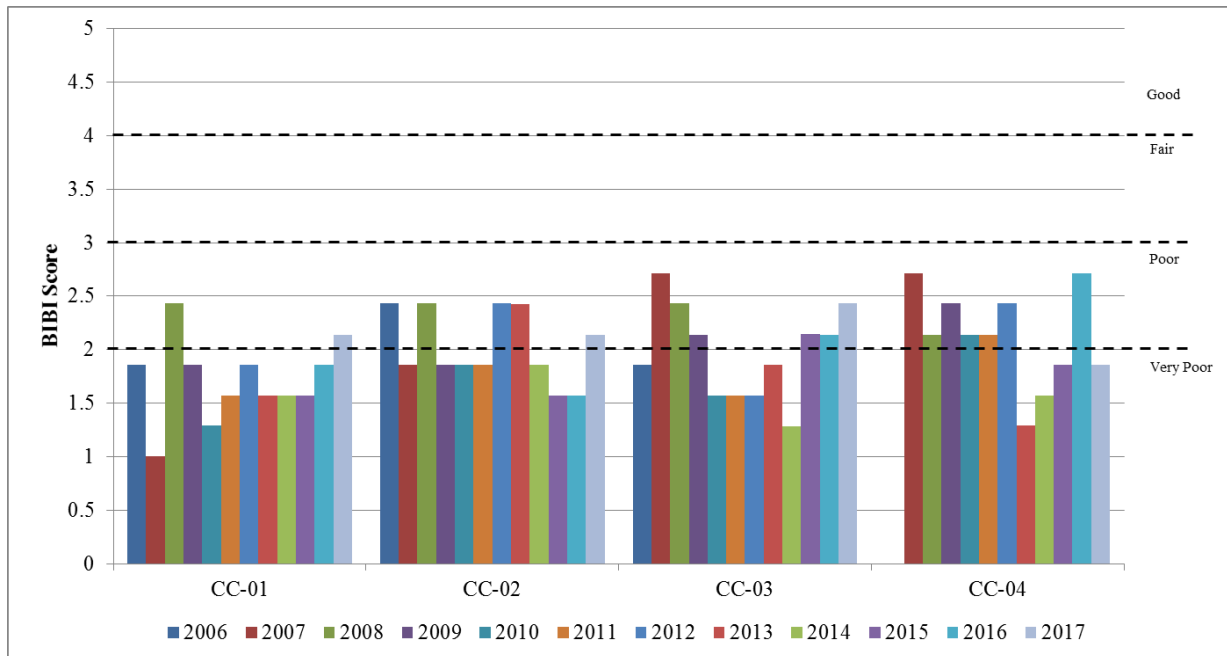


Figure 4-12. Comparison of BIBI scores from 2006 to 2017

4.3 GEOMORPHIC CONDITIONS

The Church Creek study area has a very high percentage of impervious surface cover (approximately 64 percent), and no reach was classified as a C channel, which are generally considered stable stream types due to adequate floodplain connectivity. Four reaches were classified as either F or G channels, which are more entrenched and less stable. The lower end of the Parole Plaza Tributary (XS-4) classified as an E channel and maintains some limited connectivity to its floodplain even though there are significant stormwater inputs feeding into the stream, which typically results in accelerated channel erosion and degradation. Evolution of channel type over the course of the study at each cross-section is presented in Table 4-10. It is likely that current stormwater management and wetland storage on the Church Creek mainstem, as well as the presence of an intact riparian vegetative buffer along much of the stream corridor, contributes to minimizing some of the adverse effects of the high imperviousness in the watershed. Additionally, grade controls such as the culvert at Solomon’s Island Road and cobble rip-rap armoring at XS-5 likely prevent some degradation from occurring in the channel upstream. Nonetheless, there are clear indications of channel instability (i.e., degradation, aggradation, widening) in the upper reaches of the Parole Plaza Tributary, and thus, a need for additional stormwater management to prevent further channel erosion.

Table 4-10. Past Rosgen classifications

| Cross-section | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 | 2012 | 2013 | 2014 | 2015 | 2016 | 2017 |
|----------------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|
| XS-1 | E5 | C5 | E4 | E5 → C5 | E5 → C4/5 | C4/5 → F4/5 | F5 | F4 | F5/4 | F4 | F4 | F4 |
| XS-2 | E5 | E5 | E5 | E5 | E5 | G5c | G5c | G5c | G4c | G4 | G4c | G4c |
| XS-3 | G5c | G5c | G5c | G5c | G5c | No Data | No Data | G4c | G4c | G4/3c | G4c | G4c |
| XS-4 | E5 | E5 | E5 | E5 | E5 | E5 | E5 | C5 | C5 | C5 | E5/4 | E4/5 |
| XS-5 | E5b | C5 | C5 | C5 | C3/5 | C3/5 | C3/5 | F4/3 | F3 | F3/4 | F4 | F4 |

Bankfull channel dimensions (cross-sectional area, width, depth) in the Church Creek study area showed significant departure from expected values, as derived from Maryland Coastal Plain regional relationships of bankfull channel geometry (McCandless 2003). Almost all dimensions were generally larger in the Church Creek study area (see Figures 4-13, 4-14, and 4-15), and were often more similar to relationships of bankfull channel geometry derived from gaged urban watersheds located in the Coastal Plain. These relationships were developed for an urban stream restoration project in Anne Arundel County (AADPW 2002). Values measured in 2017 were slightly higher than prior assessment results. This reflects the higher level of imperviousness in the study area, as compared to the lower impervious levels in the drainage areas used to develop the regional relationship data. The results suggest that this stream has become enlarged as a result of the high imperviousness, and is both wider and deeper than stable C and E type channels located in rural/suburban watersheds of the coastal plain. It should be noted, however, that locating bankfull elevations in the field on actively eroding, previously stabilized, or incising channels is difficult and not recommended due to unreliable and/or misleading indicators, and instead bankfull elevations should be estimated using the aforementioned regional curves (Rosgen, personal communication, May 2011). Where bankfull indicators were suspect or questionable, the indicator approximating the rural/suburban regional curve for bankfull area was used to estimate bankfull elevations. Additionally, the Rosgen method is best used on streams that are free to adjust their lateral boundaries under the current discharge regime experienced by the system (Rosgen 1996). Given the high levels of rip rap and/or concrete rubble armoring found in the reaches containing cross-sections 2, 3 and 5, the accurate determination of the bankfull indicators in the field at these locations is problematic.

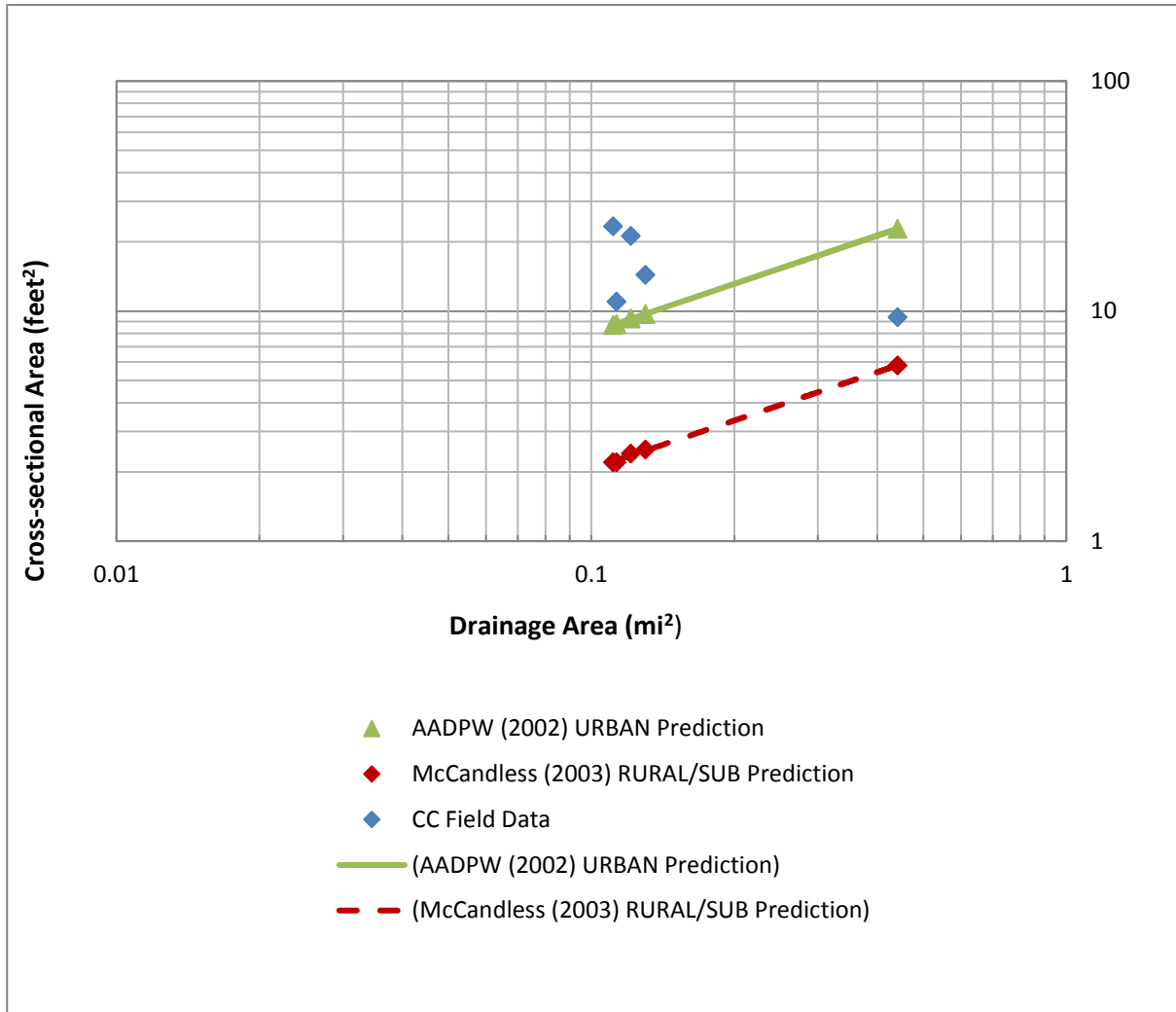


Figure 4-13. Comparison of bankfull channel cross-sectional area to drainage area (CC = Church Creek, 2017 data)

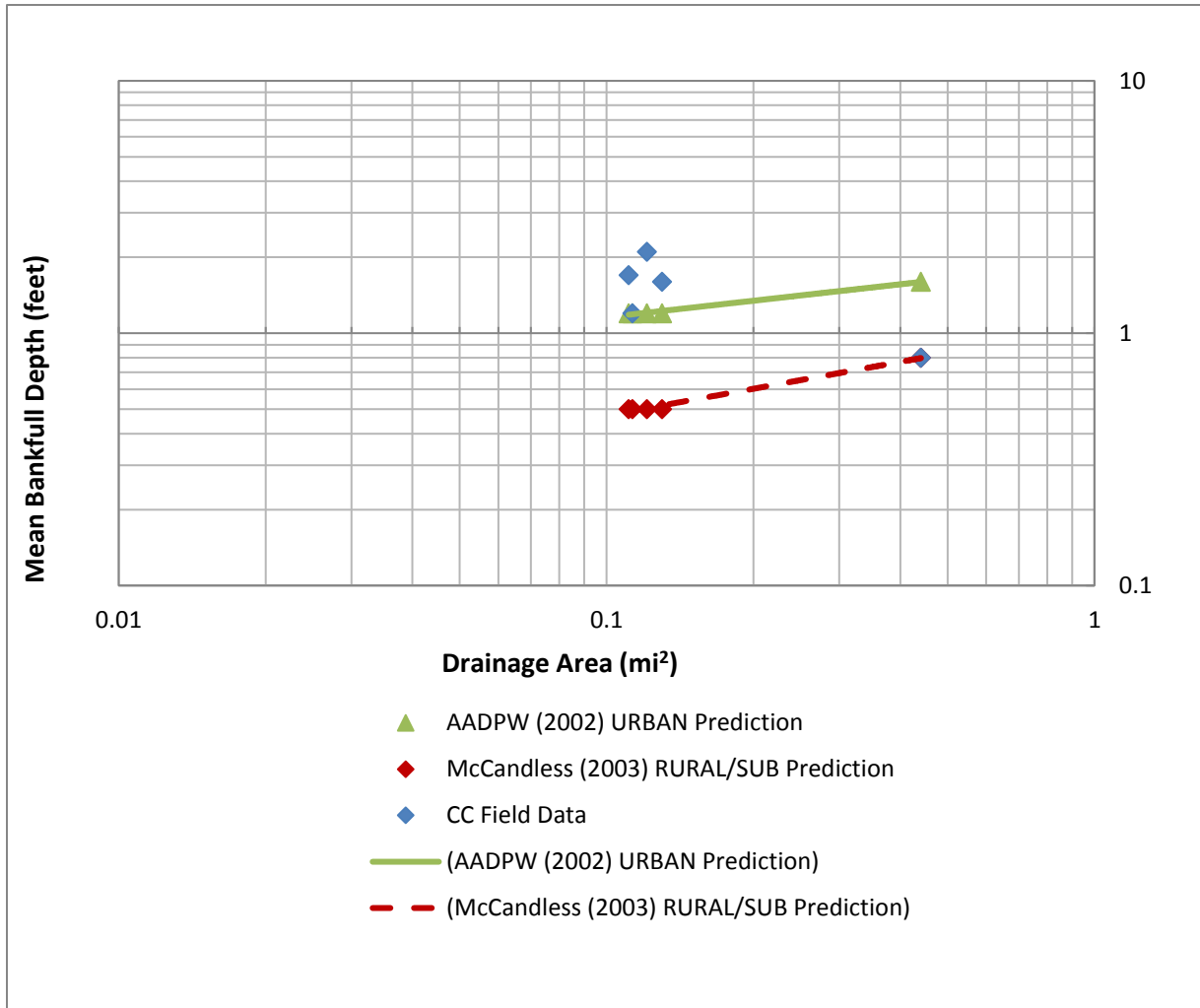


Figure 4-14. Comparison of mean bankfull depth to drainage area (CC = Church Creek, 2017 data)

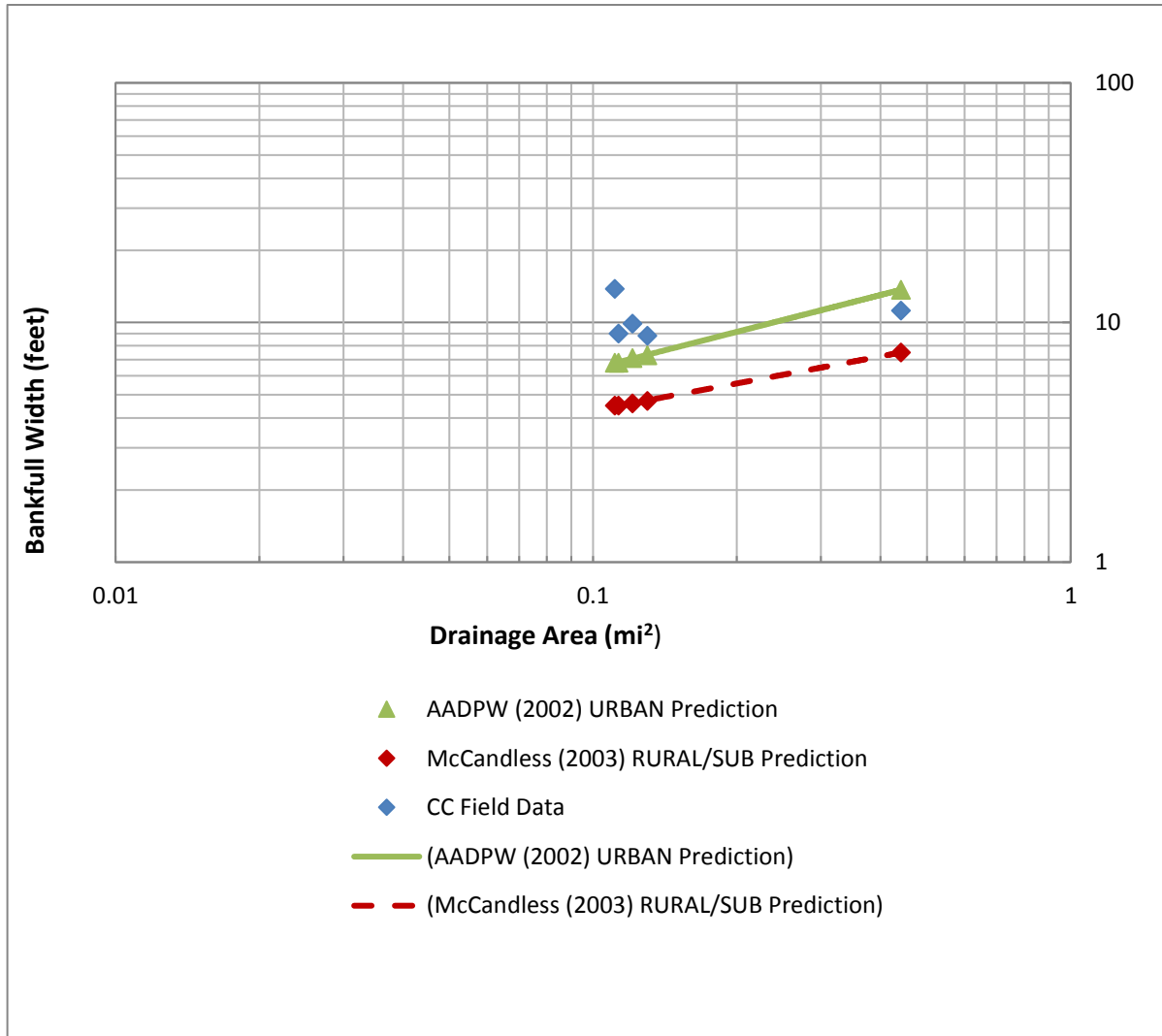


Figure 4-15. Comparison of bankfull width to drainage area (CC = Church Creek, 2017 data)

One of the five cross-sections showed enlargement from channel erosion while the other four showed aggradation as compared to baseline measurements (Table 4-11). Due to the replacement of XS-3 following channel restoration, data were compared to 2007 at this location only, whereas all other comparisons were made to 2003 data. Cross-sectional area from 2011 through 2017 was calculated using the top of bank elevation from the baseline survey in order to standardize comparisons and reduce variability among more subjective bankfull elevation reference points, or even changes that can occur to top of bank elevation from year to year. It is important to note that calculations prior to 2011 did not use the baseline reference elevation, instead they used the corresponding year's top of bank elevation for calculating cross-sectional area, and consequently these values are not directly comparable to the cross-sectional areas reported in 2011 through 2017. Comparison of baseline cross-sectional area is however comparable to 2011 through 2017 since all calculations are made using the same top of bank elevation.

| Table 4-11. Summary of cross-sectional area changes over time. | | | | | |
|---|------|-------|---------------------|------|-------|
| Cross-section ^(a) | XS-1 | XS-2 | XS-3 | XS-4 | XS-5 |
| July 2003 | 16.8 | 8.9 | ND | 14.3 | 9.7 |
| Jan 2005 | 20.7 | 10.0 | ND | 14.4 | 9.9 |
| March 2006 | 19.4 | 8.0 | ND | 18.4 | 9.5 |
| March 2007 | 19.4 | 8.9 | 19.8 | 17.4 | 9.0 |
| May 2008 | 20.1 | 10.1 | 16.7 | 18.0 | 8.9 |
| July 2009 | 19.6 | 9.8 | 21.0 | 15.4 | 8.3 |
| May 2010 | 19.8 | 10.3 | 20.4 | 16.4 | 8.5 |
| July 2011^(b) | 21.3 | 15.9 | 20.6 | 7.8 | 10.5 |
| April 2012^(b) | 21.6 | 15.4 | 19.2 | 11.7 | 5.9 |
| July 2013^(b) | 21.0 | 15.5 | 20.2 | 11.7 | 6.9 |
| June 2014^(b) | 22.4 | 16.2 | 20.6 | 6.8 | 6.7 |
| May 2015^(b) | 22.6 | 16.4 | 18.6 | 9.2 | 6.7 |
| March 2016^(b) | 25.7 | 23.0 | 18.7 | 15.7 | 6.6 |
| February 2017^(b) | 27.1 | 18.7 | 18.2 | 13.3 | 6.5 |
| % Change 2003-2017 | 61.3 | 110.1 | -8.1 ^(c) | -7.0 | -33.0 |
| % Change 2011-2017 | 27.2 | 17.6 | -11.7 | 70.5 | -38.1 |
| ^(a) All values listed here are for top of bank area and are listed in square feet ^(b) Values obtained using reference elevations (top of bank) from baseline measurements ^(c) % change from 2007 ND = No Data | | | | | |

Using the current reference elevation comparison method, the upstream cross-sections (XS-1 and XS-2) showed fairly substantial enlargement over time, with increases of approximately 61.3%, and 110.1% respectively, since baseline measurements began in 2003. The bed elevation at XS-1 appears to have dropped almost a foot since 2003 with a noticeable amount of bed scour occurring between 2014 and 2017 (Appendix F). Scouring near the right bank occurred between 2008 and 2009 but has remained stable since. The left bank however, has both widened and deepened since 2012 but looks to be stabilizing as of 2017. The channel at XS-2 has widened notably since 2003, with considerable erosion along the right bank. The left bank has been generally stable showing minimal erosion until 2016. In 2016 the channel had both widened along the left bank and deepened mid channel, although in 2017 the channel returned to more narrow and shallow conditions seen before 2016. Cross-section area comparisons between baseline and 2017 show a substantial increase with a moderate percent change occurring over the last five years of 27.2% increase at XS-1 and 17.6% increase at XS-2.

Cross-section XS-3 has had very minimal changes in cross-sectional area with just a 8.1% decrease since 2007 baseline measurements and -11.7% change between 2011 and 2017. Between 2009 and 2011, the XS-3 channel appeared to be enlarging, as the right bank and bottom of the right bank experienced some erosion and the cross-sectional area increased (Appendix F). However, during the past five years, the right bank has experienced some

aggradation (Appendix F). Between 2012 and 2016 monitoring, there has been little change apart from slight aggradation across the stream bed and toe of the right bank. In 2017 erosion began occurring behind the armored right bank and some scouring was evident on both sides of the channel bed. Cross-section XS-3 continues to have yard waste (i.e., grass clippings, leaves, and branches) dumped along the left bank.

Cross-section XS-4 has had the most variation throughout the years. Between 2010 and 2011 cross-section XS-4 had shown moderate signs of aggradation. Within the next year, the channel experienced erosion of the bed, particularly along the right hand side of the stream. In the 2013 survey, signs of aggradation were again present and the stream bed characteristics resemble those of the 2011 survey. In 2014 the stream bed remained elevated as in 2011 and 2013 however there was slight widening along the right bank. The debris jam at XS-4 which formed between 2011 and 2012 and caused sediment accumulation, was removed during stream restoration construction prior to the 2016 surveying. Consequently, the channel scoured significantly and resulted in cross-sectional area increase. This scour slowed by 2017 and the cross-sectional area decreased slightly from 2016. Cross-section XS-4 continues to be larger than in 2011 (70.5%) but is 7.0% smaller than 2003.

Cross-section XS-5 has been armored with cobble-sized rip rap in its bed to protect the sanitary sewer line. Between 2012 and 2013, XS-5 appears to have eroded by several inches of sediment, most notably near the left bank. The cross-sectional area has decreased by 38.1% since 2011. During the past three years, however, there has been little change in both stream bed elevation and bank stability (Appendix F). Cross-sectional area has remained relatively stable from 2014 to 2017 with little to no change year to year.

4.4 GENERAL CONCLUSIONS

Based upon the data collected in 2017, most stream water quality parameters measured have improved when compared to pre-construction and earlier post-construction monitoring years, but biological and physical conditions within the Church Creek study area have not improved and remain in a degraded and impaired condition. Although the stream channel has been stabilized along several reaches, the effects on biota are yet to be seen from such efforts. In 2016, stream restoration occurred downstream of XS-4, on an unnamed tributary of Church Creek, and upstream of XS-5 on the mainstem Church Creek. All of the CC-04 and part of the CC-03 biological monitoring sites were within the restored reach of stream. At Church Creek, annual average EMCs for most pollutants and annual loads for all pollutants were lower in 2017 compared to 2016. The reduced loadings and EMCs may have been the result of stream restoration; however, given the size of the restored area in relation to the overall watershed, water quality improvement may be difficult to discern from natural variations in pollutant levels, especially in the short timespan in which post-restoration data are available. The reduction in loads of pollutants at Church Creek may also be due, in part, to a 17% decline in total discharge as a result of lower annual rainfall between 2016 and 2017. Over time, the restoration project should result in less sediment transported downstream, increased stability at physical monitoring stations, and could positively affect the biota at monitoring stations through habitat improvement. In the year since restoration was completed, cross-section 5, downstream of the

restored reach has maintained stability in its geomorphic parameters including consistent cross-sectional area. No considerable changes have been observed in water chemistry, biota, or physical habitat downstream of the restored reach; while increased categorical changes in PHI were observed in 2017, these categorical ratings were still within the historical rankings for the sites. Future monitoring efforts will be used to evaluate the effects of this restoration.

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APPENDIX A

**STORM EVENT HYDROGRAPHS, NARRATIVES AND COMPOSITE
SAMPLING METHOD TECHNICAL MEMORANDUM**



TECHNICAL MEMORANDUM

TO: Janis Markusic, AACO DPW
FROM: James Tomlinson
DATE: 5/12/08
SUBJECT: Proposed Modifications to Sampling Procedures
Church Creek/Parole Plaza NPDES Monitoring 2008
KCI Job Order No. 01-032333.38

Dear Ms. Markusic,

Monitoring at the Parole Plaza station has historically entailed collecting storm water samples from a 60" corrugated metal pipe (CMP) outfall draining the Parole Plaza site. The rising, peak, and falling limb of each storm event was sampled and analyzed in the lab. In the summer of 2007, a newly-installed 54" reinforced concrete pipe (RCP) was connected to the Church Creek drainage network (refer to Figure 1). The new 54" RCP drains the *Annapolis Towne Center at Parole* construction site and contains portions of the drainage that had historically been passing through the 60" CMP. Based on plan reviews, on-site discussions with County staff /inspectors, and discussions with Mr. Matthew Kelly of Greenhorne & O'Mara Inc. (Head Engineer) it was determined that runoff from the *Annapolis Towne Center at Parole* site would be discharging from both the 60" CMP and 54" RCP outfalls under post-construction conditions.

In order to maintain consistency in the characterization of the watershed, samples are being collected from both pipes. As in the 2006-2007 reporting year, the 2007-2008 reporting year has also encountered the challenge of determining the active drainage configuration during each sampled event. To address this problem, samples were collected for each limb of an event and taken to the laboratory for each pipe separately. In 2007, storm samples were collected and analyzed for both outfall pipes for three storm events (9/11/07, 10/9/07, and 10/19/07) at the Parole Plaza station. In 2008, storm samples were collected and analyzed for both outfall pipes for two storm events (1/10/08 and 1/17/08) at the Parole Plaza station.

Stage/discharge relationships were developed for both pipes, to determine the discharge from each pipe based on field-measured depths. The rating tables were used to determine discharges and total flow volumes for each limb of the storm. The reported concentrations for each limb were then weighted by the flow for that limb of the storm to determine the event mean concentration for each pipe. The event mean concentrations for each pipe were then flow weighted again for the total discharge through each pipe to determine the total event mean concentration for the *Annapolis Towne Center at Parole* site.

In March 2008, the sampling methodology was modified such that only one flow-weighted composite sample per event limb is collected and analyzed at the lab. The goal of compositing the discharge from both pipes is to ensure a comprehensive analysis of the storm water leaving the *Annapolis Towne Center at Parole* site during and after construction. The flow-weighted composite sample for each limb of an event is determined by the combined discharge of the 60" CMP and 54"RCP outfalls. Stage/discharge relationships were developed for both pipes, to determine the discharge from each pipe based on field-measured depths. The relationships are based on modeled values and are currently being validated by field-measured data. A sample

Proposed Modifications to Sampling Procedures
Church Creek/Parole Plaza NPDES Monitoring 2008
KCI Job Order No. 01-032333.38

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5/12/08

is collected for each limb of a storm by proportioning the appropriate volume of RCP/CMP sample into a sample container, based on each pipes contribution to the total runoff. To determine the appropriate amount of sample to collect from each pipe for each limb, the discharge from the CMP and RCP is computed and totaled. The CMP and RCP discharge is then divided by the total discharge to determine the percent contribution to the total discharge. Once the percent contribution is determined, then volumetric containers are utilized to proportion each pipe sample to create one composite sample. The composite sample is then distributed into the laboratory bottles.

At this time, three storm events (3/19/08, 4/20/08, and 4/26/08) have been sampled utilizing the flow-weighted composite method. Currently, no automated equipment is available for the RCP. Level, pH, and temperature are taken manually during sampled events. All of the flow data and discharge characteristics for each outfall pipe are recorded during the storm event into a laptop spreadsheet that automatically calculates the correct sample proportions. General notes, field-measured velocity data, and a hydrograph are also recorded on the laptop spreadsheet during the storm event (see Figure 2).

As discussed in the field on April 10, 2008, KCI feels that this is an appropriate method for maintaining consistency with the historical data, while reducing the costs of sampling and testing both outfalls individually.

If you have any questions or comments, please feel free to contact me at the number below.

Sincerely,

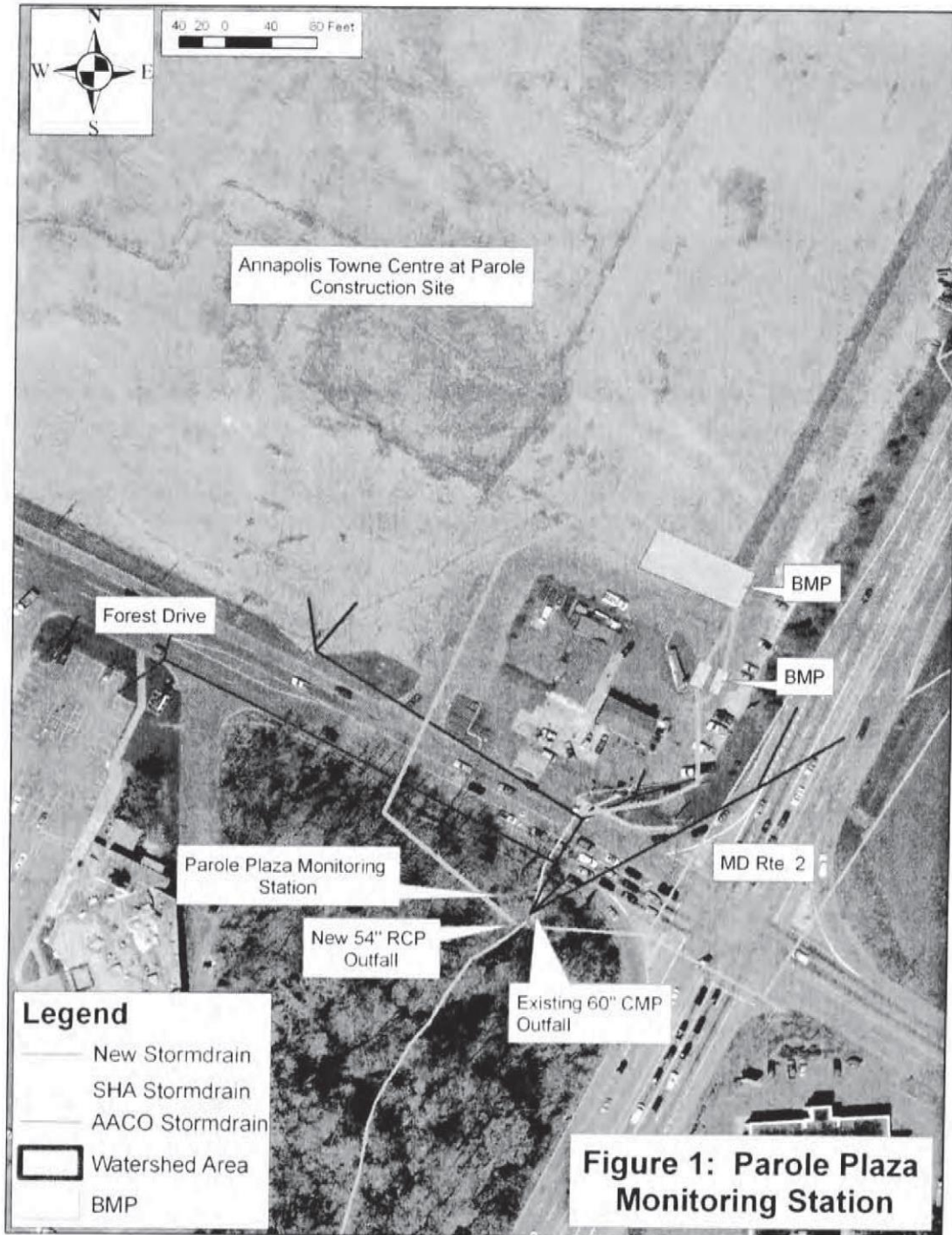


James A. Tomlinson, PE
Project Manager
(410) 316-7864

ND/jt

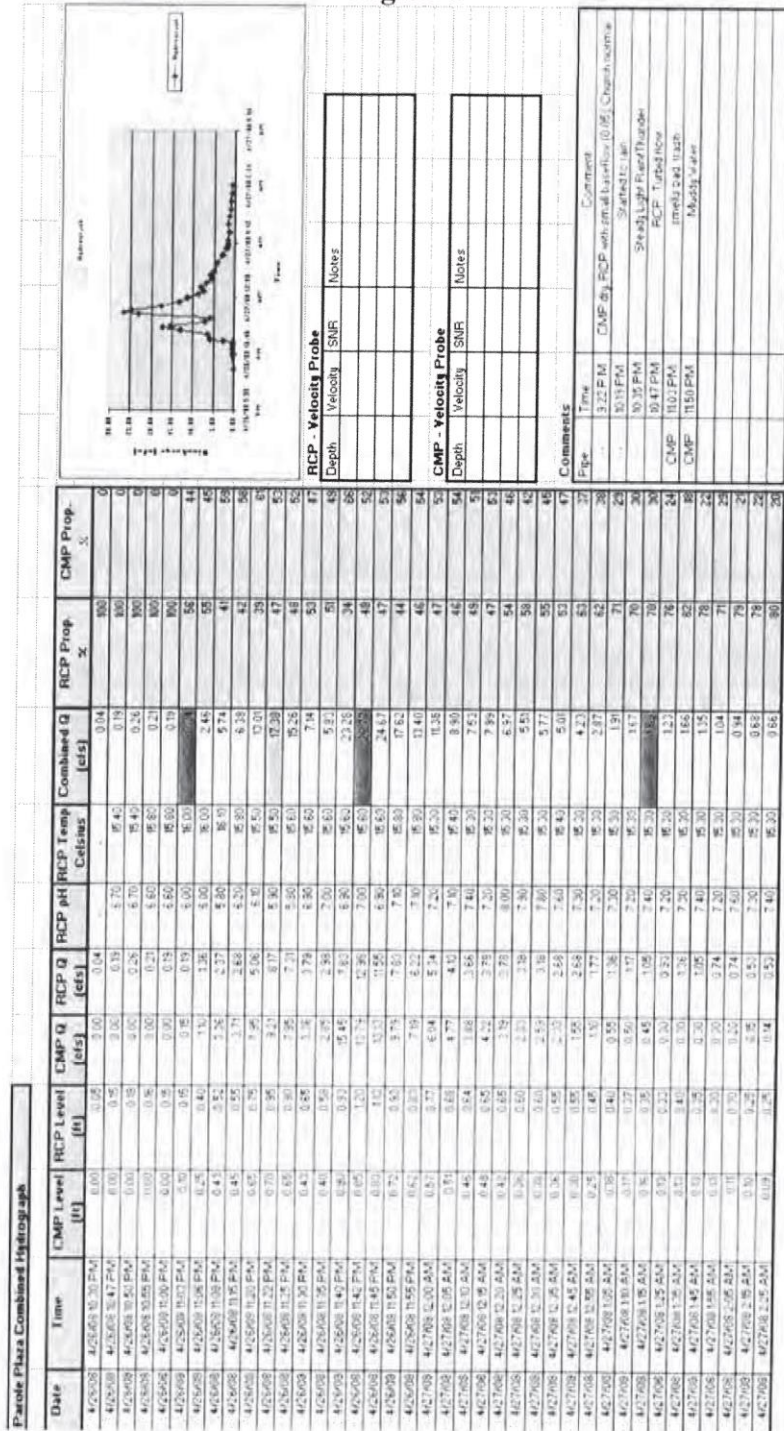
cc: Christopher Victoria, AACo DPW
Nathan Drescher/KCI

Proposed Modifications to Sampling Procedures
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Proposed Modifications to Sampling Procedures
 Church Creek/Parole Plaza NPDES Monitoring 2008
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Figure 2



Storm Event Narratives

- July 28, 2016 - At Church Creek, only four of the parameter EMCs were greater than the average concentrations of the storms captured since December 12, 2012. The EMC of BOD exceeded its long-term historical average by 142%, which included a peak limb concentration of 20 mg/L. The other three parameter EMCs (total Kjeldahl nitrogen, nitrate-nitrite, and lead) exceeded their historical average values for storm runoff by smaller percentages. Zinc had high concentrations during the rising limb (217 µg/L) and peak limb (122 µg/L) of this storm, however neither concentration exceeded the threshold to trigger reanalysis by the laboratory. Three of the parameter EMCs at Parole Plaza during the storm were greater than the corresponding average concentrations of the storms captured since December 12, 2012. The EMCs of BOD total Kjeldahl nitrogen, and zinc exceeded their historical averages by 46%, 29.8%, and 3.3%, respectively. The rising limb concentrations of BOD (25 mg/L) and total Kjeldahl nitrogen (2.7 mg/L) were the cause of above-average EMCs for this storm. Individual zinc concentrations were higher than the long-term average during the rising and peak limbs; however, the EMC was generally comparable to the historical average value.
- September 19, 2016 Storm - At Church Creek, field staff inadvertently transferred sample water into the incorrect set of bottles. Field staff went back to the site, and poured the remaining volume into the correct sample bottles, but there was insufficient volume to analyze for hardness. At Church Creek, three parameter EMCs were greater than the average concentrations of the storms captured since December 12, 2012: TSS (41.9%), lead (61.7%), and *E. coli* (132.9%). The higher than average EMCs for TSS and *E. coli* resulted from the higher than average sample concentrations during the rising and peak, and all three limbs, respectively. For lead, the concentrations during each limb were comparable to each other, but were all above the long-term average. At Parole Plaza, field staff noted an oily odor coming from the CMP during the event. The EMCs of only two parameters, TSS (9.0%) and total phosphorus (30.4%), were greater than the corresponding average concentrations of the storms captured since December 12, 2012. Total phosphorus and TSS concentrations during the rising and peak limbs were higher than the applicable long-term averages and contributed to the overall higher EMC values. No concentrations of these analytes triggered reanalysis by the laboratory.
- September 27, 2016 Storm - Total Kjeldahl nitrogen, TSS, lead, and *E. coli* EMCs were greater than the average concentrations of the storms captured since December 12, 2012 at Church Creek. The EMC for *E. coli* was greater than the corresponding historical average by 194%. During the rising limb, the *E. coli* concentration was 91,000 MPN, which was the highest concentration recorded during the last four years of sampling at Church Creek. The total phosphorus EMC was similar to the historical average value. Similar to Church Creek, the *E. coli* EMC at Parole Plaza was higher than the average concentration of storms captured since December 12, 2012. However, individual limb concentrations did not include record concentrations as was the case at Church Creek. The total Kjeldahl nitrogen EMC was similar to (i.e., 4.5% higher than) the historical average value.

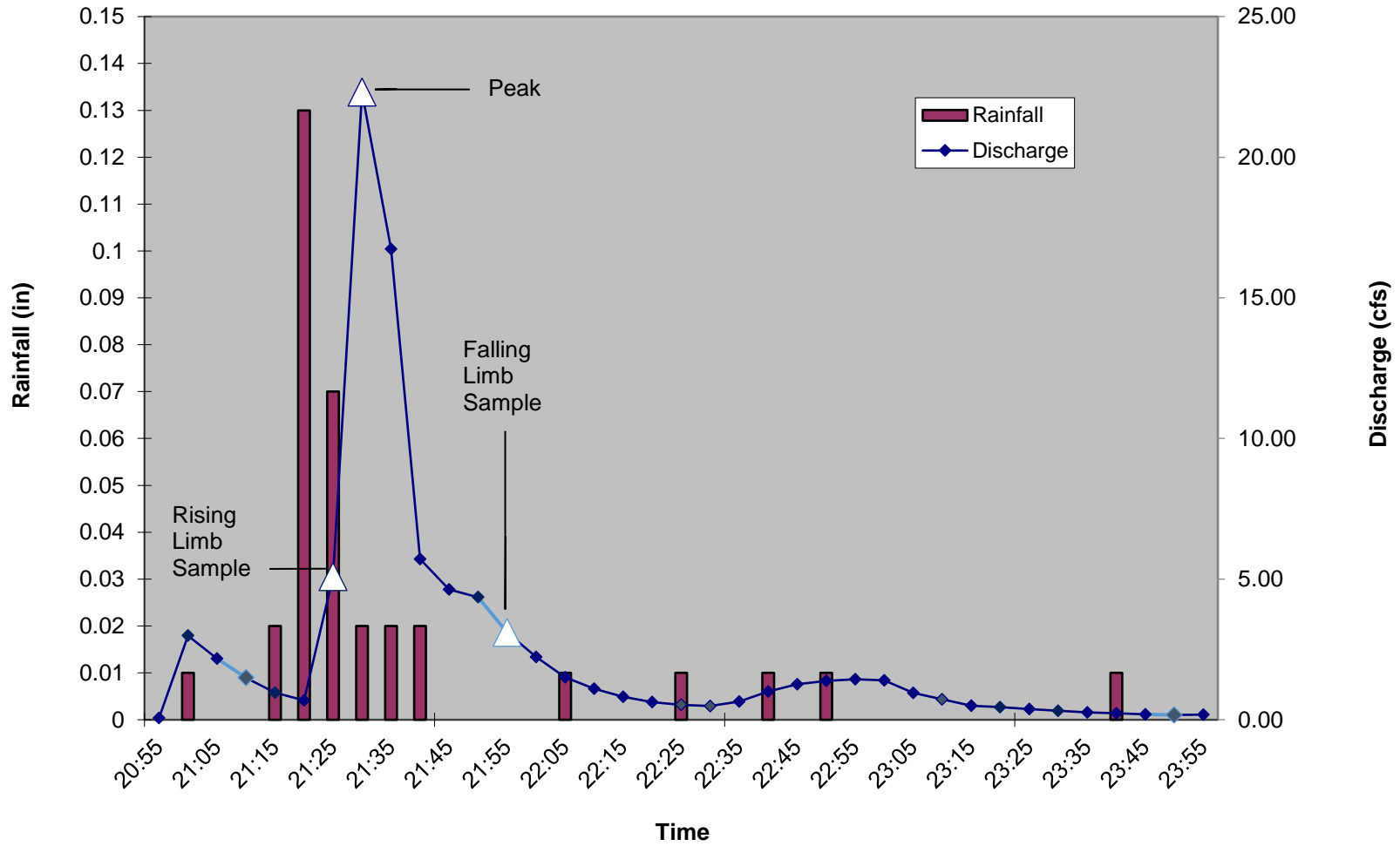
- November 9, 2017 - Six of the parameter EMCs were greater than the long-term average concentrations of the storms captured since December 12, 2012, at Church Creek. The EMCs for total Kjeldahl nitrogen (TKN), nitrate-nitrite, total phosphorus, and hardness exceeded their long-term historical averages by over 100%. The rising limb concentrations for these parameters were possibly the main factor in the above-average EMCs for this storm. At Parole Plaza, five of the parameter EMCs during the storm were greater than the corresponding average concentrations of the storms captured since December 12, 2012. The EMCs of BOD, TKN, total phosphorus, and zinc exceeded their historical averages by a small percentage. The unusual high concentrations of lead during all three limbs at Parole Plaza and the blank samples caused doubt in the results therefore they have been removed from the EMC spreadsheet for both sites. Also during this storm event, field staff noticed the smell of cooking oil coming from the RCP.
- November 29, 2016 Storm - Because October and November were very dry months with a total rainfall of only 2.8 inches, this storm was accepted by Anne Arundel County even though it did not meet the 0.1" minimum rainfall requirement. At Church Creek, only two parameter EMCs were greater than the average concentrations of the storms captured since December 12, 2012: nitrate-nitrite (56.3%), and total phosphorus (108.6%). The higher than average EMC for total phosphorus resulted from the higher than average sample concentration during the peak limb. The individual concentration of 2.60 mg/L for total phosphorus during the peak limb was the highest since December 12, 2012. During this storm event, at Parole Plaza a large amount of leaves flowed through the RCP, and affected the measured level and the EMCs calculations; therefore no EMC calculations were completed for this storm at Parole Plaza. During the storm, field staff continually cleared the leaves from the RCP to maintain accurate flow measurements. Concentrations of TPH, zinc and BOD during the rising and peak limbs were higher than the detection limit. The *E. coli* concentration was very high during the peak limb.
- December 6, 2016 Storm - None of the parameter EMCs were greater than the average concentrations of the storms captured since December 12, 2012 at both stations. At Church Creek, most concentrations were higher during the peak limb than the rising and falling limbs, but were not high enough to cause EMCs to exceed long-term averages. At Parole Plaza, the rising limb concentrations of most parameters were higher than the peak and falling limbs. None of the individual concentrations during this event were identified as a concern.
- January 23, 2017 Storm – At Church Creek, five of the parameter EMCs were greater than the long-term average concentrations of the storms captured since December 12, 2012. The EMCs for total metals (copper, lead, and zinc) exceeded their long-term historical averages by over 100%. The highest concentrations of metals during this event occurred during the peak limb. The TSS EMC exceeded its long-term historical average (by less than 25%), which may have contributed to the higher than average concentrations of the metals. At Parole Plaza, EMCs for the same five parameters as Church Creek, plus phosphorus, were greater than the corresponding average concentrations of the storms captured since December 12, 2012. The TSS and lead EMCs exceeded their historical averages by over 100%. Maximum concentrations for both

parameters occurred during the peak limb. During this storm event, field staff noticed that the water discharging from the pipes was gray and smelled like gasoline and oil; however, TPH results were below detection limits.

- February 28, 2017 Storm - At Church Creek, all parameter EMCs were less than the average concentrations of the storms captured since December 12, 2012. During this event, duplicate samples were collected at Church Creek during the falling limb. All of the results were within 20% of each other. At Parole Plaza, the EMC of only one parameter (TPH) was greater (69.8%) than the corresponding average concentration of the storms captured since December 12, 2012. Specifically, concentrations of TPH during the rising and peak limbs were higher than the detection limit. During this storm, the field crew inadvertently failed to program the In-situ sonde to record pH from the CMP; therefore, the pH readings on the EMC spreadsheet are only from the RCP for this event.
- March 31, 2017 Storm - During the March 31, 2017 event, the copper EMC was greater than the average concentration of the storms captured since December 12, 2012 at Church Creek, but only by 7.5%. Concentrations of copper were higher than the detection limit during all three limbs which contributed to the overall higher EMC value, but none exceeded concentrations that would be considered a concern. During this storm, at Parole Plaza, the same parameters from the January 23, 2017 storm had greater EMCs than the corresponding average concentrations of the storms captured since December 12, 2012. Total phosphorus, TSS, copper, lead, zinc, and hardness exceeded their historical averages by 6.3% to 34.1%.
- April 25, 2017 - During this storm event, *E. coli* samples were now being analyzed by new laboratory (Water Testing Labs of Maryland). Since it was the first time analyzing the *E. coli* samples they were not aware that the samples needed to be diluted; therefore, none of the samples were diluted and none of results were included in the EMC calculations. At Church Creek, the only parameter EMC that was greater than the average concentrations of storms captured since December 12, 2012 was copper. The EMC of copper was greater than the corresponding historical averages by 32.8%. At Parole Plaza, five parameter EMCs (TKN, total phosphorus, copper, zinc, and TPH) were greater than the average concentrations of the storms captured since December 12, 2012. The parameters greater than the corresponding historical averages ranged between 12.85 to 52.7%. The TPH concentration was 7 mg/L during the rising and peak limb of the storm event causing the EMC to increase.
- May 12, 2017 - None of parameter EMCs were greater than the average concentrations of the storms captured since December 12, 2012 at both stations. None of the parameter concentrations were unusually high or stood out as a problem at either station.
- June 19, 2017 - At Church Creek, six parameter EMCs were greater than the average concentrations of the storms captured since December 12, 2012. EMCs for BOD, TSS, copper, lead, zinc and *E. coli* were greater than the corresponding historical averages by 18.5% to 226.2%. The concentration of metals were elevated during the rising limb which contributed to the higher EMCs; have been requested to be reanalyzed by the

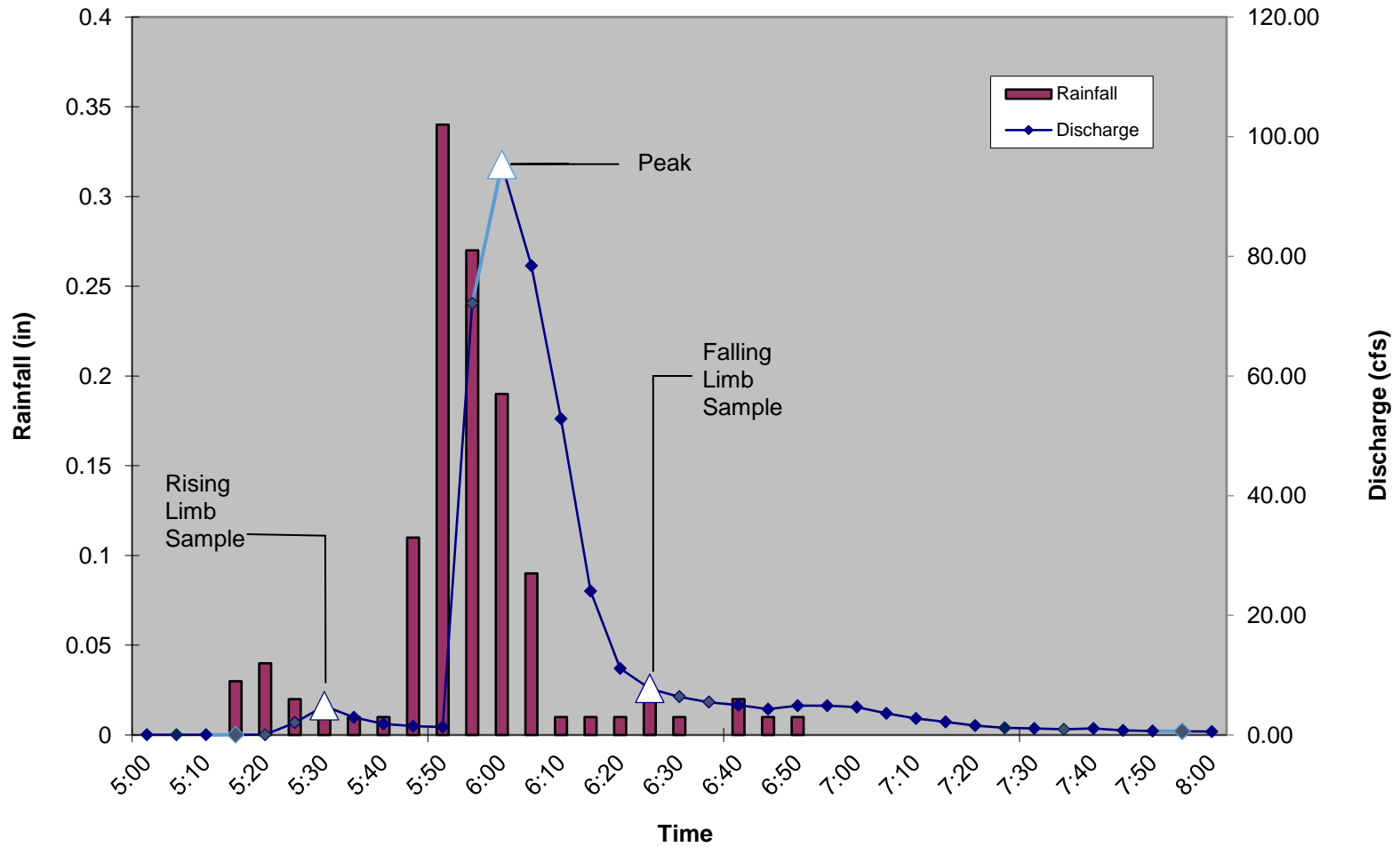
Martel along with TPH because they are over their parameter threshold for the laboratory. Copper, zinc, and TPH EMCs were also greater than the average concentrations of the storms captured since December 12, 2012 at Parole Plaza which can be explained by high concentrations during the rising limb. The EMC for TPH was greater than the corresponding historical averages by 420.7%. All of these parameters have been requested to be reanalyzed as well.

Hydrograph for July 28, 2016 Storm Parole Plaza



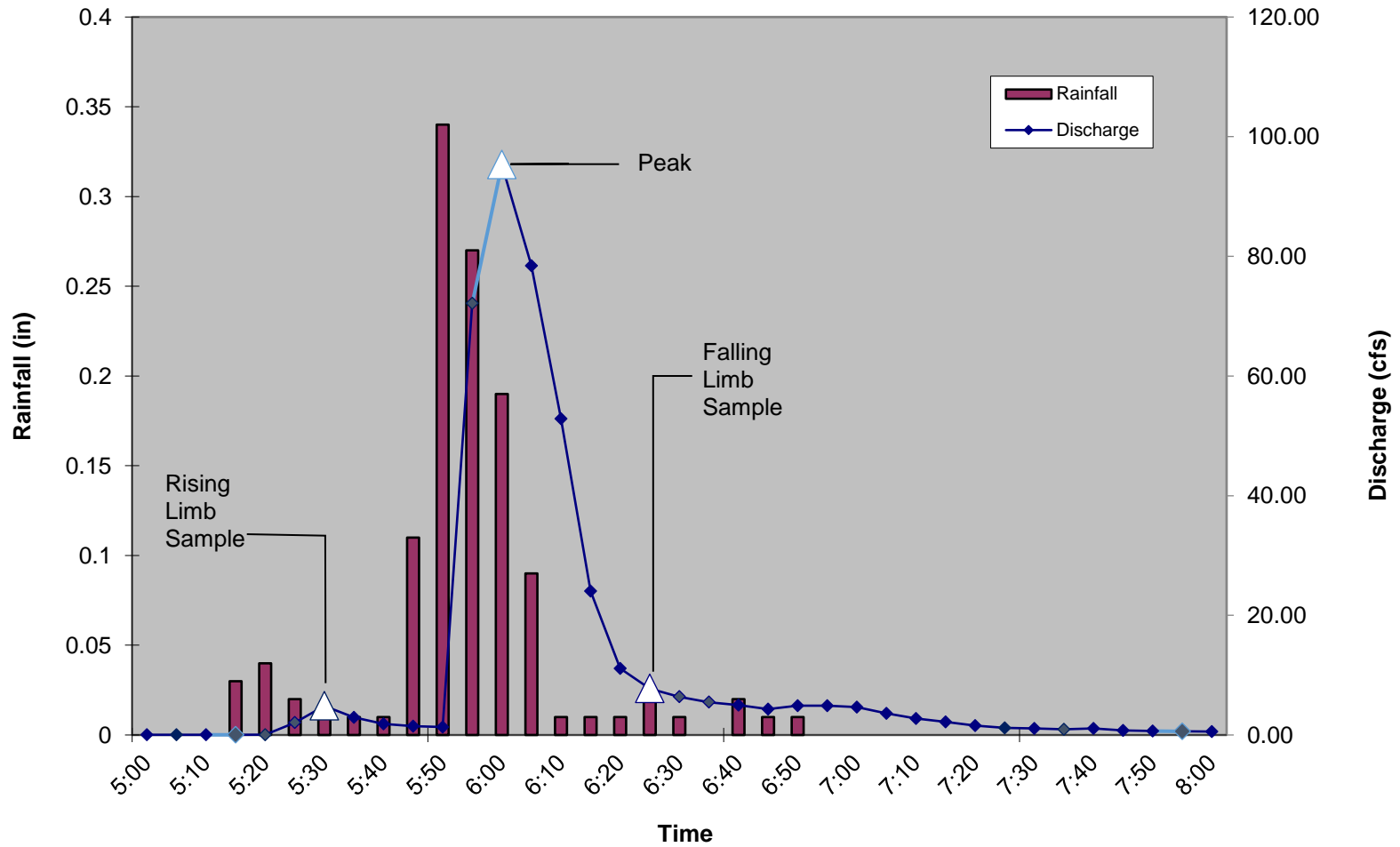
Hydrograph for September 19, 2016 Storm Parole Plaza

A-12

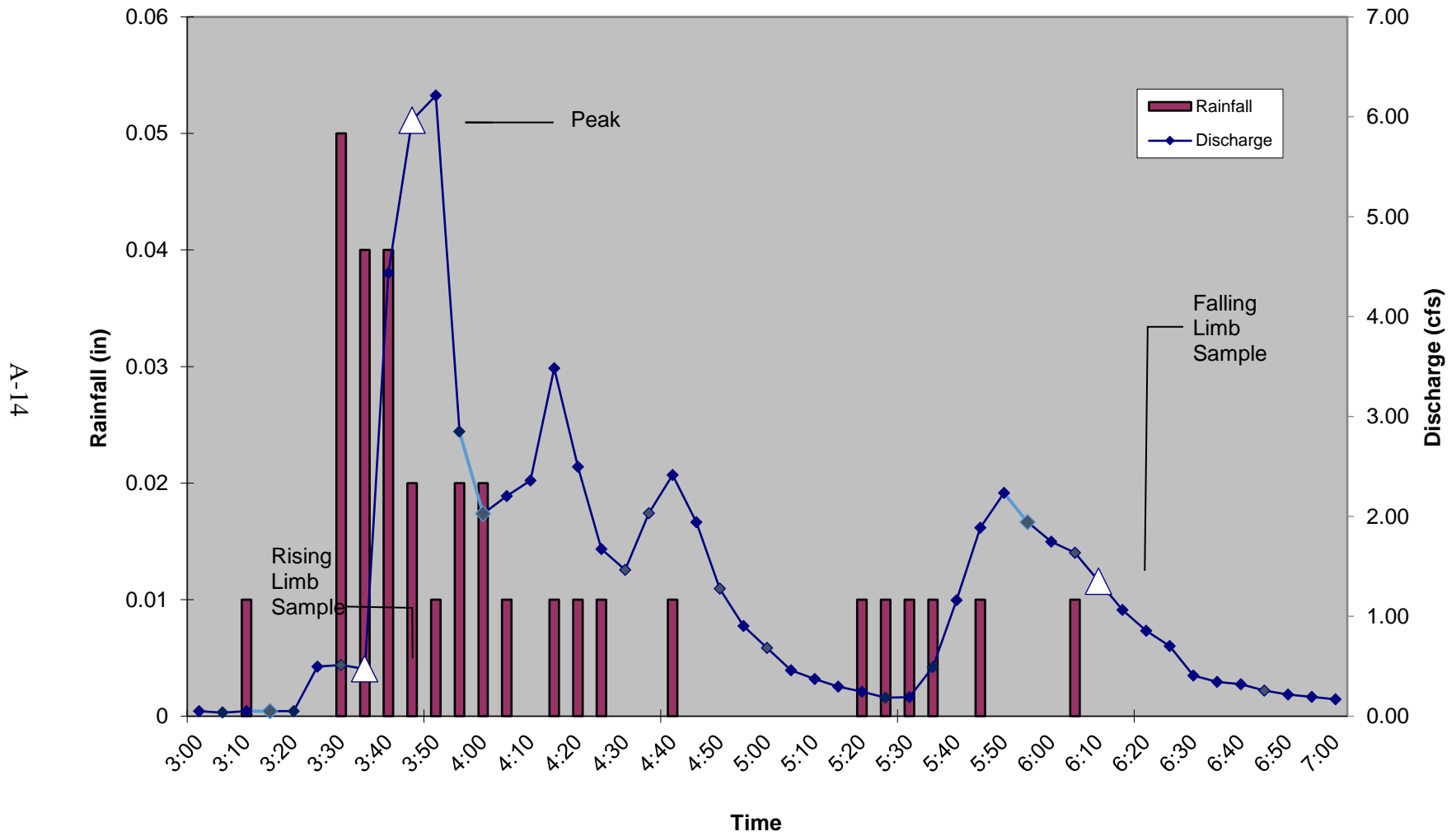


Hydrograph for September 19, 2016 Storm Parole Plaza

A-13

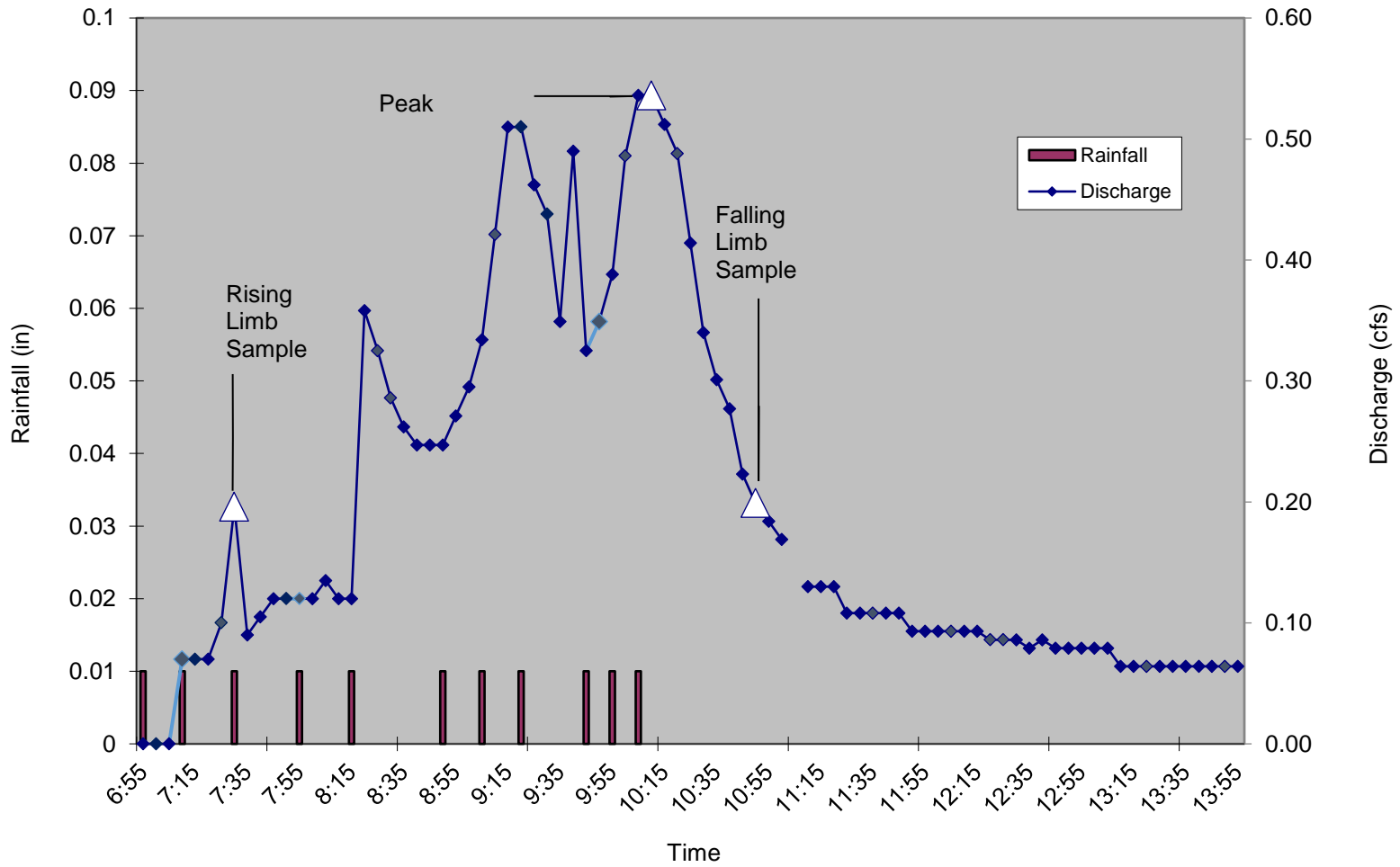


Hydrograph for September 27, 2016 Storm Parole Plaza

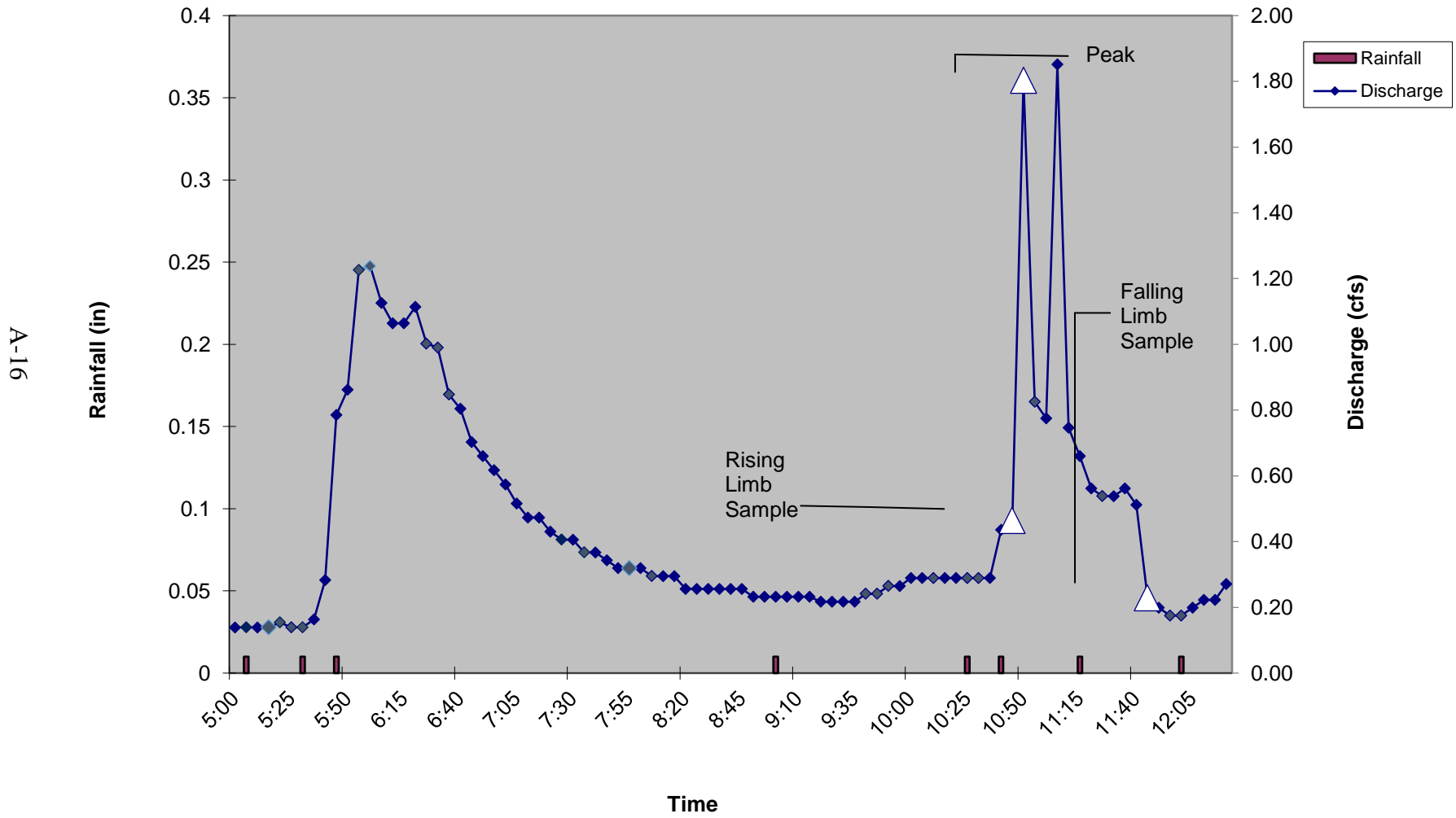


Hydrograph for November 9, 2016 Storm Parole Plaza

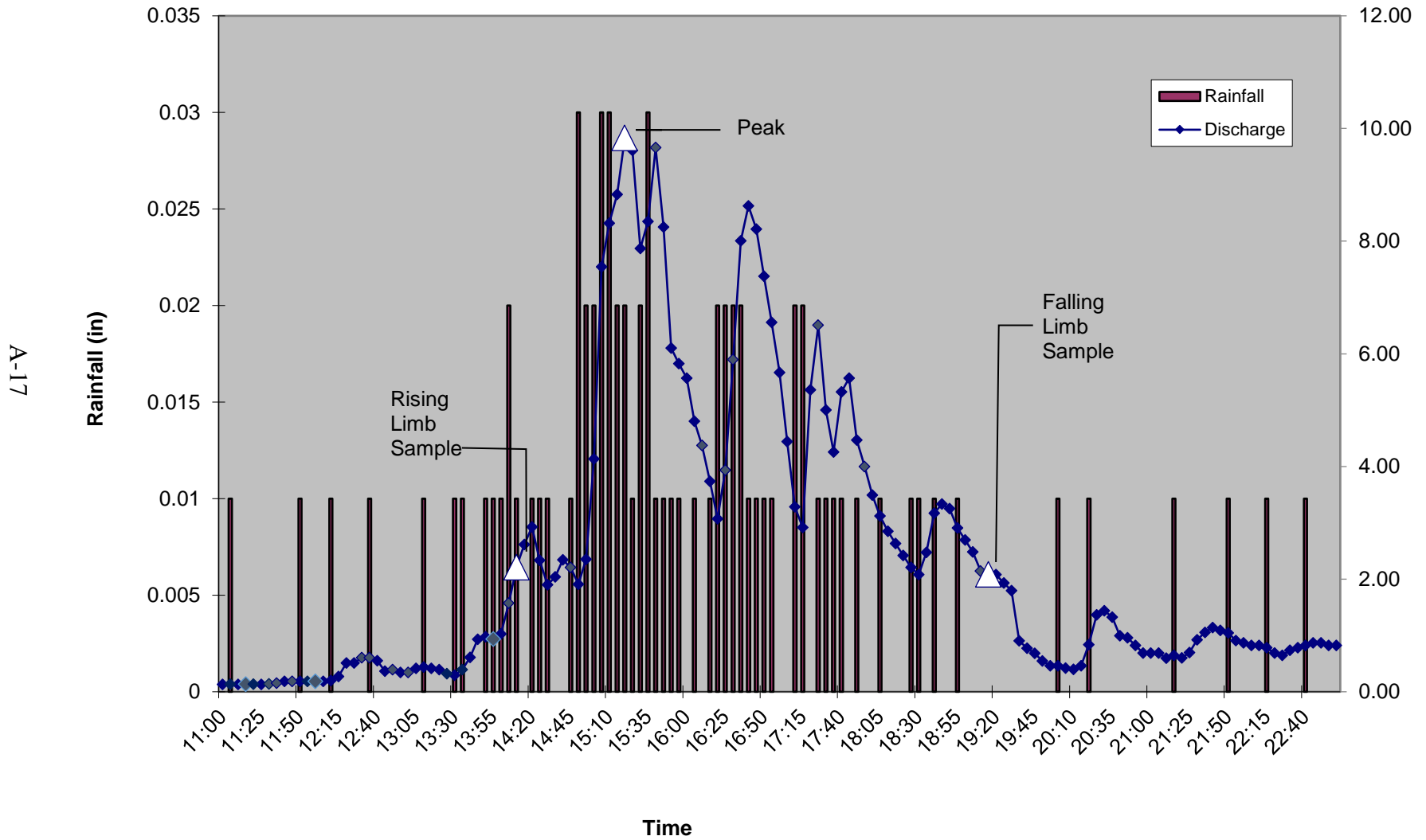
A-15



Hydrograph for November 29, 2016 Storm Parole Plaza

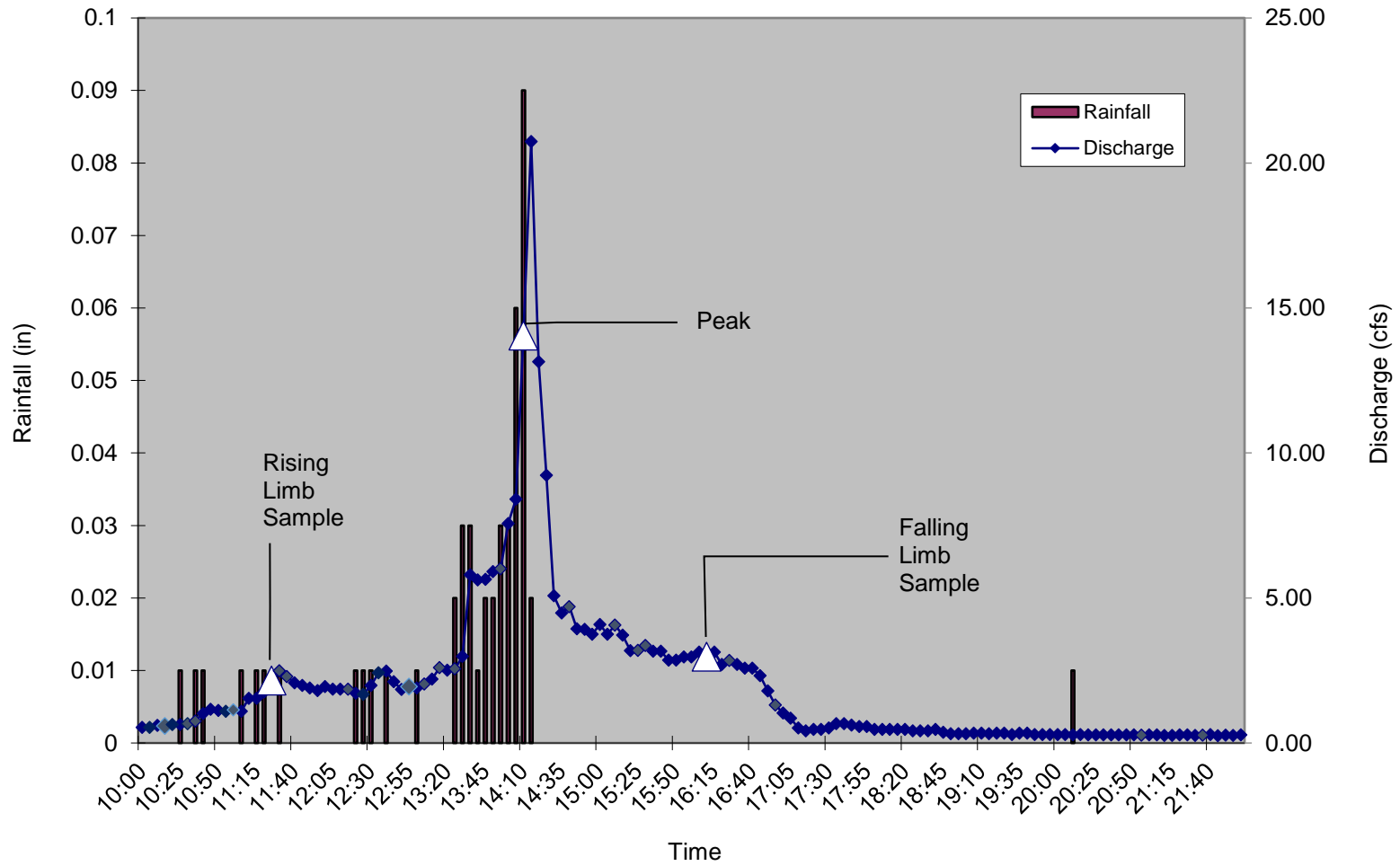


Hydrograph for December 6, 2016 Storm Parole Plaza

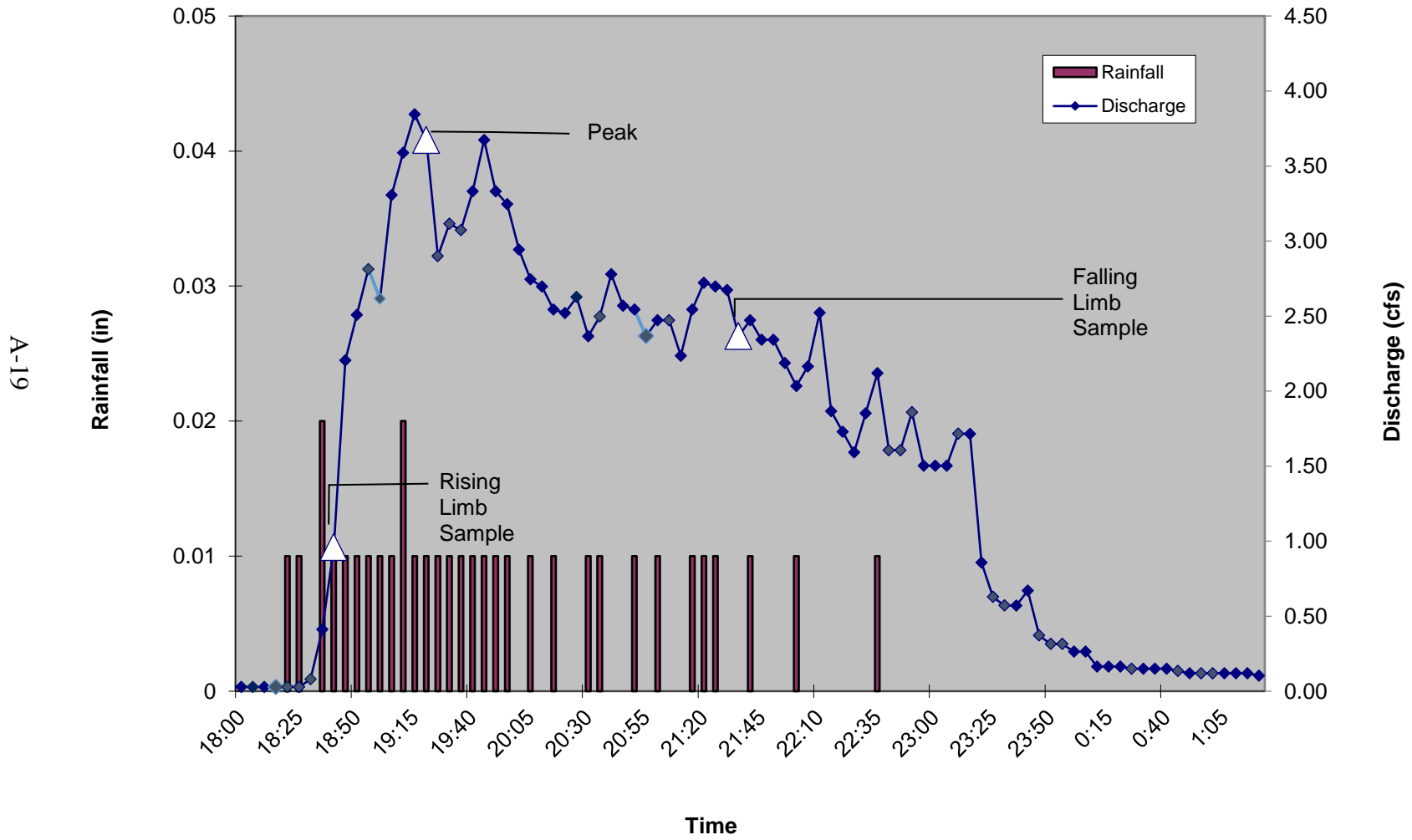


Hydrograph for January 23, 2017 Storm Parole Plaza

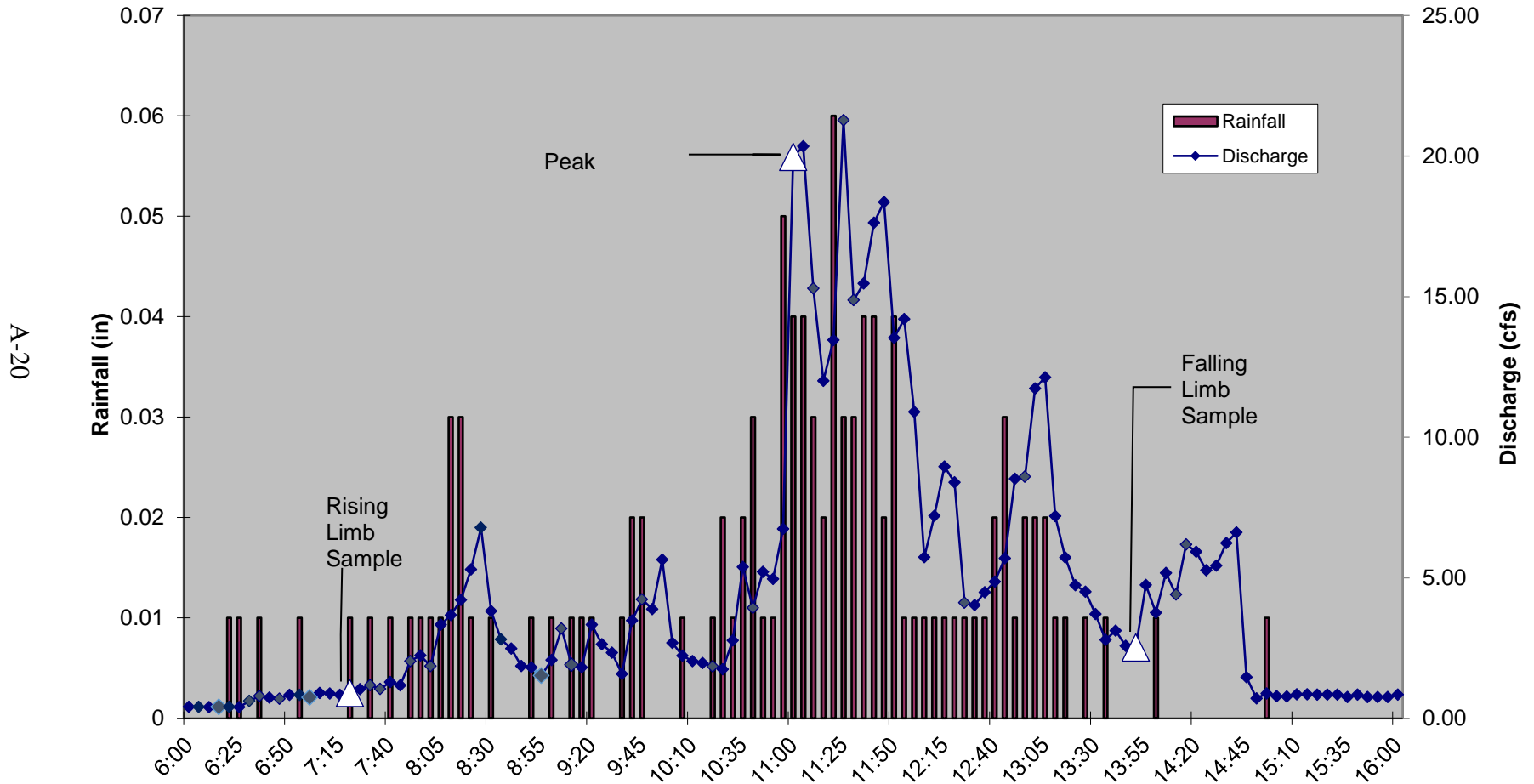
81-V



Hydrograph for February 28, 2017 Storm Parole Plaza

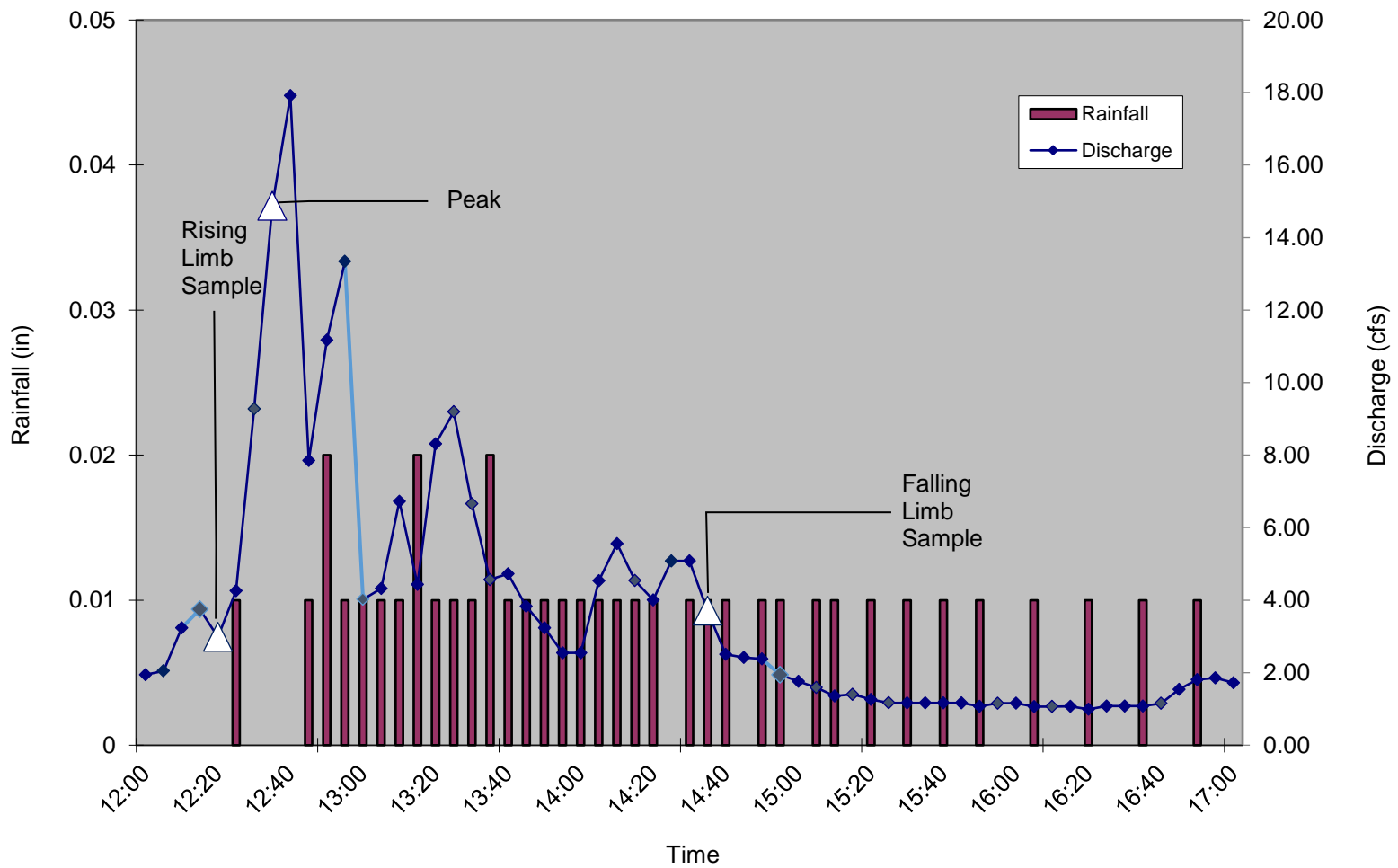


Hydrograph for March 31, 2017 Storm Parole Plaza



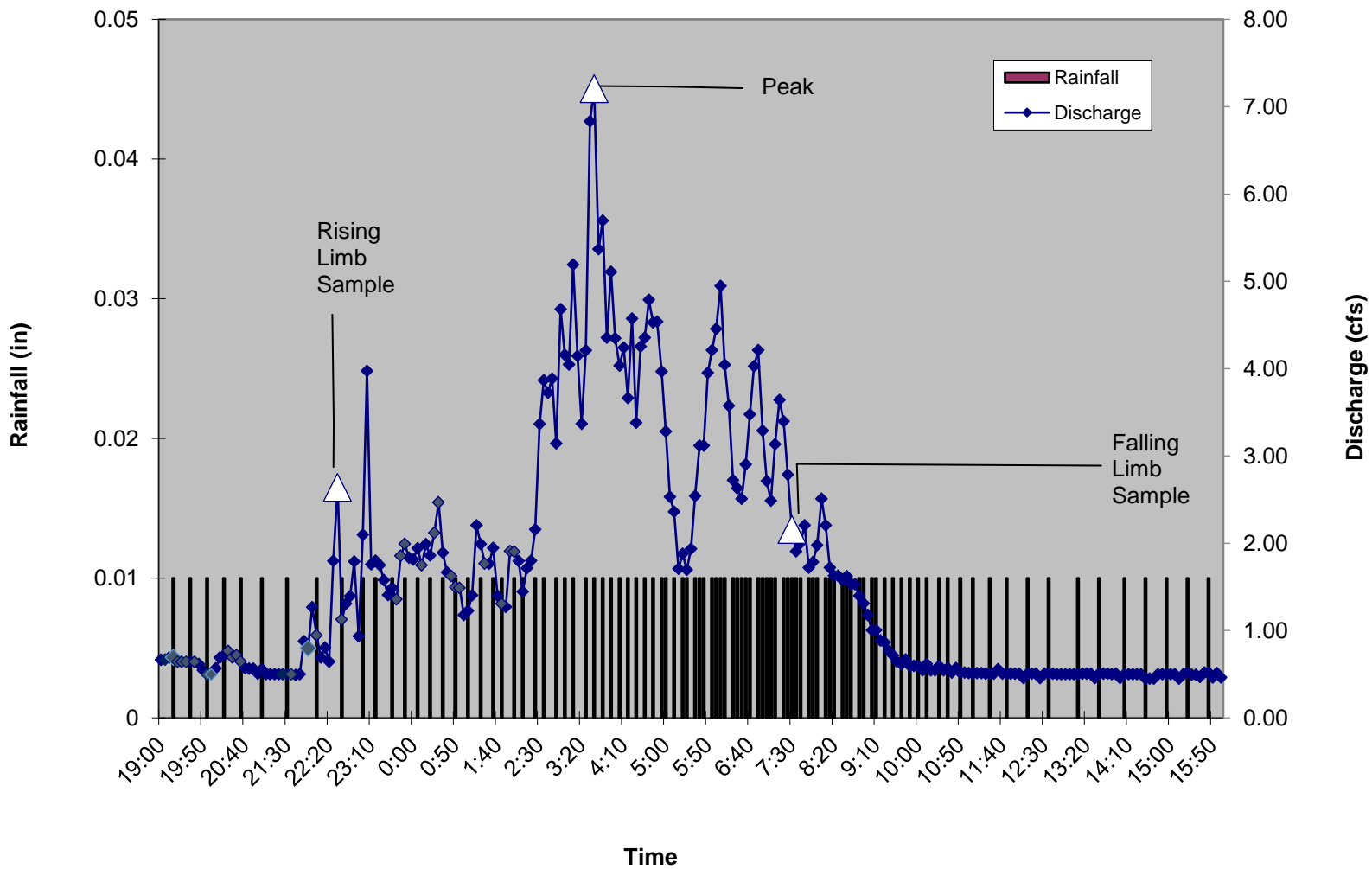
Hydrograph for April 25, 2017 Storm Parole Plaza

A-21



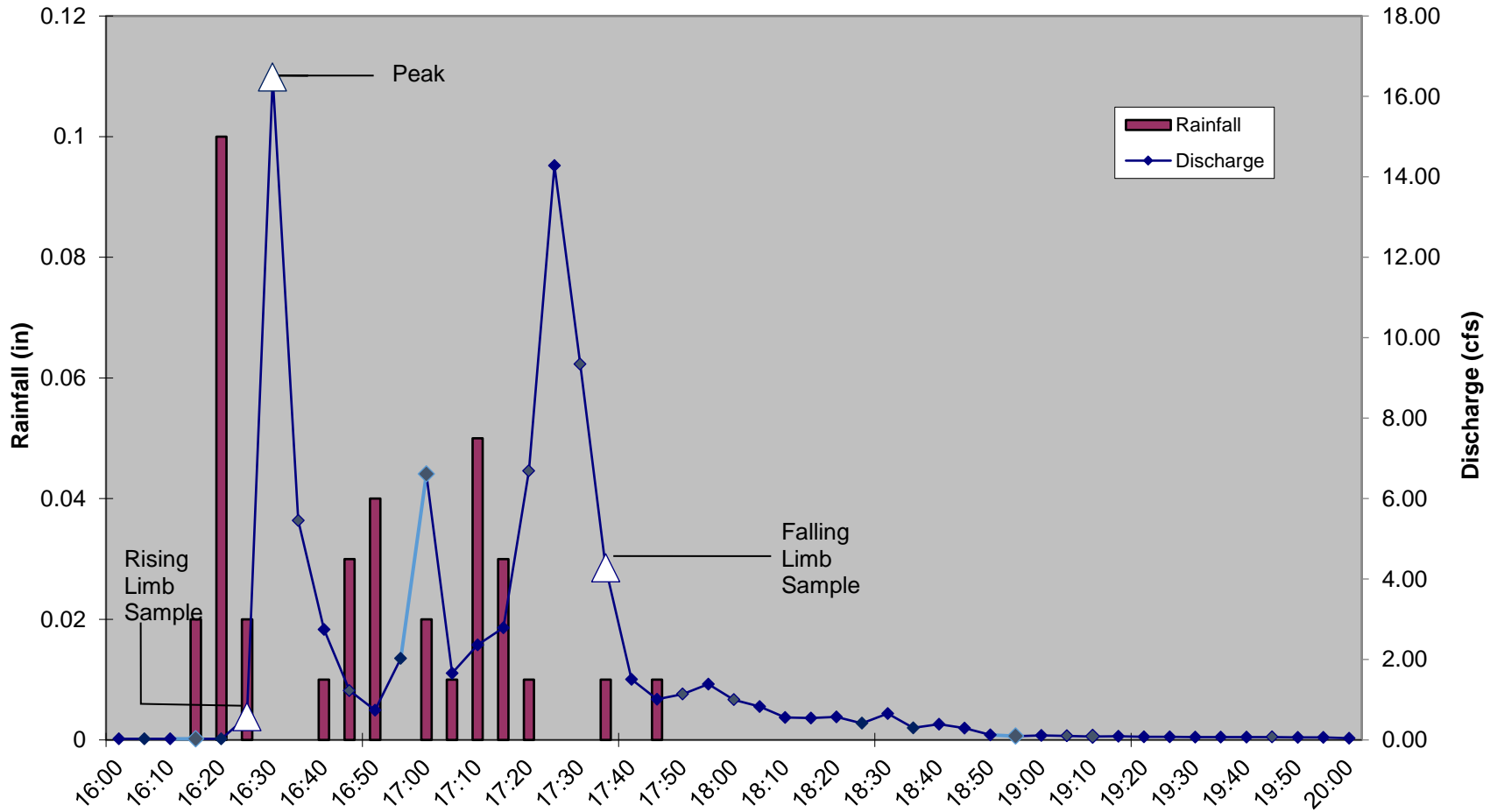
Hydrograph for May 12, 2017 Storm Parole Plaza

A-22



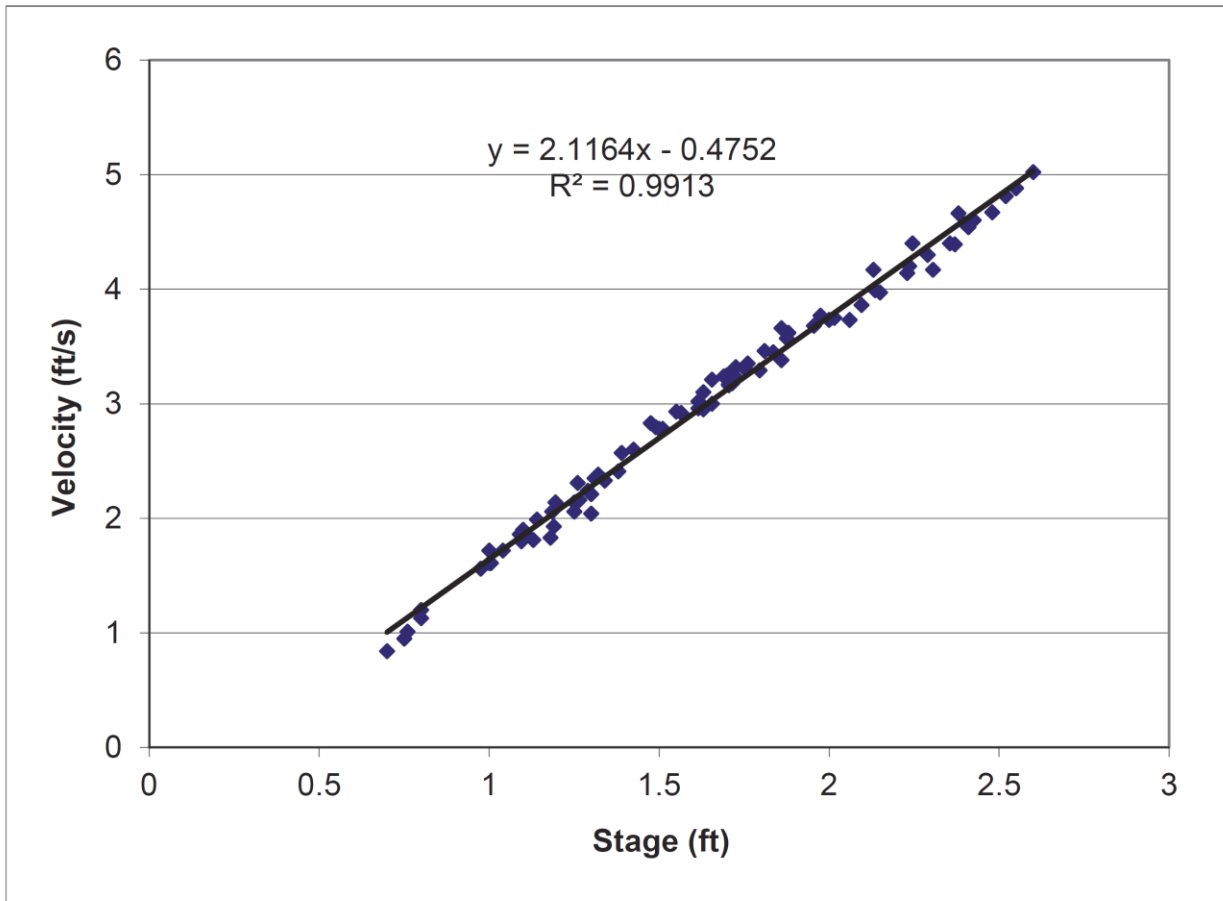
Hydrograph for June 19, 2017 Storm Parole Plaza

A-23

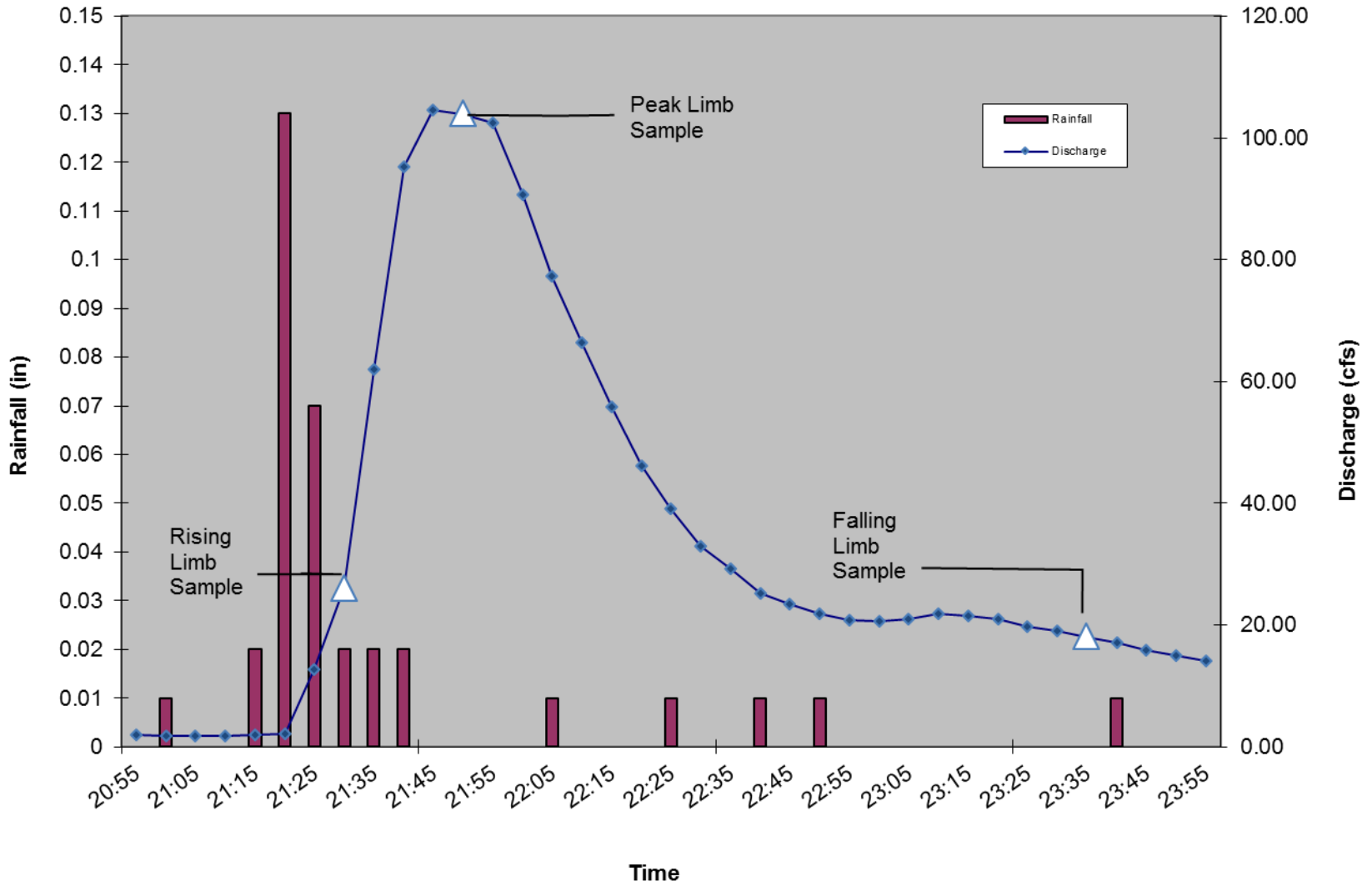


Church Creek Discharge Rating Table

| Stage (ft) | Flow Area (ft²) | Wetted Perimeter (ft) | Top Width (ft) | Velocity, (ft/s) | Discharge (cfs) |
|-------------------|-----------------------------------|------------------------------|-----------------------|-------------------------|------------------------|
| 0.0 | 0 | 0 | 0 | 0.00 | 0.00 |
| 0.1 | 0.21 | 2.77 | 2.75 | 0.00 | 0.00 |
| 0.2 | 0.53 | 3.62 | 3.57 | 0.00 | 0.00 |
| 0.3 | 0.91 | 4.03 | 3.93 | 0.16 | 0.15 |
| 0.4 | 1.31 | 4.36 | 4.19 | 0.37 | 0.49 |
| 0.5 | 1.75 | 4.83 | 4.58 | 0.58 | 1.02 |
| 0.6 | 2.22 | 5.09 | 4.73 | 0.79 | 1.76 |
| 0.7 | 2.70 | 5.34 | 4.88 | 1.01 | 2.72 |
| 0.8 | 3.20 | 5.73 | 5.20 | 1.22 | 3.90 |
| 0.9 | 3.74 | 6.14 | 5.54 | 1.43 | 5.35 |
| 1.0 | 4.31 | 6.48 | 5.81 | 1.64 | 7.07 |
| 1.1 | 4.90 | 6.75 | 5.98 | 1.85 | 9.08 |
| 1.2 | 5.50 | 7.01 | 6.16 | 2.06 | 11.35 |
| 1.3 | 6.13 | 7.28 | 6.33 | 2.28 | 13.95 |
| 1.4 | 6.77 | 7.53 | 6.49 | 2.49 | 16.84 |
| 1.5 | 7.43 | 7.80 | 6.66 | 2.70 | 20.06 |
| 1.6 | 8.10 | 8.08 | 6.86 | 2.91 | 23.58 |
| 1.7 | 8.80 | 8.37 | 7.06 | 3.12 | 27.48 |
| 1.8 | 9.51 | 8.65 | 7.26 | 3.33 | 31.71 |
| 1.9 | 10.25 | 8.93 | 7.44 | 3.55 | 36.35 |
| 2.0 | 11.00 | 9.15 | 7.52 | 3.76 | 41.33 |
| 2.1 | 11.75 | 9.35 | 7.54 | 3.97 | 46.64 |
| 2.2 | 12.51 | 9.55 | 7.57 | 4.18 | 52.30 |
| 2.3 | 13.26 | 9.75 | 7.60 | 4.39 | 58.24 |
| 2.4 | 14.03 | 9.96 | 7.63 | 4.60 | 64.60 |
| 2.5 | 14.79 | 10.16 | 7.65 | 4.82 | 71.23 |
| 2.6 | 15.56 | 10.36 | 7.68 | 5.03 | 78.23 |
| 2.7 | 16.33 | 10.56 | 7.71 | 5.24 | 85.55 |
| 2.8 | 17.10 | 10.76 | 7.73 | 5.45 | 93.21 |
| 2.9 | 17.87 | 10.96 | 7.76 | 5.66 | 101.19 |
| 3.0 | 18.65 | 11.17 | 7.79 | 5.87 | 109.55 |
| 3.1 | 19.43 | 11.37 | 7.81 | 6.09 | 118.24 |
| 3.2 | 20.21 | 11.57 | 7.84 | 6.30 | 127.27 |
| 3.3 | 21.00 | 11.77 | 7.87 | 6.51 | 136.69 |
| 3.4 | 21.79 | 11.97 | 7.89 | 6.72 | 146.44 |
| 3.5 | 22.58 | 12.18 | 7.92 | 6.93 | 156.53 |
| 3.6 | 23.37 | 12.38 | 7.95 | 7.14 | 166.95 |
| 3.7 | 24.17 | 12.58 | 7.98 | 7.36 | 177.78 |

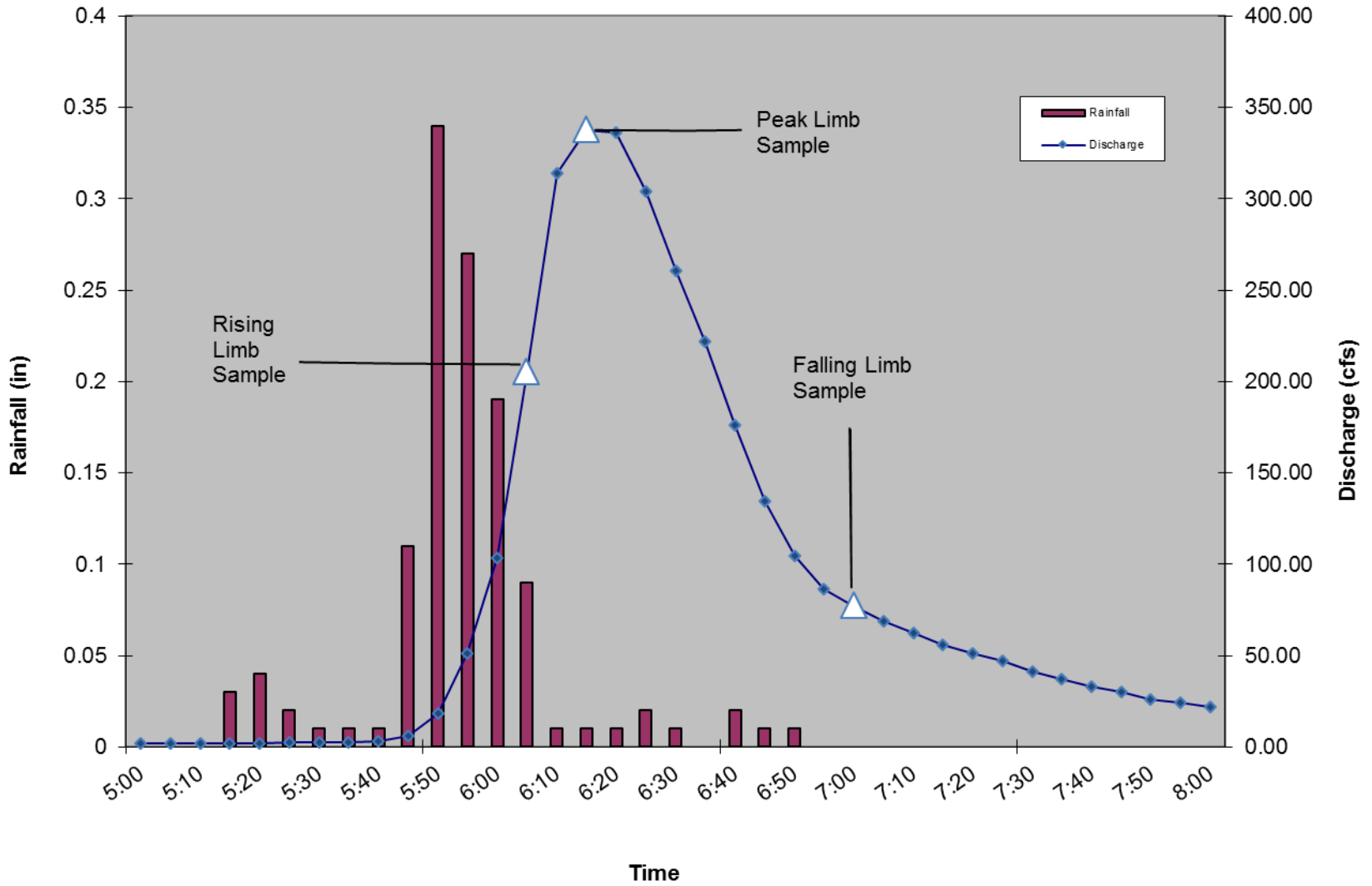


Hydrograph for July 28, 2016 Storm Church Creek



A-26

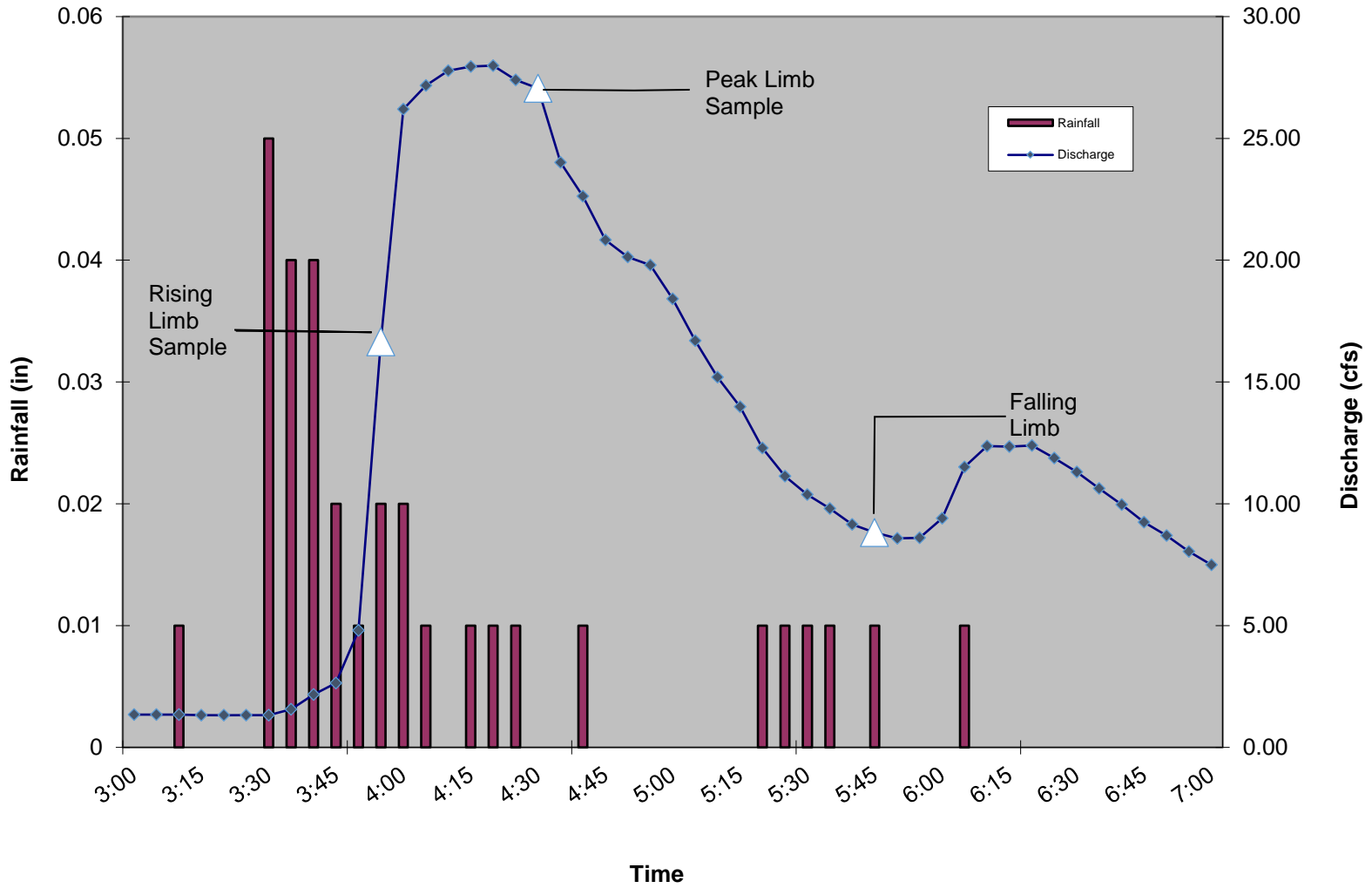
Hydrograph September 19, 2016 Storm Church Creek



A-27

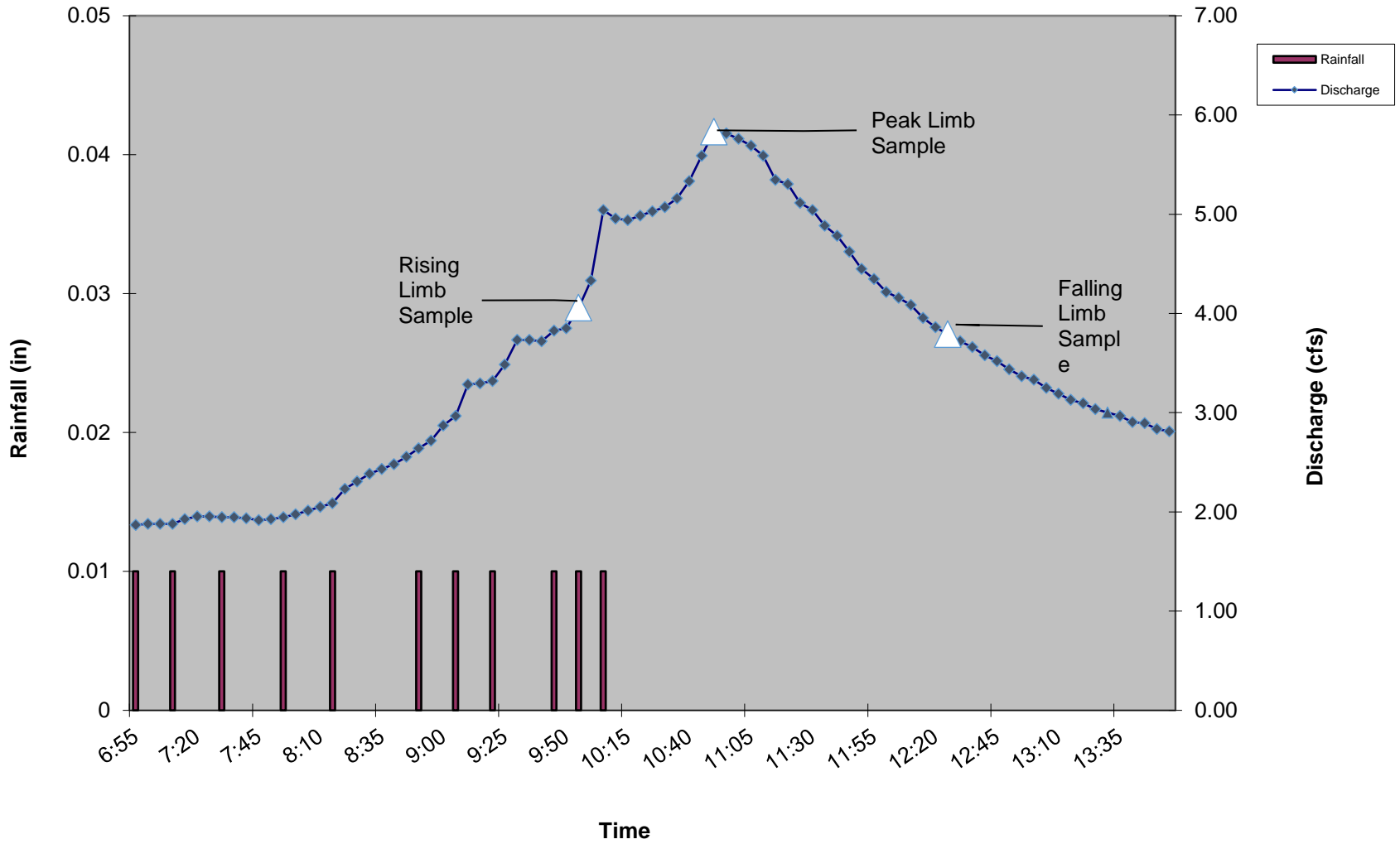
Hydrograph for September 27, 2016 Storm Church Creek

A-28

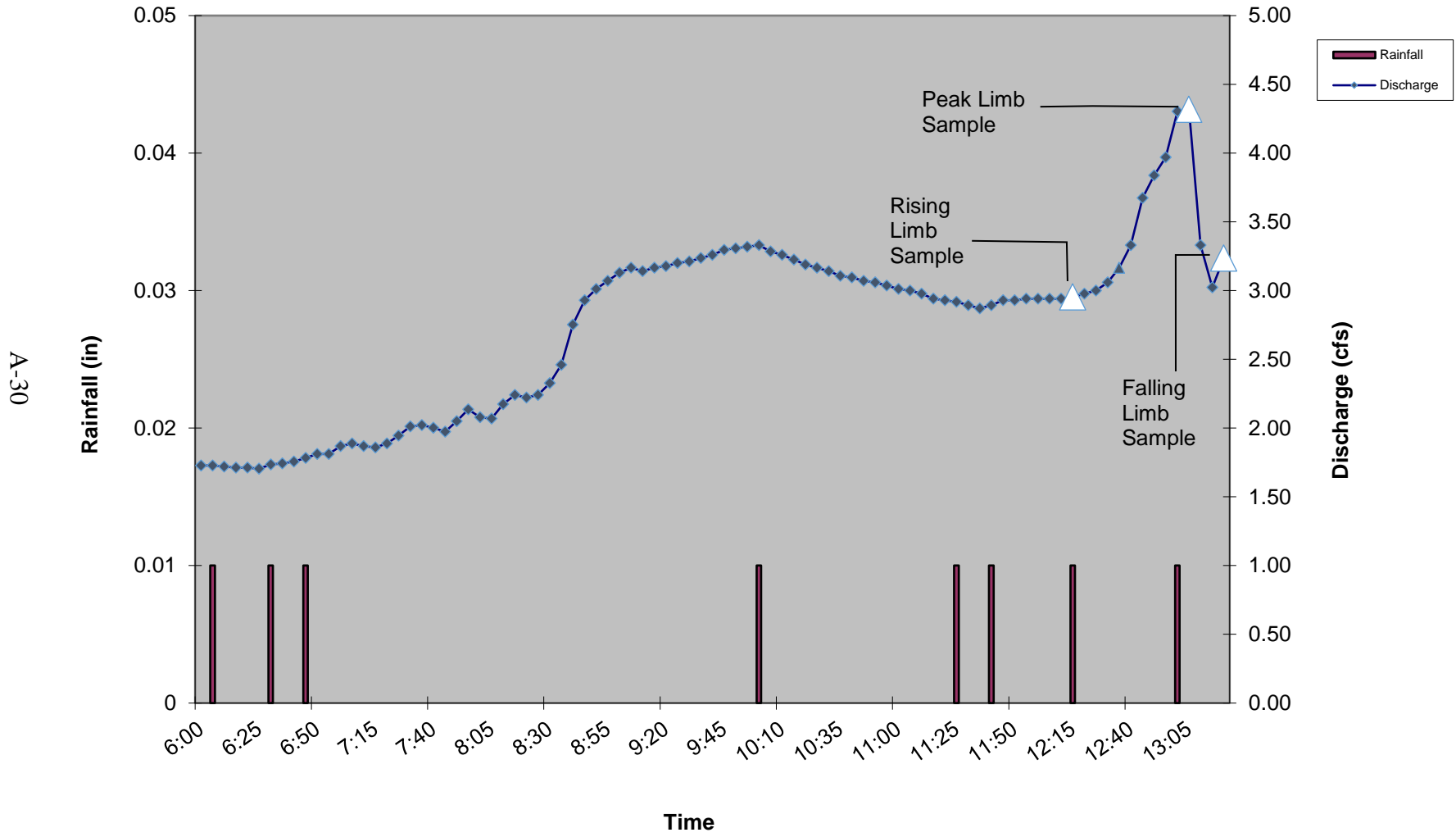


Hydrograph for November 9, 2016 Storm Church Creek

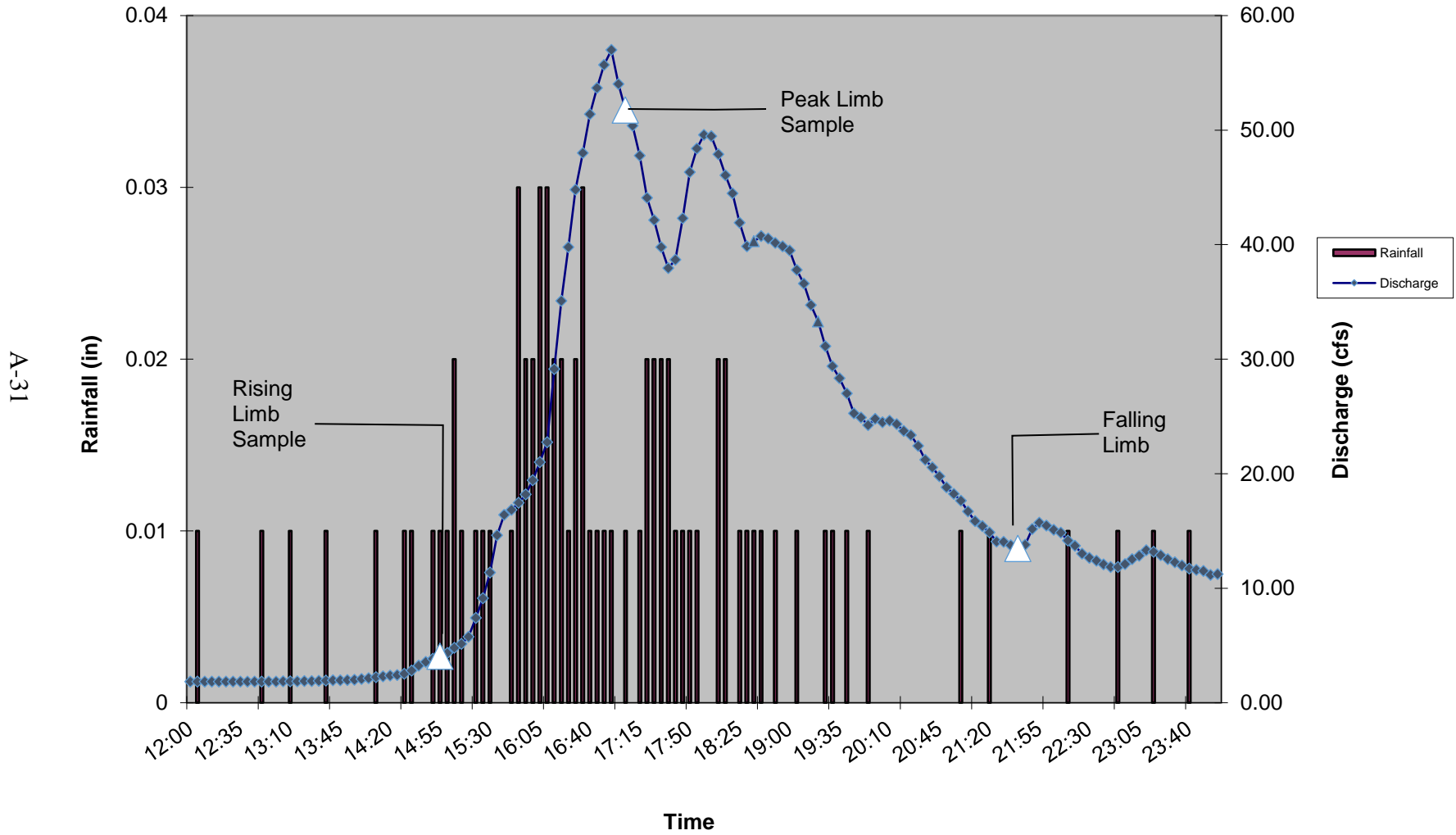
A-29



Hydrograph November 29, 2016 Storm Church Creek

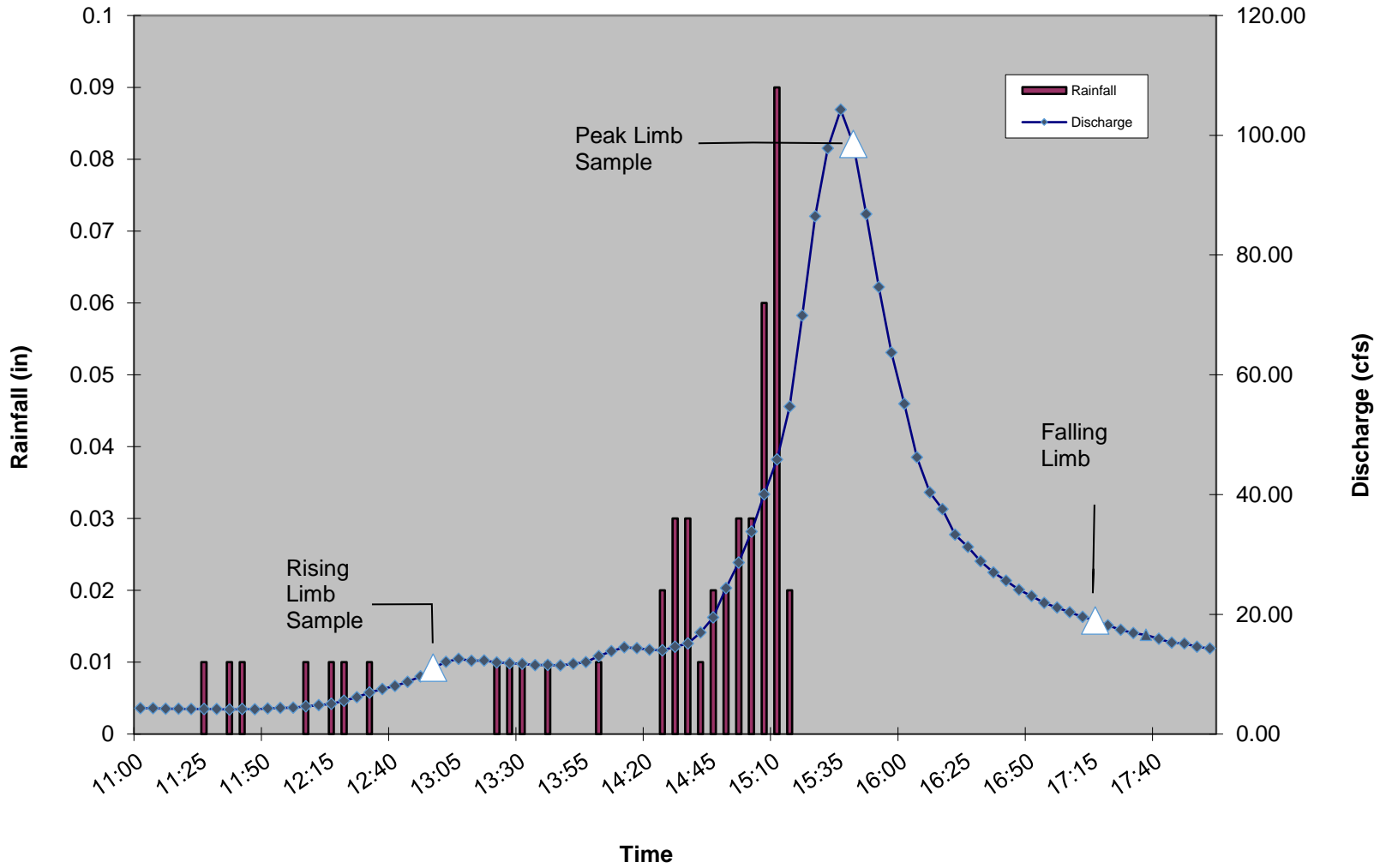


Hydrograph for December 6, 2016 Storm Church Creek

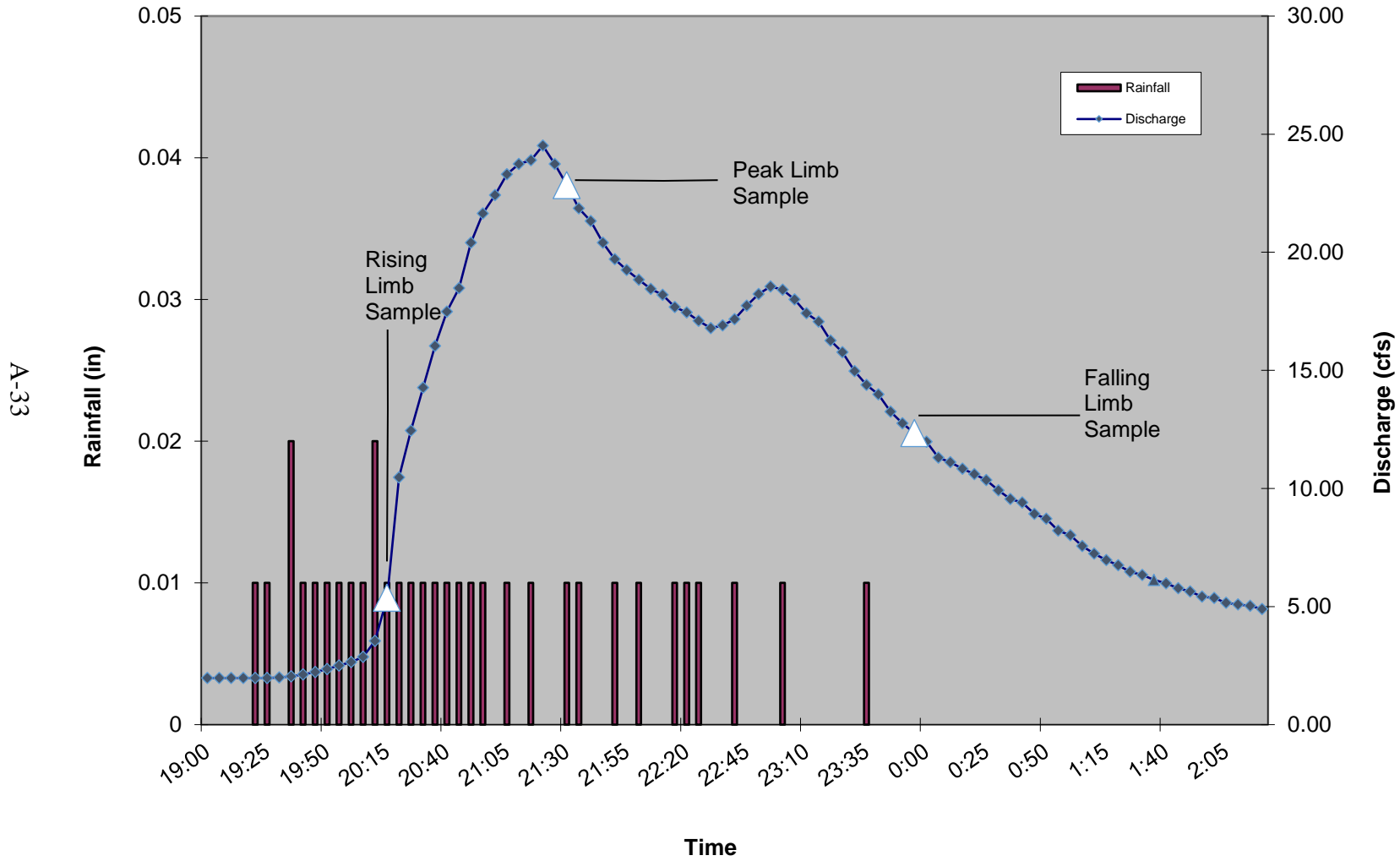


Hydrograph for January 23, 2017 Storm Church Creek

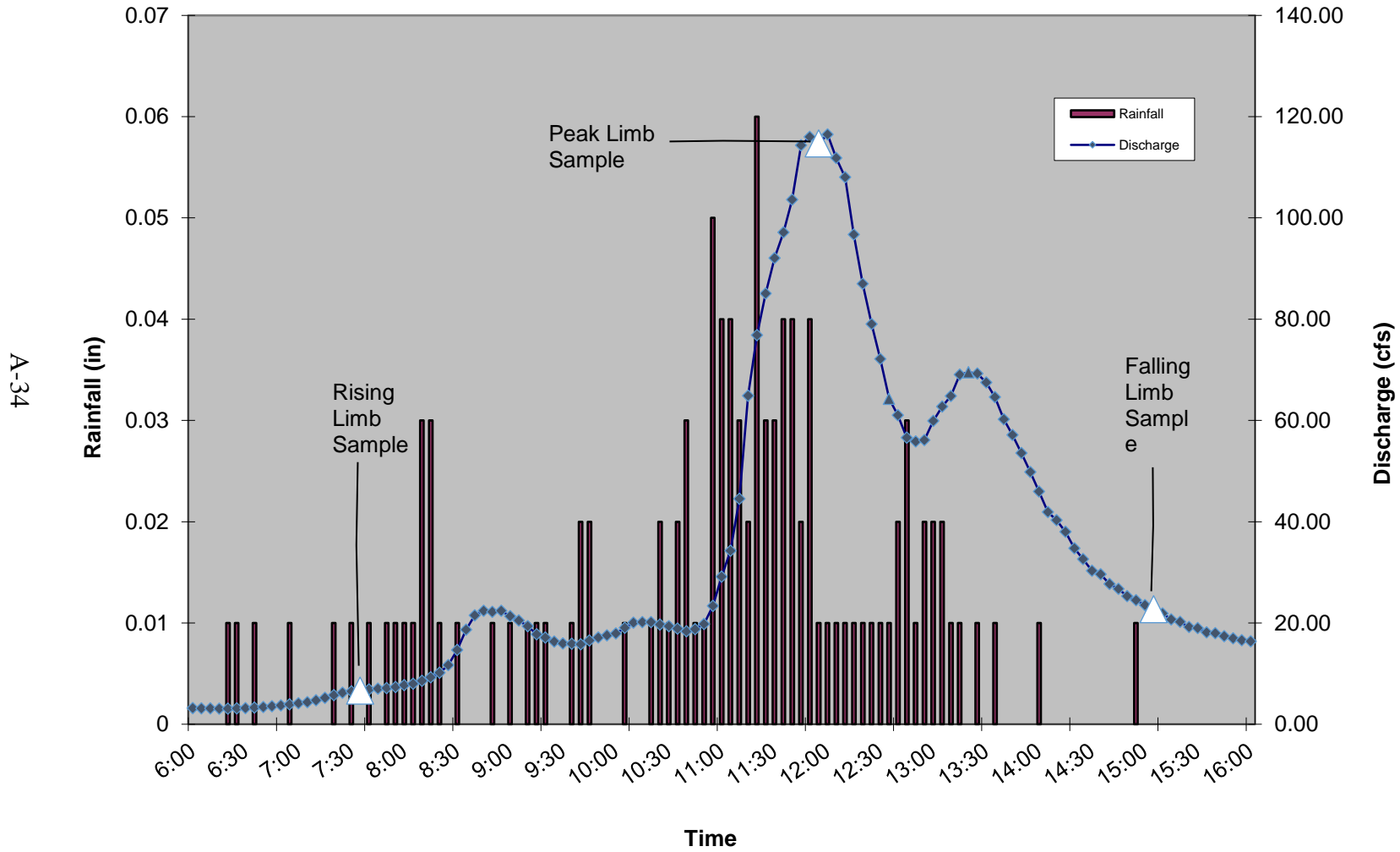
A-32



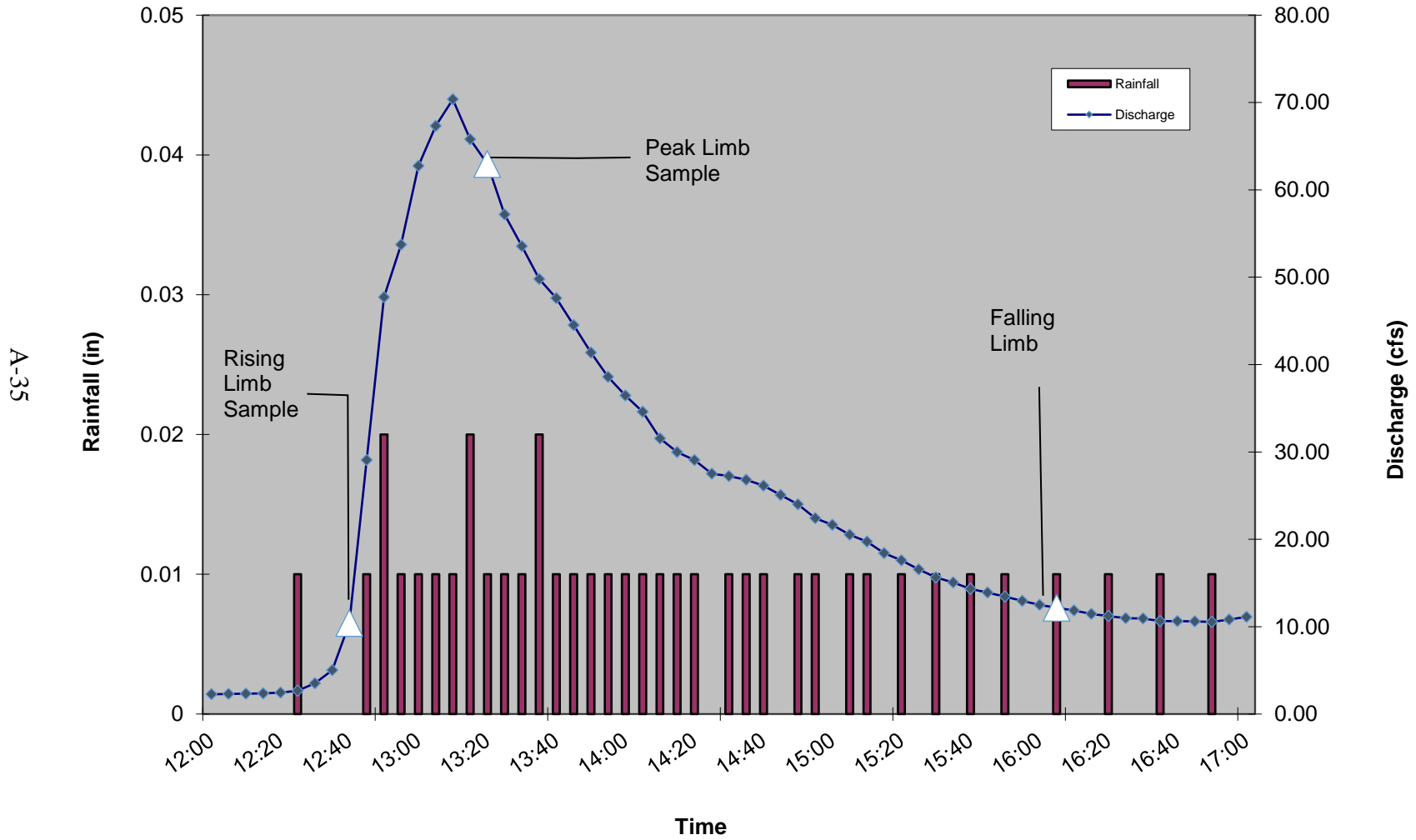
Hydrograph February 28, 2017 Storm Church Creek



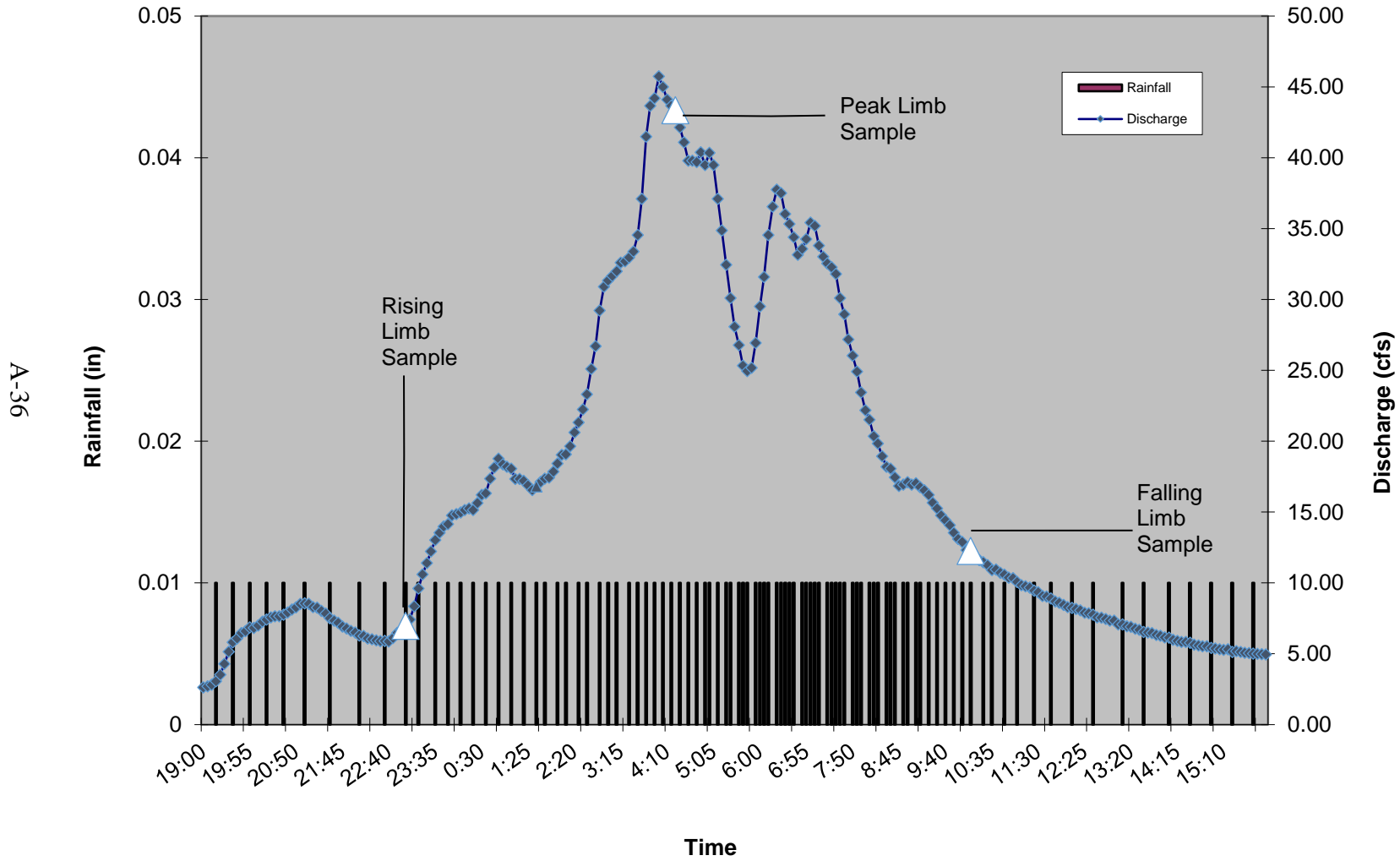
Hydrograph for March 31, 2017 Storm Church Creek



Hydrograph for April 25, 2017 Storm Church Creek

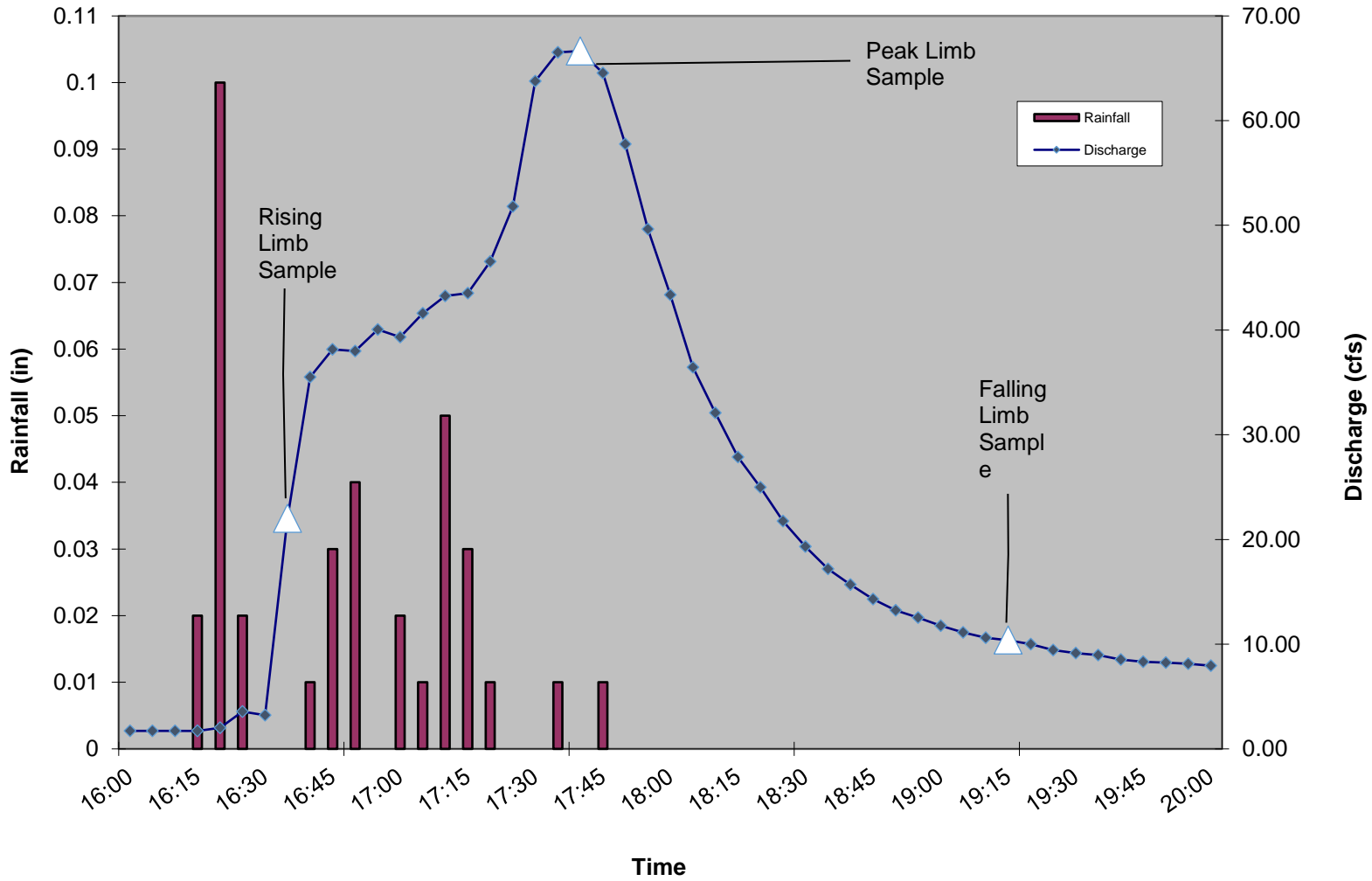


Hydrograph May 12, 2017 Storm Church Creek



Hydrograph for June 19, 2017 Storm Church Creek

A-37



APPENDIX B
MASTER TAXA LIST

| Order | Family | Genus | Taxon | FFG ^(a) | Habit ^(b) | Tolerance Value ^(c) |
|------------------|------------------|----------------|----------------|--------------------|----------------------|--------------------------------|
| Amphipoda | Crangonyctidae | Crangonyx | Crangonyx | Collector | sp | 6.7 |
| Amphipoda | Gammaridae | Gammarus | Gammarus | Shredder | sp | 6.7 |
| Arhynchobdellida | Erpobdellidae | | Erpobdellidae | Predator | sp | 10 |
| Basommatophora | Ancylidae | Ferrissia | Ferrissia | Scraper | cb | 7 |
| Basommatophora | Lymnaeidae | Fossaria | Fossaria | Scraper | cb | 6.9 |
| Basommatophora | Lymnaeidae | Pseudosuccinea | Pseudosuccinea | Collector | cb | 6.3 |
| Basommatophora | Physidae | Physa | Physa | Scraper | cb | 7 |
| Diptera | Chironomidae | Chironomini | Chironomini | | | 5.9 |
| Diptera | Chironomidae | Chironomus | Chironomus | Collector | bu | 4.6 |
| Diptera | Chironomidae | Conchapelopia | Conchapelopia | Predator | sp | 6.1 |
| Diptera | Chironomidae | Cricotopus | Cricotopus | Shredder | cn, bu | 9.6 |
| Diptera | Chironomidae | Dicrotendipes | Dicrotendipes | Collector | bu | 9 |
| Diptera | Chironomidae | Orthocladius | Orthocladius | Collector | sp, bu | 9.2 |
| Diptera | Chironomidae | Polypedilum | Polypedilum | Shredder | cb, cn | 6.3 |
| Diptera | Chironomidae | Potthastia | Potthastia | Collector | sp | 0 |
| Diptera | Chironomidae | Rheocricotopus | Rheocricotopus | Collector | sp | 6.2 |
| Diptera | Empididae | Hemerodromia | Hemerodromia | Predator | sp, bu | 7.9 |
| Diptera | Psychodidae | | Psychodidae | | | 4 |
| Diptera | Psychodidae | Psychoda | Psychoda | Collector | bu | 4 |
| Diptera | Sciomyzidae | | Sciomyzidae | Predator | bu | 6 |
| Diptera | Simuliidae | | Simuliidae | Filterer | cn | 3.2 |
| Diptera | Tipulidae | Tipula | Tipula | Shredder | bu | 6.7 |
| Haplotaxida | Enchytraeidae | | Enchytraeidae | Collector | bu | 9.1 |
| Haplotaxida | Naididae | | Naididae | Collector | bu | 8.5 |
| Hoploneurata | Tetrastemmatidae | Prostoma | Prostoma | Predator | | 7.3 |
| Isopoda | Asellidae | Caecidotea | Caecidotea | Collector | sp | 2.6 |
| Lumbricida | Lumbricidae | | Lumbricidae | Collector | | 10 |
| Lumbriculida | Lumbriculidae | | Lumbriculidae | Collector | bu | 6.6 |
| Odonata | Aeshnidae | Boyeria | Boyeria | Predator | cb, sp | 6.3 |
| Odonata | Calopterygidae | Calopteryx | Calopteryx | Predator | cb | 8.3 |
| Odonata | Coenagrionidae | Argia | Argia | Predator | cn, cb, sp | 9.3 |
| Odonata | Coenagrionidae | Ischnura | Ischnura | Predator | cb | 9 |
| Rhynchobdellida | Glossiphoniidae | Helobdella | Helobdella | Predator | sp | 6 |
| Trichoptera | Hydropsychidae | Cheumatopsyche | Cheumatopsyche | Filterer | cn | 6.5 |
| Tricladida | Dugesidae | Girardia | Girardia | Predator | sp | 9.3 |
| Tubificida | Tubificidae | | Tubificidae | Collector | cn | 8.4 |
| Tubificida | Tubificidae | Limnodrilus | Limnodrilus | Collector | cn | 8.6 |
| Veneroida | Pisidiidae | | Sphaeriidae | Filterer | bu | 6.5 |
| Veneroida | Pisidiidae | Pisidium | Pisidium | Filterer | bu | 5.7 |

^(a) Functional Feeding Group

^(b) Primary habit or form of locomotion, includes bu - burrower, cn - clinger, cb - climber, sk - skater, sp - sprawler, sw - swimmer
Some information for the particular taxa was not available.

^(c) Tolerance Values, based on Hilsenhoff, modified for Maryland

APPENDIX C
BIOLOGICAL ASSESSMENT RESULTS

Church Creek Site CC-01

Sampled: 4/12/2017

Biological Condition

Benthic Macroinvertebrate IBI

| | |
|------------------|------|
| Narrative Rating | Poor |
| BIBI Score | 2.14 |

| Metric | Value | Score |
|-----------------------|-------|-------|
| Total Taxa | 10 | 1 |
| EPT Taxa | 0 | 1 |
| Number Ephemeroptera | 0 | 1 |
| % Intolerant to Urban | 0 | 1 |
| % Ephemeroptera | 0 | 1 |
| Scraper Taxa | 2 | 5 |
| % Climbers | 64.35 | 5 |

Benthic Macroinvertebrate Taxa List

| Taxa | Count |
|---------------|-------|
| Argia | 9 |
| Boyeria | 3 |
| Enchytraeidae | 3 |
| Fossaria | 1 |
| Limnodrilus | 2 |
| Naididae | 13 |
| Physa | 61 |
| Prostoma | 5 |
| Psychodidae | 1 |
| Sciomyzidae | 4 |
| Tubificidae | 13 |

Physical Habitat

Maryland Biological Stream Survey PHI

| | |
|------------------|--------------------|
| Narrative Rating | Partially Degraded |
| PHI Score | 67.41 |

| Metric | Score |
|-----------------------|--------|
| Drainage area (acres) | 70.40 |
| Remoteness | 18.60 |
| Percent Shading | 68.32 |
| Epifaunal Substrate | 69.16 |
| Instream Habitat | 92.64 |
| Instream Wood Debris | 100.00 |
| Bank Stability | 55.74 |

Rapid Bioassessment Protocol

| | |
|------------------|----------------------|
| Narrative Rating | Partially Supporting |
| RBP Score | 74 |

| Metric | Score |
|-----------------------------|------------------|
| Epifaunal Substrate / Cover | 11 |
| Embeddedness | 11 |
| Velocity / Depth Regime | 10 |
| Sediment Deposition | 13 |
| Channel Flow Status | 8 |
| Channel Alteration | 19 |
| Frequency of Riffles | 16 |
| Bank Stability | 5(Left)/4(Right) |
| Vegetative Protection | 7(Left)/5(Right) |
| Riparian Veg Zone Width | 9(Left)/6(Right) |

Water Chemistry

| | |
|------------------------------|-------|
| Dissolved Oxygen (mg/L) | 13.86 |
| pH | 6.29 |
| Specific Conductance (µS/cm) | 2319 |
| Temperature (°C) | 17.6 |
| Turbidity (NTUs) | 17.5 |

Church Creek Site CC-02

Sampled: 4/12/2017

Biological Condition

Benthic Macroinvertebrate IBI

| | |
|------------------|------|
| Narrative Rating | Poor |
| BIBI Score | 2.14 |

| Metric | Value | Score |
|-----------------------|-------|-------|
| Total Taxa | 21 | 3 |
| EPT Taxa | 1 | 1 |
| Number Ephemeroptera | 0 | 1 |
| % Intolerant to Urban | 1.63 | 1 |
| % Ephemeroptera | 0 | 1 |
| Scraper Taxa | 1 | 3 |
| % Climbers | 11.38 | 5 |

Benthic Macroinvertebrate Taxa List

| Taxa | Count |
|----------------|-------|
| Argia | 3 |
| Caecidotea | 1 |
| Calopteryx | 2 |
| Cheumatopsyche | 1 |
| Chironomus | 1 |
| Conchapelopia | 1 |
| Crangonyx | 1 |
| Cricotopus | 13 |
| Cricotopus | 5 |
| Dicrotendipes | 2 |
| Gammarus | 53 |
| Girardia | 1 |
| Hemerodromia | 1 |
| Ischnura | 1 |
| Lumbriculidae | 5 |
| Naididae | 10 |
| Physa | 1 |
| Pisidium | 3 |
| Polypedilum | 5 |
| Polypedilum | 2 |
| Potthastia | 1 |
| Rheocricotopus | 1 |
| Sphaeriidae | 1 |
| Tubificidae | 8 |

Physical Habitat

Maryland Biological Stream Survey PHI

| | |
|------------------|----------|
| Narrative Rating | Degraded |
| PHI Score | 60.97 |

| Metric | Score |
|-----------------------|--------|
| Drainage area (acres) | 282.24 |
| Remoteness | 23.05 |
| Percent Shading | 68.32 |
| Epifaunal Substrate | 48.50 |
| Instream Habitat | 78.42 |
| Instream Wood Debris | 87.41 |
| Bank Stability | 60.10 |

Rapid Bioassessment Protocol

| | |
|------------------|----------------------|
| Narrative Rating | Partially Supporting |
| RBP Score | 61 |

| Metric | Score |
|-----------------------------|------------------|
| Epifaunal Substrate / Cover | 10 |
| Embeddedness | 8 |
| Velocity / Depth Regime | 9 |
| Sediment Deposition | 11 |
| Channel Flow Status | 11 |
| Channel Alteration | 13 |
| Frequency of Riffles | 10 |
| Bank Stability | 3(Left)/4(Right) |
| Vegetative Protection | 5(Left)/4(Right) |
| Riparian Veg Zone Width | 7(Left)/7(Right) |

Water Chemistry

| | |
|------------------------------|------|
| Dissolved Oxygen (mg/L) | 7.86 |
| pH | 7.04 |
| Specific Conductance (µS/cm) | 468 |
| Temperature (°C) | 15.4 |
| Turbidity (NTUs) | 16.2 |

Church Creek Site CC-03

Sampled: 4/17/2017

Biological Condition

Benthic Macroinvertebrate IBI

| | |
|------------------|------|
| Narrative Rating | Poor |
| BIBI Score | 2.43 |

| Metric | Value | Score |
|-----------------------|-------|-------|
| Total Taxa | 18 | 3 |
| EPT Taxa | 0 | 1 |
| Number Ephemeroptera | 0 | 1 |
| % Intolerant to Urban | 0.78 | 1 |
| % Ephemeroptera | 0 | 1 |
| Scraper Taxa | 2 | 5 |
| % Climbers | 14.84 | 5 |

Benthic Macroinvertebrate Taxa List

| Taxa | Count |
|---------------|-------|
| Argia | 5 |
| Caecidotea | 1 |
| Cricotopus | 32 |
| Cricotopus | 9 |
| Enchytraeidae | 1 |
| Erpobdellidae | 4 |
| Ferrissia | 2 |
| Gammarus | 9 |
| Helobdella | 1 |
| Hemerodromia | 1 |
| Ischnura | 1 |
| Limnodrilus | 4 |
| Naididae | 29 |
| Orthocladius | 4 |
| Orthocladius | 2 |
| Physa | 8 |
| Polypedilum | 3 |
| Prostoma | 1 |
| Psychoda | 1 |
| Simuliidae | 1 |
| Tubificidae | 9 |

Physical Habitat

Maryland Biological Stream Survey PHI

| | |
|------------------|--------------------|
| Narrative Rating | Partially Degraded |
| PHI Score | 71.72 |

| Metric | Score |
|-----------------------|--------|
| Drainage area (acres) | 282.24 |
| Remoteness | 20.96 |
| Percent Shading | 45.47 |
| Epifaunal Substrate | 100.00 |
| Instream Habitat | 100.00 |
| Instream Wood Debris | 93.33 |
| Bank Stability | 70.56 |

Rapid Bioassessment Protocol

| | |
|------------------|----------------------|
| Narrative Rating | Partially Supporting |
| RBP Score | 70 |

| Metric | Score |
|-----------------------------|------------------|
| Epifaunal Substrate / Cover | 15 |
| Embeddedness | 11 |
| Velocity / Depth Regime | 12 |
| Sediment Deposition | 16 |
| Channel Flow Status | 15 |
| Channel Alteration | 12 |
| Frequency of Riffles | 6 |
| Bank Stability | 6(Left)/8(Right) |
| Vegetative Protection | 5(Left)/4(Right) |
| Riparian Veg Zone Width | 4(Left)/4(Right) |

Water Chemistry

| | |
|------------------------------|------|
| Dissolved Oxygen (mg/L) | 6.08 |
| pH | 6.67 |
| Specific Conductance (µS/cm) | 479 |
| Temperature (°C) | 17 |
| Turbidity (NTUs) | 15.7 |

Church Creek Site CC-04

Sampled: 4/12/2017

Biological Condition

Benthic Macroinvertebrate IBI

| | |
|------------------|-----------|
| Narrative Rating | Very Poor |
| BIBI Score | 1.86 |

| Metric | Value | Score |
|-----------------------|-------|-------|
| Total Taxa | 16 | 3 |
| EPT Taxa | 0 | 1 |
| Number Ephemeroptera | 0 | 1 |
| % Intolerant to Urban | 5.04 | 1 |
| % Ephemeroptera | 0 | 1 |
| Scraper Taxa | 1 | 3 |
| % Climbers | 6.72 | 3 |

Benthic Macroinvertebrate Taxa List

| Taxa | Count |
|----------------|-------|
| Caecidotea | 6 |
| Chironomini | 1 |
| Chironomus | 3 |
| Conchapelopia | 4 |
| Conchapelopia | 1 |
| Cricotopus | 4 |
| Cricotopus | 22 |
| Erpobdellidae | 2 |
| Gammarus | 3 |
| Limnodrilus | 5 |
| Lumbricidae | 1 |
| Lumbriculidae | 4 |
| Naididae | 30 |
| Orthocladius | 1 |
| Physa | 2 |
| Pisidium | 1 |
| Polypedilum | 4 |
| Pseudosuccinea | 2 |
| Tipula | 1 |
| Tubificidae | 22 |

Physical Habitat

Maryland Biological Stream Survey PHI

| | |
|------------------|--------------------|
| Narrative Rating | Partially Degraded |
| PHI Score | 67.92 |

| Metric | Score |
|-----------------------|--------|
| Drainage area (acres) | 110.53 |
| Remoteness | 20.96 |
| Percent Shading | 40.96 |
| Epifaunal Substrate | 72.03 |
| Instream Habitat | 99.12 |
| Instream Wood Debris | 100.00 |
| Bank Stability | 74.42 |

Rapid Bioassessment Protocol

| | |
|------------------|------------|
| Narrative Rating | Supporting |
| RBP Score | 78 |

| Metric | Score |
|-----------------------------|------------------|
| Epifaunal Substrate / Cover | 13 |
| Embeddedness | 11 |
| Velocity / Depth Regime | 8 |
| Sediment Deposition | 8 |
| Channel Flow Status | 14 |
| Channel Alteration | 19 |
| Frequency of Riffles | 16 |
| Bank Stability | 8(Left)/7(Right) |
| Vegetative Protection | 5(Left)/5(Right) |
| Riparian Veg Zone Width | 8(Left)/9(Right) |

Water Chemistry

| | |
|------------------------------|------|
| Dissolved Oxygen (mg/L) | 5.58 |
| pH | 6.49 |
| Specific Conductance (µS/cm) | 392 |
| Temperature (°C) | 19.3 |
| Turbidity (NTUs) | 13.2 |

| Select physical habitat parameters (raw scores) 2017 | | | |
|--|------------------------------|-------------------------|-------------------------|
| Site | Epifaunal Substrate (0 – 20) | Instream Habitat (0-20) | Embeddedness (0 – 100%) |
| CC-01 | 7 | 10 | 80 |
| CC-02 | 5 | 10 | 80 |
| CC-03 | 15 | 16 | 35 |
| CC-04 | 8 | 12 | 90 |

APPENDIX D
QA/QC INFORMATION

Quality Assurance/Quality Control Summary for NPDES Monitoring Activities

This section describes all Quality Assurance/Quality Control (QA/QC) procedures implemented for this project including field sampling, laboratory sorting and subsampling, data entry, metric calculation, final IBI calculation, geomorphic field sampling, and classification of stream types.

Field Sampling

Initial QA/QC procedures for benthic macroinvertebrate field sampling included formal training for field crew leaders in MBSS Sampling Protocols. All field crew members have attended at least one MBSS Spring Index Period Training. At least one crew member extensively trained and certified in MBSS sampling protocols was present for each field sampling day. Also during field sampling, each data sheet was double checked for completeness and sample bottle labels were double checked for accuracy. Geomorphic assessment field crews have more than one year of experience conducting similar assessment using the Rosgen Stream Classification Methodology, and the Geomorphic Field Leader has completed two or more levels of Rosgen training.

Geomorphic assessment survey equipment is calibrated annually and regularly inspected to ensure proper functioning. Cross-section and profile data were digitally plotted and analyzed in Ohio Department of Natural Resources (ODNR) Reference Reach Spreadsheet Version 4.3L for accuracy.

Water quality QA/QC procedures included calibration of the YSI multiprobe meter daily during the sampling season. Dissolved oxygen probe membranes were inspected regularly and replaced when dirty or damaged.

Laboratory Sorting and Subsampling

Sorting QA/QC was conducted on one sample since only seven samples were collected for this survey (The four samples from Church Creek are analyzed concurrently with three samples taken in Picture Spring Branch). This check consisted of entirely resorting the sorted grid cells of one randomly selected sample. This QC met the sorting efficiency criterion of 90%, so no further action was required. As a taxonomic QC, one sample was re-identified completely by another Versar SFS-certified taxonomist following the same identification methods stated above. The Percent Difference in Enumeration (PDE) and the Percent Taxonomic Disagreement (PTD) were calculated, and no further action was required since both the PDE and PTD met MBSS and County MQO requirements.

Data Entry

All data entered were double checked by someone other than the person who performed the initial data entry. Any errors found during QA/QC were corrected to ensure 100% accuracy of the data.

Metric and IBI Calculations

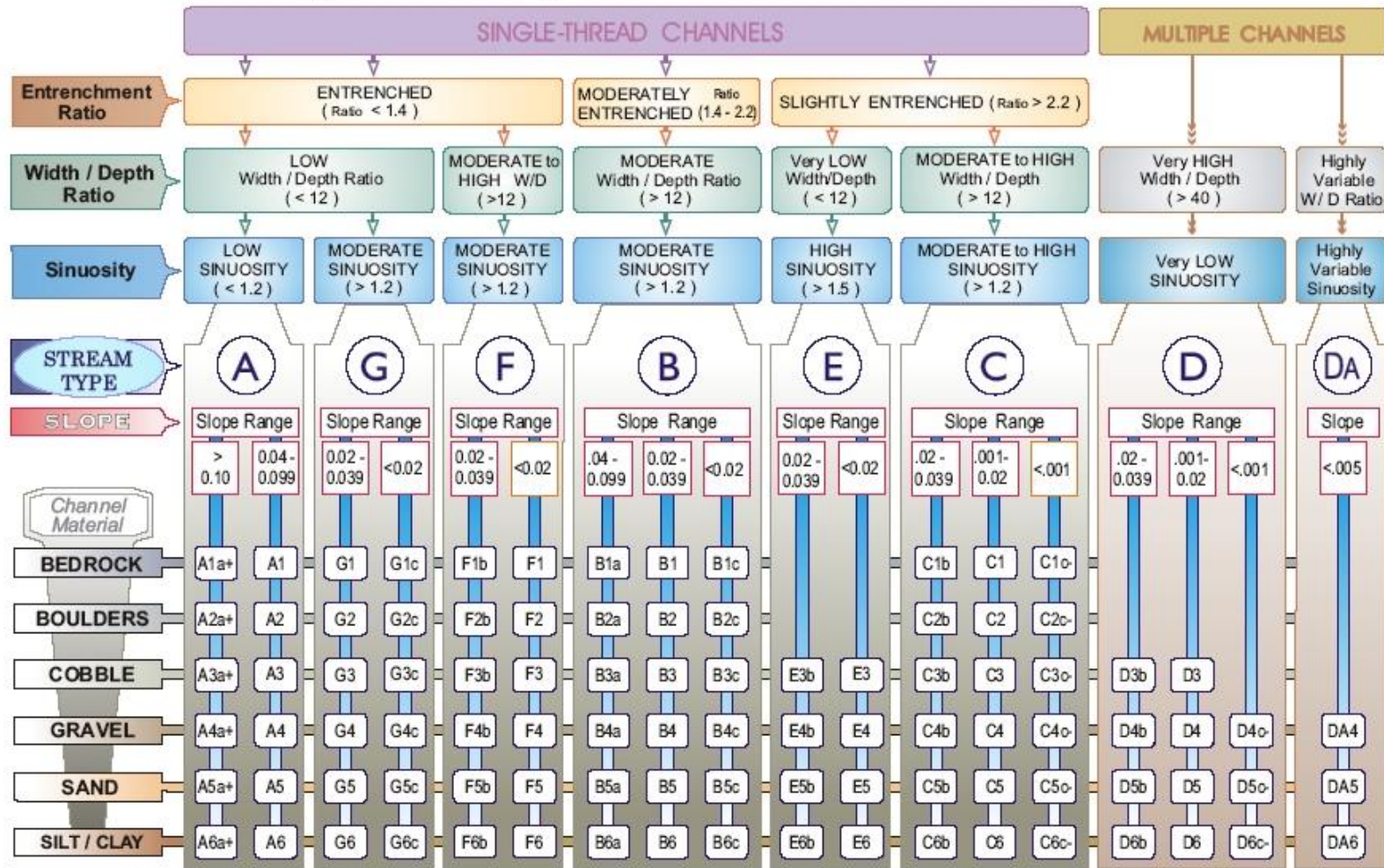
Ten percent of metric and IBI calculations were checked by hand using a calculator to ensure correct calculation by the Access database. Any discrepancies were addressed at that time.

Identification of Stream Types

All stream types were determined by hand based on the methods of the Rosgen Stream Classification (Rosgen 1996). Due to the natural variability, or continuum, of streams, adjustments in the values of Width Depth Ratio (+/- 2.0) and Entrenchment Ratio (+/-0.2) are allowed, which may result in assigning a different stream type. Therefore, all stream types assigned were checked and any necessary adjustments were made.

APPENDIX E
ROSGEN CLASSIFICATION SCHEME

The Key to the Rosgen Classification of Natural Rivers



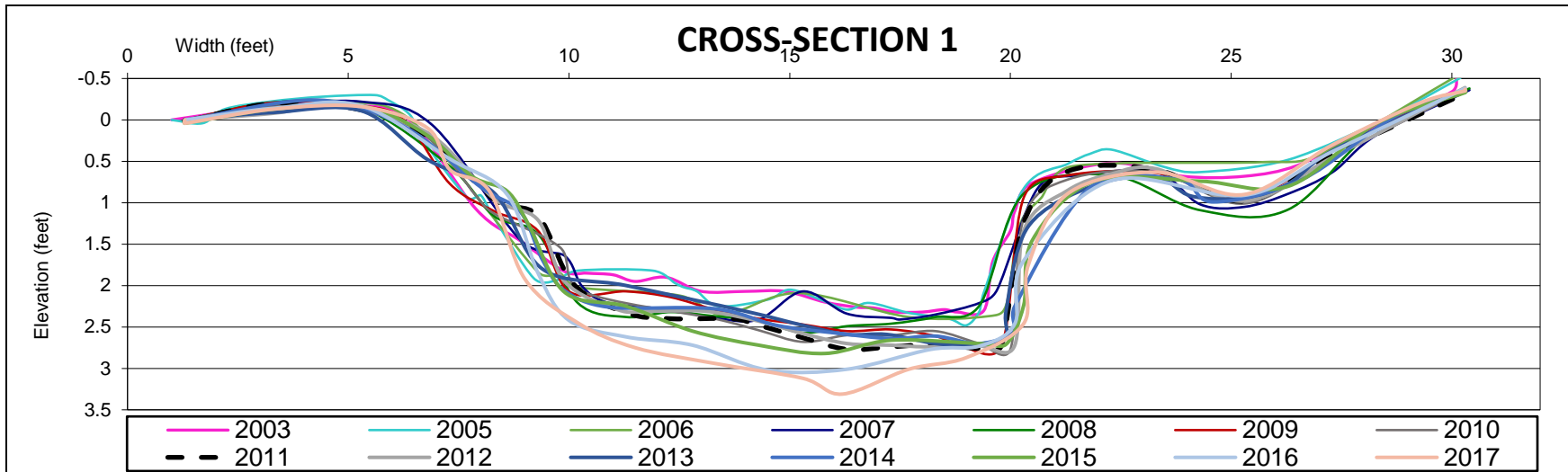
KEY to the **ROSGEN** CLASSIFICATION of NATURAL RIVERS. As a function of the "continuum of physical variables" within stream reaches, values of **Entrenchment** and **Sinuosity** ratios can vary by +/- 0.2 units; while values for **Width / Depth** ratios can vary by +/- 2.0 units.

Source: Rosgen, D.L. 1996. *Applied River Morphology*. Wildland Hydrology, Pagosa Springs, CO.

APPENDIX F
GEOMORPHOLOGICAL DATA

Church Creek 2017 Geomorphic Assessment Results Summary

| Assessment Parameter | Cross-section | | | | |
|---------------------------------------|----------------------------|--------------------------|--------------------------|--------------------------|----------------------------|
| | XS-1 Glide @ sta 3+70.5 | XS-2 Glide @ sta 6+82 | XS-3 Pool @ sta 11+00 | XS-4 Pool @ sta 13+53 | XS-5 Riffle @ sta 17+10 |
| Classification | F4 | G4c | G4c | E4/5 | F4 |
| Bankfull Width (ft) | 13.8 | 9.0 | 9.9 | 8.8 | 11.2 |
| Mean Depth (ft) | 1.7 | 1.2 | 2.1 | 1.6 | 0.8 |
| Bankfull X-Sec Area (sq ft) | 23.4 | 11.0 | 21.2 | 14.4 | 9.4 |
| Width:Depth Ratio | 8.1 | 7.4 | 4.7 | 5.4 | 13.4 |
| Flood-Prone Width (ft) | 29.0 | 22.8 | 20.9 | 23.5 | 19.6 |
| Entrenchment Ratio | 2.1 | 2.5 | 2.1 | 2.7 | 1.7 |
| D50(mm) | 8 | 9.5 | 1.3 | 5.5 | 15 |
| Water Surface Slope (ft/ft) | 0.00097 | 0.028 | 0.01 | 0.0028 | 0.0079 |
| Sinuosity | <1.2 | <1.2 | <1.2 | <1.2 | <1.2 |
| Drainage Area (mi²) | 0.111 | 0.113 | 0.121 | 0.130 | 0.441 |
| Adjustments? | Sin ↑, ER ↓, W/D ↑ | Sin ↑, ER ↓ | Sin ↑, ER ↓ | Sin ↑ | Sin ↑, ER ↓ |



F-4



Upstream View

Downstream View

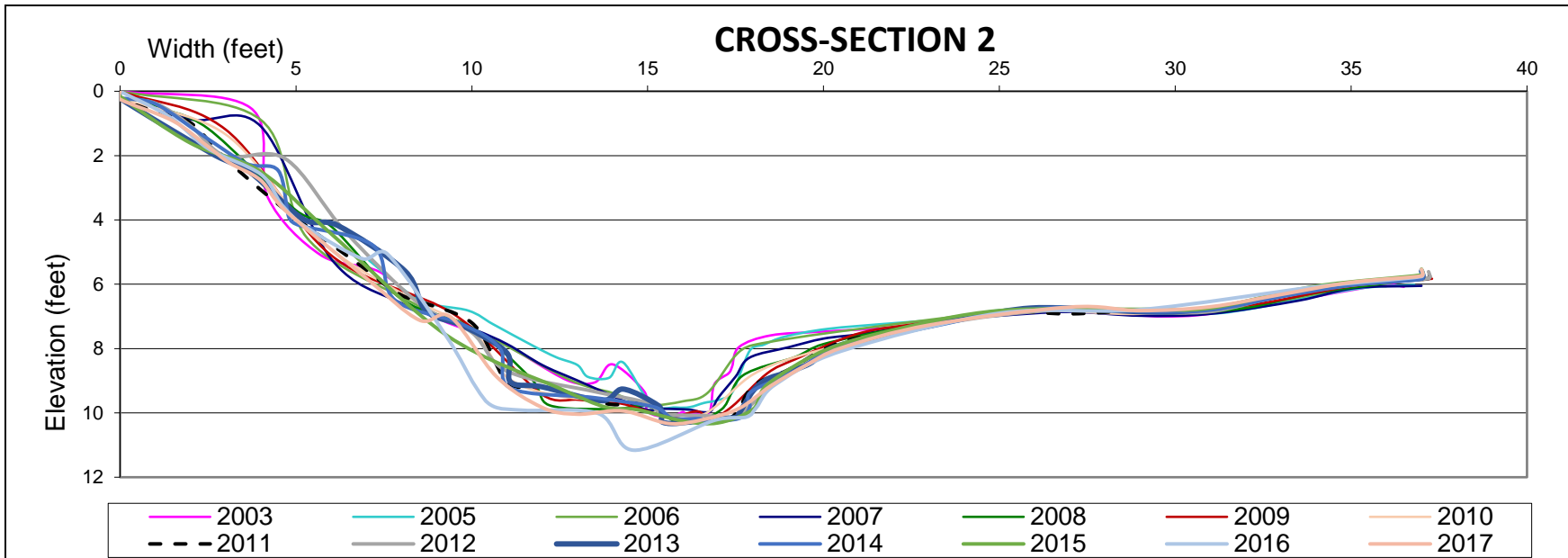


Left Bank View

Right Bank View

2017 Geomorphic Assessment Results

| | |
|--|--------------------|
| Bankfull Width (W_{bkt}) (feet) | 13.8 |
| Mean Depth (d_{bkt}) (feet) | 1.7 |
| Bankfull Cross-sectional Area (A_{bkt}) (feet ²) | 23.4 |
| Width/Depth Ratio (W_{bkt}/d_{bkt}) | 8.1 |
| Width of Flood-prone Area (W_{fpa}) (feet) | 29.0 |
| Entrenchment Ratio (ER) = W_{fpa}/W_{bkt} | 2.1 |
| Channel Materials D_{50} (millimeters) | 8 |
| Water Surface Slope (S) | 0.00097 |
| Sinuosity (K) = stream length/valley length | <1.2 |
| Adjustments? | Sin ↑, ER ↓, W/D ↑ |
| STREAM TYPE | F4 |



Upstream View

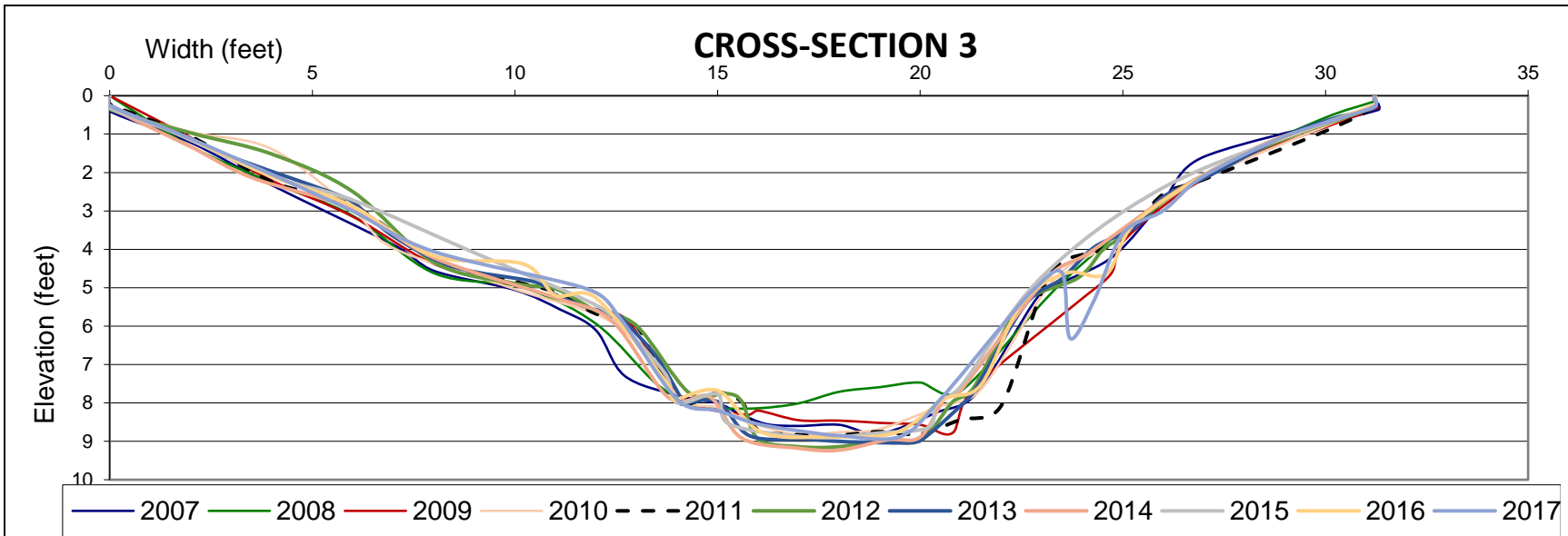
Downstream View





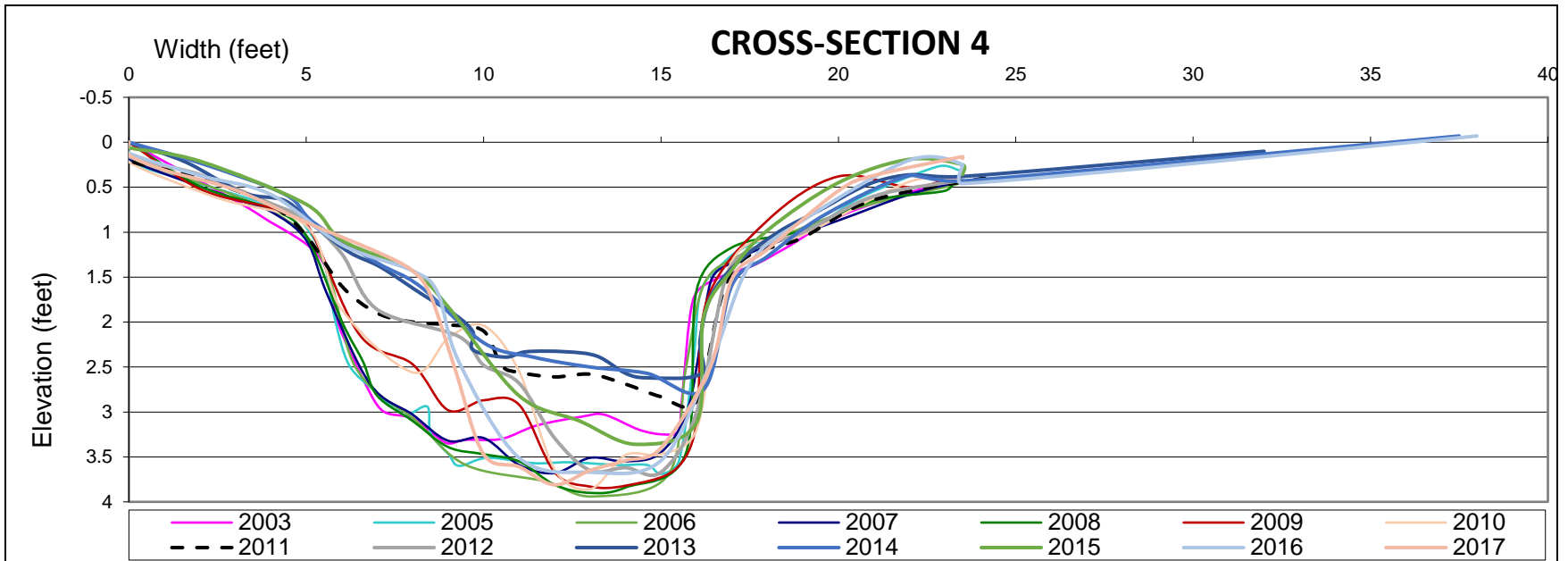
Left Bank View

Right Bank View

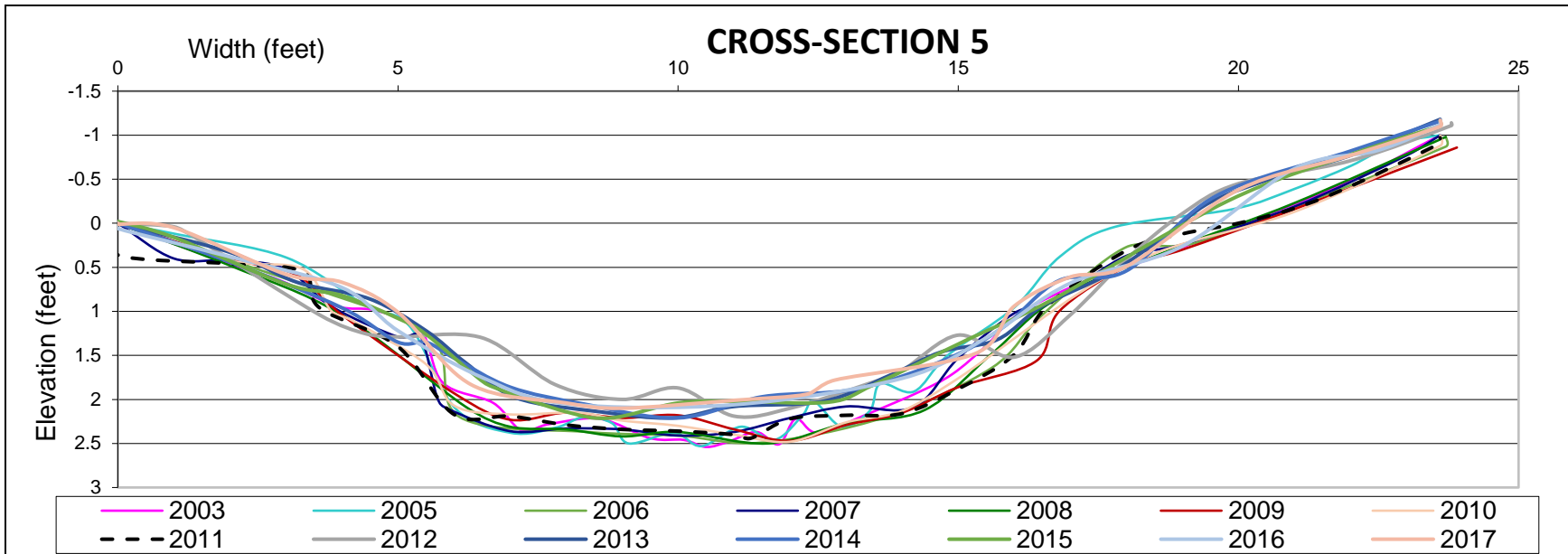
| 2017 Geomorphic Assessment Results | |
|--|-------------|
| Bankfull Width (W_{bkt}) (feet) | 9.0 |
| Mean Depth (d_{bkt}) (feet) | 1.2 |
| Bankfull Cross-sectional Area (A_{bkt}) (feet ²) | 11.0 |
| Width/Depth Ratio (W_{bkt}/d_{bkt}) | 7.4 |
| Width of Flood-prone Area (W_{fpa}) (feet) | 22.8 |
| Entrenchment Ratio (ER) = W_{fpa}/W_{bkt} | 2.5 |
| Channel Materials D_{50} (millimeters) | 9.5 |
| Water Surface Slope (S) | 0.028 |
| Sinuosity (K) = stream length/valley length | <1.2 |
| Adjustments? | Sin ↑, ER ↓ |
| STREAM TYPE | G4c |







| | | | | | |
|--|--|---|--|--|-------------|
|  | |  | | 2017 Geomorphic Assessment Results | |
| Upstream View | | Downstream View | | Bankfull Width (W_{bkt}) (feet) | 9.9 |
| | | | | Mean Depth (d_{bkt}) (feet) | 2.1 |
| Left Bank View | | Right Bank View | | Bankfull Cross-sectional Area (A_{bkt}) (feet ²) | 21.2 |
| | | | | Width/Depth Ratio (W_{bkt}/d_{bkt}) | 4.7 |
| Left Bank View | | Right Bank View | | Width of Flood-prone Area (W_{fpa}) (feet) | 20.9 |
| | | | | Entrenchment Ratio (ER) = W_{fpa}/W_{bkt} | 2.1 |
| Left Bank View | | Right Bank View | | Channel Materials D_{50} (millimeters) | 1.3 |
| | | | | Water Surface Slope (S) | 0.01 |
| Left Bank View | | Right Bank View | | Sinuosity (K) = stream length/valley length | <1.2 |
| | | | | Adjustments? | Sin ↑, ER ↓ |
| Left Bank View | | Right Bank View | | STREAM TYPE | G4c |

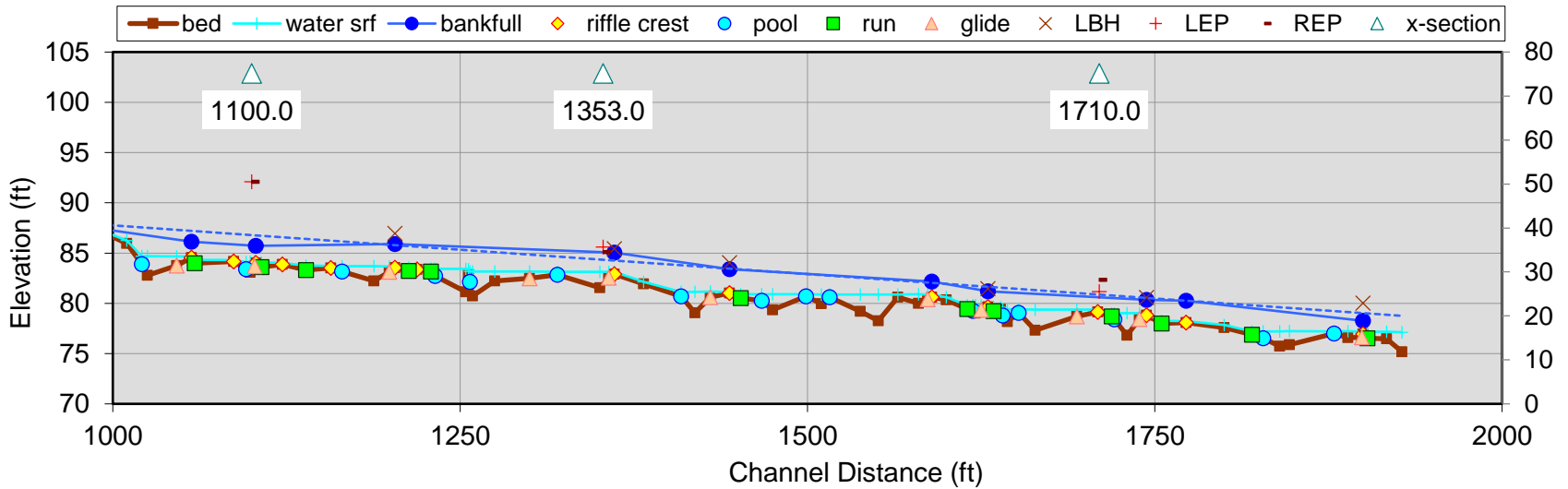
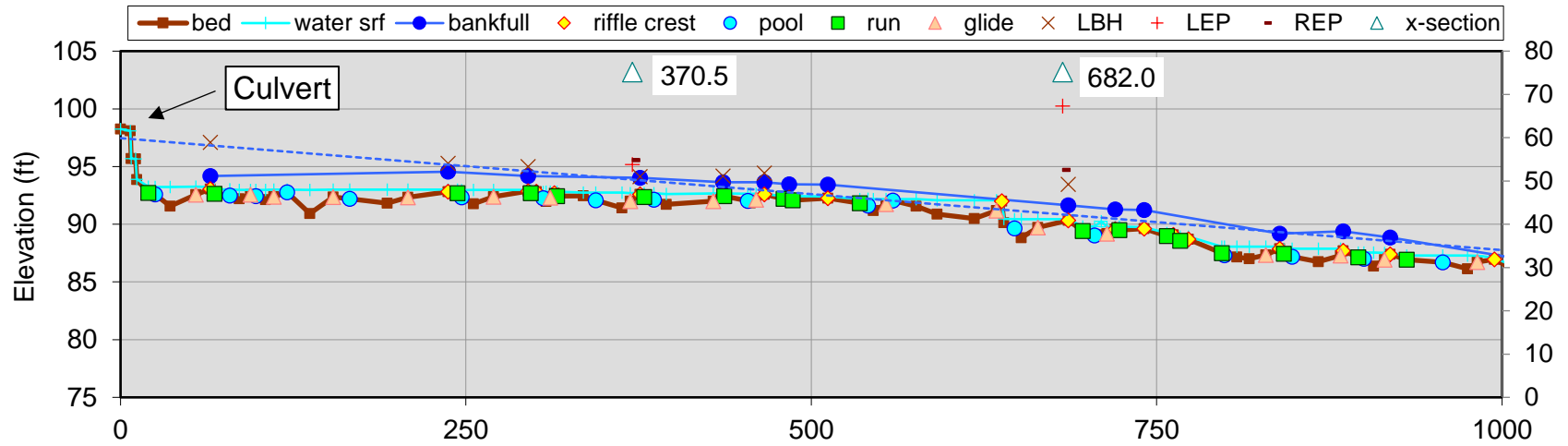


| | | | |
|--|--|--|-------------|
| | | 2017 Geomorphic Assessment Results | |
| | | Bankfull Width (W_{bkf}) (feet) | 8.8 |
| | | Mean Depth (d_{bkf}) (feet) | 1.6 |
| | | Bankfull Cross-sectional Area (A_{bkf}) (feet ²) | 14.4 |
| | | Width/Depth Ratio (W_{bkf}/d_{bkf}) | 5.4 |
| | | Width of Flood-prone Area (W_{fpa}) (feet) | 23.5 |
| | | Entrenchment Ratio (ER) = W_{fpa}/W_{bkf} | 2.7 |
| | | Channel Materials D_{50} (millimeters) | 5.5 |
| | | Water Surface Slope (S) | 0.0028 |
| | | Sinuosity (K) = stream length/valley length | <1.2 |
| | | Adjustments? | Sin ↑ |
| | | STREAM TYPE | E4/5 |

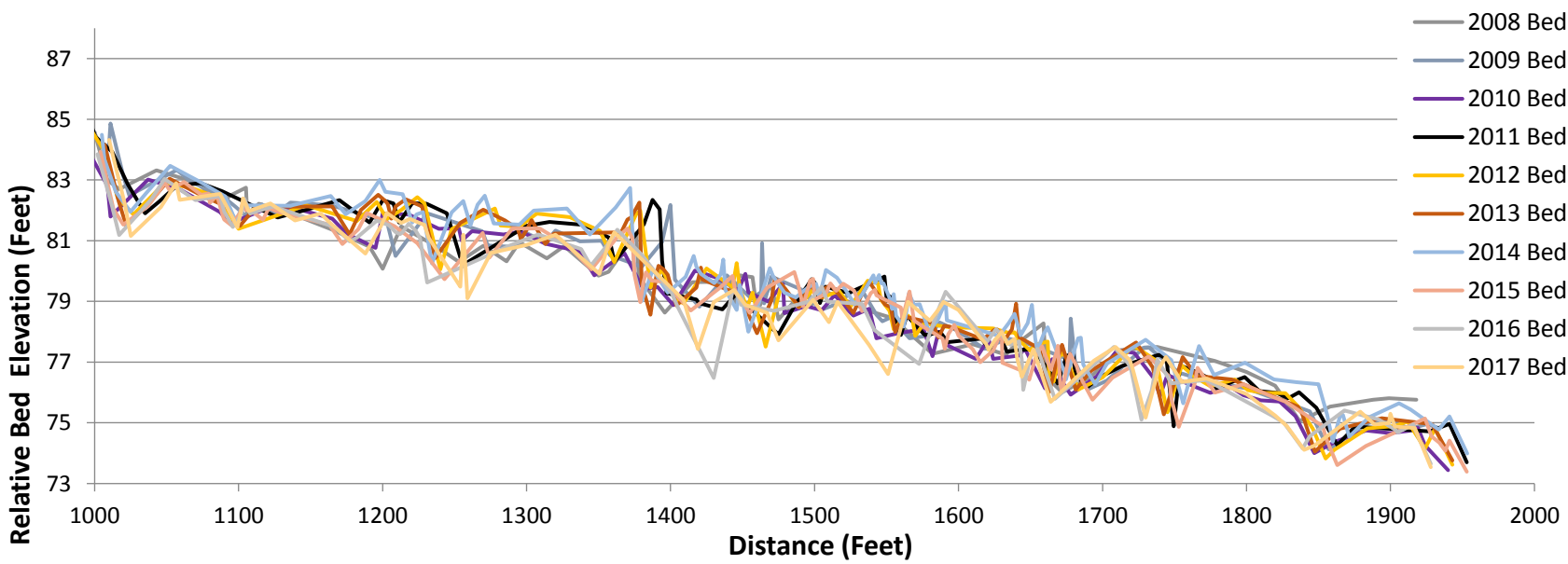
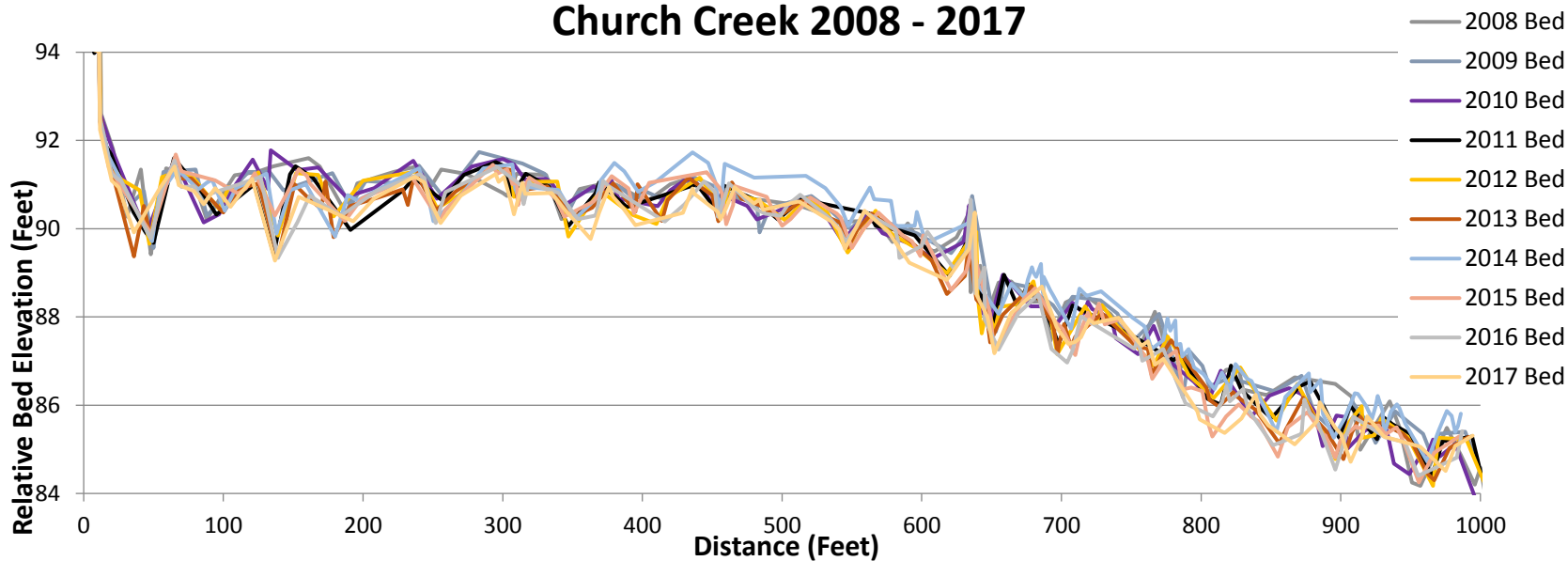


| | | | | | |
|---|--|--|--|--|-------------|
|  | |  | | 2017 Geomorphic Assessment Results | |
| <p style="text-align: center;">Upstream View</p> | | <p style="text-align: center;">Downstream View</p> | | Bankfull Width (W_{bkf}) (feet) | 11.2 |
| | | | | Mean Depth (d_{bkf}) (feet) | 0.8 |
|  | |  | | Bankfull Cross-sectional Area (A_{bkf}) (feet ²) | 9.4 |
| | | | | Width/Depth Ratio (W_{bkf}/d_{bkf}) | 13.4 |
| <p style="text-align: center;">Left Bank View</p> | | <p style="text-align: center;">Right Bank View</p> | | Width of Flood-prone Area (W_{fpa}) (feet) | 19.6 |
| | | | | Entrenchment Ratio (ER) = W_{fpa}/W_{bkf} | 1.7 |
| <p style="text-align: center;">Left Bank View</p> | | <p style="text-align: center;">Right Bank View</p> | | Channel Materials D_{50} (millimeters) | 15 |
| | | | | Water Surface Slope (S) | 0.0079 |
| <p style="text-align: center;">Left Bank View</p> | | <p style="text-align: center;">Right Bank View</p> | | Sinuosity (K) = stream length/valley length | <1.2 |
| | | | | Adjustments? | Sin ↑, ER ↓ |
| | | | | STREAM TYPE | F4 |

Church Creek Longitudinal Profile



Church Creek 2008 - 2017



F-10

APPENDIX G
CHEMICAL MONITORING RESULTS

**Anne Arundel County NPDES
Sampling and EMC Data – 2017 Reporting Year
Parole Plaza Station**

| Sampler | ID | Jurisdiction | Date | Time | Site | Outfall or Instream | Storm or Baseflow | Depth | Duration | Intensity | Temperature - field | Flow | pH - field | BOD | BOD | dt for Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | dt for Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | dt for Total Phosphorus | Total Phosphorus | Total Phosphorus | dt for TSS | TSS | TSS | dt for Copper | Copper | Copper | dt for Lead | Lead | Lead | dt for Zinc | Zinc | Zinc | dt for TPH | TPH | TPH | dt for E-COLI | E-COLI | E-COLI | dt for HARDNESS | HARDNESS | HARDNESS | | |
|----------------------------------|----|--------------|------------|------|------|---------------------|-------------------|--------|----------|-----------|---|-------|-------------|----------|-----------|--------------------------------|-------------------------|-------------------------|-----------------------------|----------------------|----------------------|-------------------------|------------------|------------------|-------------|----------|-----------|---------------|------------|-------------|-------------|------------|------------|-------------|------------|------------|------------|------------|------------|---------------|------------|--------------|-----------------|------------|-----------|-----------|-----------|
| | | | | | | | | Inches | Hours | In/Hr | oF | CF | | mg/L | (l) mg/L | (dt) mg/L | mg/L | (l) mg/L | (dt) mg/L | mg/L | (l) mg/L | (dt) mg/L | mg/L | (l) mg/L | (dt) mg/L | mg/L | (l) mg/L | (dt) mg/L | ug/L | (l) ug/L | (dt) ug/L | ug/L | (l) ug/L | (dt) ug/L | ug/L | (l) ug/L | (dt) ug/L | MPN | (l) MPN | (dt) MPN | mg/L | (l) mg/L | (dt) mg/L | | | | |
| Versar | 1 | AP | 7/28/2016 | 2125 | 101 | O | S | 0.34 | 3.0 | 0.11 | 80.38 | 3262 | 7.28 | 2 | 25 | 25 | 0.5 | 2.7 | 2.7 | 0.05 | 0.61 | 0.61 | 0.01 | 0.14 | 0.14 | 1 | 97 | 97 | 2.0 | 24.1 | 24.1 | 2.0 | 6.2 | 6.2 | 20 | 214 | 214 | 5.0 | 0.0 | 5.0 | 10 | 959 | 959 | 1 | 45 | 45 | |
| Versar | 2 | AP | 7/28/2016 | 2130 | 101 | O | S | | | | 80.22 | 4109 | 7.41 | 2 | 16 | 16 | 0.5 | 1.3 | 1.3 | 0.05 | 0.58 | 0.58 | 0.01 | 0.07 | 0.07 | 1 | 32 | 32 | 2.0 | 18.4 | 18.4 | 2.0 | 2.7 | 2.7 | 20 | 221 | 221 | 5.0 | 0.0 | 5.0 | 10 | 0 | 10 | 1 | 25 | 25 | |
| Versar | 3 | AP | 7/28/2016 | 2155 | 101 | O | S | | | | 79.76 | 13243 | 7.46 | 2 | 4 | 4 | 0.5 | 0.7 | 0.7 | 0.05 | 0.54 | 0.54 | 0.01 | 0.05 | 0.05 | 1 | 4 | 4 | 2.0 | 16.3 | 16.3 | 2.0 | 0.0 | 2.0 | 20 | 115 | 115 | 5.0 | 0.0 | 5.0 | 10 | 0 | 10 | 1 | 20 | 20 | |
| Event Mean Concentration: | | | | | | | | | | | 79.95 | | 7.42 | 2 | 10 | 10 | 0.5 | 1.1 | 1.1 | 0.05 | 0.56 | 0.56 | 0.01 | 0.07 | 0.07 | 1 | 24 | 24 | 2.0 | 18.0 | 18.0 | 2.0 | 1.5 | 2.8 | 2.0 | 152 | 152 | 5.0 | 0.0 | 5.0 | 10 | 152 | 160 | 1 | 25 | 25 | |
| Versar | 1 | AP | 9/19/2016 | 530 | 101 | O | S | 1.22 | 3.0 | 0.41 | 76.41 | 1389 | 8.02 | 4 | 24 | 24 | 0.5 | 2.2 | 2.2 | 0.05 | 0.84 | 0.84 | 0.01 | 0.25 | 0.25 | 1 | 80 | 80 | 2.0 | 34.8 | 34.8 | 2.0 | 4.6 | 4.6 | 20 | 404 | 404 | 5.0 | 0.0 | 5.0 | 100 | 0 | 100 | 1 | 50 | 50 | |
| Versar | 2 | AP | 9/19/2016 | 600 | 101 | O | S | | | | 73.85 | 38968 | 7.99 | 4 | 0 | 4 | 0.5 | 0.8 | 0.8 | 0.05 | 0.19 | 0.19 | 0.01 | 0.17 | 0.17 | 1 | 97 | 97 | 2.0 | 24.1 | 24.1 | 2.0 | 4.7 | 4.7 | 20 | 239 | 239 | 5.0 | 0.0 | 5.0 | 100 | 3550 | 3550 | 1 | 29 | 29 | |
| Versar | 3 | AP | 9/19/2016 | 625 | 101 | O | S | | | | 73.78 | 65408 | 7.69 | 4 | 0 | 4 | 0.5 | 0.8 | 0.8 | 0.05 | 0.66 | 0.66 | 0.01 | 0.16 | 0.16 | 1 | 21 | 21 | 2.0 | 13.0 | 13.0 | 2.0 | 0.0 | 2.0 | 20 | 86 | 86 | 5.0 | 0.0 | 5.0 | 100 | 0 | 100 | 1 | 21 | 21 | |
| Event Mean Concentration: | | | | | | | | | | | 73.84 | | 7.80 | 4 | 0 | 4 | 0.5 | 0.8 | 0.8 | 0.05 | 0.49 | 0.49 | 0.01 | 0.16 | 0.16 | 1 | 50 | 50 | 2.0 | 17.4 | 17.4 | 2.0 | 1.8 | 3.0 | 2.0 | 147 | 147 | 5.0 | 0.0 | 5.0 | 100 | 1308 | 1371 | 1 | 24 | 24 | |
| Versar | 1 | AP | 9/27/2016 | 335 | 101 | O | S | 0.32 | 4.5 | 0.07 | 67.74 | 437 | 7.89 | 4 | 10 | 10 | 0.5 | 3.0 | 3.0 | 0.05 | 0.74 | 0.74 | 0.01 | 0.30 | 0.30 | 1 | 83 | 83 | 2.0 | 14.6 | 14.6 | 2.0 | 0.0 | 2.0 | 20 | 262 | 262 | 5.0 | 0.0 | 5.0 | 100 | 410 | 410 | 1 | 34 | 34 | |
| Versar | 2 | AP | 9/27/2016 | 340 | 101 | O | S | | | | 68.02 | 2297 | 7.98 | 4 | 8 | 8 | 0.5 | 1.6 | 1.6 | 0.05 | 0.43 | 0.43 | 0.01 | 0.15 | 0.15 | 1 | 70 | 70 | 2.0 | 14.9 | 14.9 | 2.0 | 3.4 | 3.4 | 20 | 171 | 171 | 5.0 | 0.0 | 5.0 | 100 | 980 | 980 | 1 | 28 | 28 | |
| Versar | 3 | AP | 9/27/2016 | 610 | 101 | O | S | | | | 73.37 | 15156 | 8.58 | 4 | 0 | 4 | 0.5 | 0.8 | 0.8 | 0.02 | 0.40 | 0.40 | 0.01 | 0.06 | 0.06 | 1 | 2 | 2 | 2.0 | 11.8 | 11.8 | 2.0 | 0.0 | 2.0 | 20 | 116 | 116 | 5.0 | 0.0 | 5.0 | 100 | 46110 | 46110 | 1 | 26 | 26 | |
| Event Mean Concentration: | | | | | | | | | | | 72.55 | | 8.49 | 4 | 1 | 5 | 0.5 | 1.0 | 1.0 | 0.02 | 0.41 | 0.41 | 0.01 | 0.08 | 0.08 | 1 | 13 | 13 | 2.0 | 12.3 | 12.3 | 2.0 | 0.4 | 2.2 | 2.0 | 127 | 127 | 5.0 | 0.0 | 5.0 | 100 | 39201 | 39201 | 1 | 26 | 26 | |
| Versar | 1 | AP | 11/9/2016 | 630 | 101 | O | S | 0.11 | 7.0 | 0.02 | 56.99 | 122 | 8.07 | 4 | 11 | 11 | 0.5 | 1.8 | 1.8 | 0.05 | 1.40 | 1.40 | 0.01 | 0.15 | 0.15 | 1 | 10 | 10 | 2.0 | 18.1 | 18.1 | 2.0 | *** | *** | *** | 20 | 233 | 233 | 5.0 | 0.0 | 5.0 | 10 | 341 | 341 | 1 | 41 | 41 |
| Versar | 2 | AP | 11/9/2016 | 910 | 101 | O | S | | | | 59.75 | 3436 | 7.45 | 4 | 8 | 8 | 0.5 | 1.2 | 1.2 | 0.05 | 0.61 | 0.61 | 0.01 | 0.16 | 0.16 | 1 | 20 | 20 | 2.0 | 22.0 | 22.0 | 2.0 | *** | *** | *** | 20 | 183 | 183 | 5.0 | 0.0 | 5.0 | 10 | 63 | 63 | 1 | 41 | 41 |
| Versar | 3 | AP | 11/9/2016 | 950 | 101 | O | S | | | | 58.92 | 877 | 7.46 | 4 | 6 | 6 | 0.5 | 1.2 | 1.2 | 0.05 | 0.92 | 0.92 | 0.01 | 0.12 | 0.12 | 1 | 12 | 12 | 2.0 | 18.1 | 18.1 | 2.0 | *** | *** | *** | 20 | 166 | 166 | 5.0 | 0.0 | 5.0 | 10 | 645 | 645 | 1 | 59 | 59 |
| Event Mean Concentration: | | | | | | | | | | | 59.51 | | 7.47 | 4 | 8 | 8 | 0.5 | 1.2 | 1.2 | 0.05 | 0.69 | 0.69 | 0.01 | 0.15 | 0.15 | 1 | 18 | 18 | 2.0 | 21.1 | 21.1 | 2.0 | *** | *** | *** | 20 | 181 | 181 | 5.0 | 0.0 | 5.0 | 10 | 186 | 186 | 1 | 45 | 45 |
| Versar | 1 | AP | 11/29/2016 | 1045 | 101 | O | S | 0.08 | 8.0 | 0.01 | 58.01 | 8836 | 7.35 | 4 | 18 | 18 | 0.5 | 2.5 | 2.5 | 0.05 | 3.00 | 3.00 | 0.01 | 0.36 | 0.36 | 1 | 38 | 38 | 2.0 | 39.2 | 39.2 | 2.0 | 4.7 | 4.7 | 20 | 303 | 303 | 5.0 | 6.0 | 6.0 | 10 | 1314 | 1314 | 1 | 140 | 140 | |
| Versar | 2 | AP | 11/29/2016 | 1050 | 101 | O | S | | | | 56.65 | 340 | 7.87 | 4 | 18 | 18 | 0.5 | 2.5 | 2.5 | 0.05 | 1.40 | 1.40 | 0.01 | 0.49 | 0.49 | 1 | 89 | 89 | 2.0 | 38.3 | 38.3 | 2.0 | 5.5 | 5.5 | 20 | 663 | 663 | 5.0 | 7.0 | 7.0 | 10 | 24196 | 24196 | 1 | 50 | 50 | |
| Versar | 3 | AP | 11/29/2016 | 1145 | 101 | O | S | | | | 57.15 | 2576 | 7.80 | 4 | 12 | 12 | 0.5 | 1.9 | 1.9 | 0.05 | 1.40 | 1.40 | 0.01 | 0.30 | 0.30 | 1 | 35 | 35 | 2.0 | 40.5 | 40.5 | 2.0 | 3.1 | 3.1 | 20 | 249 | 249 | 5.0 | 10.0 | 10.0 | 10 | 4611 | 4611 | 1 | 76 | 76 | |
| Event Mean Concentration: | | | | | | | | | | | No EMCs were calculated for this storm – leaf litter in RCP affected flow record | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| Versar | 1 | AP | 12/6/2016 | 1410 | 101 | O | S | 0.78 | 12.0 | 0.07 | 47.24 | 4990 | 8.65 | 4 | 6 | 6 | 0.5 | 1.2 | 1.2 | 0.05 | 0.44 | 0.44 | 0.01 | 0.15 | 0.15 | 1 | 62 | 62 | 2.0 | 27.1 | 27.1 | 2.0 | 4.0 | 4.0 | 20 | 228 | 228 | 5.0 | 6.0 | 6.0 | 100 | 850 | 850 | 1 | 60 | 60 | |
| Versar | 2 | AP | 12/6/2016 | 1520 | 101 | O | S | | | | 45.88 | 16642 | 8.92 | 4 | 0 | 4 | 0.5 | 0.8 | 0.8 | 0.05 | 0.22 | 0.22 | 0.01 | 0.10 | 0.10 | 1 | 36 | 36 | 2.0 | 17.0 | 17.0 | 2.0 | 2.4 | 2.4 | 20 | 133 | 133 | 5.0 | 0.0 | 5.0 | 100 | 4960 | 4960 | 1 | 30 | 30 | |
| Versar | 3 | AP | 12/6/2016 | 1915 | 101 | O | S | | | | 48.31 | 68958 | 8.59 | 4 | 0 | 4 | 0.5 | 0.7 | 0.7 | 0.05 | 0.36 | 0.36 | 0.01 | 0.09 | 0.09 | 1 | 10 | 10 | 2.0 | 10.7 | 10.7 | 2.0 | 0.0 | 2.0 | 20 | 80 | 80 | 5.0 | 0.0 | 5.0 | 100 | 2180 | 2180 | 1 | 30 | 30 | |
| Event Mean Concentration: | | | | | | | | | | | 47.80 | | 8.65 | 4 | 0 | 4 | 0.5 | 0.7 | 0.7 | 0.05 | 0.34 | 0.34 | 0.01 | 0.10 | 0.10 | 1 | 18 | 18 | 2.0 | 12.8 | 12.8 | 2.0 | 0.7 | 2.2 | 2.0 | 98 | 98 | 5.0 | 0.3 | 5.1 | 100 | 2617 | 2617 | 1 | 32 | 32 | |
| Versar | 1 | AP | 1/23/2017 | 1125 | 101 | O | S | 0.49 | 12.0 | 0.04 | 47.25 | 5161 | 7.70 | 2 | 5 | 5 | 0.5 | 1.0 | 1.0 | 0.05 | 0.29 | 0.29 | 0.01 | 0.08 | 0.08 | 1 | 94 | 94 | 2.0 | 32.3 | 32.3 | 2.0 | 5.1 | 5.1 | 20 | 300 | 300 | 5.0 | 0.0 | 5.0 | 10 | 3255 | 3255 | 1 | 79 | 79 | |
| Versar | 2 | AP | 1/23/2017 | 1410 | 101 | O | S | | | | 43.73 | 31808 | 7.46 | 2 | 6 | 6 | 0.5 | 1.9 | 1.9 | 0.05 | 0.17 | 0.17 | 0.01 | 0.29 | 0.29 | 1 | 340 | 340 | 2.0 | 62.5 | 62.5 | 2.0 | 14.8 | 14.8 | 20 | 341 | 341 | 5.0 | 0.0 | 5.0 | 10 | 3448 | 3448 | 1 | 140 | 140 | |
| Versar | 3 | AP | 1/23/2017 | 1610 | 101 | O | S | | | | 44.97 | 37197 | 6.65 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.40 | 0.40 | 0.01 | 0.13 | 0.13 | 1 | 58 | 58 | 2.0 | 15.2 | 15.2 | 2.0 | 2.0 | 2.0 | 20 | 120 | 120 | 5.0 | 0.0 | 5.0 | 10 | 9804 | 9804 | 1 | 83 | 83 | |
| Event Mean Concentration: | | | | | | | | | | | 44.59 | | 7.07 | 2 | 3 | 4 | 0.5 | 0.9 | 1.1 | 0.05 | 0.29 | 0.29 | 0.01 | 0.20 | | | | | | | | | | | | | | | | | | | | | | | |

| Sampler | ID | Jurisdiction | Date | Time | Site | Outfall or Instream | Storm or Baseflow | Depth | Duration | Intensity | Temperature - field | Flow | pH - field | dt for BOD | BOD | BOD | dt for Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | dt for Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | dt for Total Phosphorus | Total Phosphorus | Total Phosphorus | dt for TSS | TSS | TSS | dt for Copper | Copper | Copper | dt for Lead | Lead | Lead | dt for Zinc | Zinc | Zinc | dt for TPH | TPH | TPH | dt for E-COLI | E-COLI | E-COLI | dt for HARDNESS | HARDNESS | HARDNESS |
|---|----|--------------|-----------|------|------|---------------------|-------------------|--------|----------|-----------|---------------------|-------|------------|------------|----------|-----------|--------------------------------|-------------------------|-------------------------|-----------------------------|----------------------|----------------------|-------------------------|------------------|------------------|------------|----------|-----------|---------------|----------|-----------|-------------|----------|-----------|-------------|----------|-----------|------------|----------|-----------|---------------|---------|----------|-----------------|----------|-----------|
| | | | | | | | | Inches | Hours | In/Hr | oF | CF | pH | mg/L | (0) mg/L | (dt) mg/L | mg/L | (0) mg/L | (dt) mg/L | mg/L | (0) mg/L | (dt) mg/L | mg/L | (0) mg/L | (dt) mg/L | mg/L | (0) mg/L | (dt) mg/L | µg/L | (0) µg/L | (dt) µg/L | µg/L | (0) µg/L | (dt) µg/L | µg/L | (0) µg/L | (dt) µg/L | mg/L | (0) mg/L | (dt) mg/L | MPN | (0) MPN | (dt) MPN | mg/L | (0) mg/L | (dt) mg/L |
| Versar | 1 | AP | 4/25/2017 | 1220 | 101 | O | S | 0.39 | 5.0 | 0.08 | 58.31 | 3449 | 7.75 | 2 | 10 | 10 | 0.5 | 2.0 | 2.0 | 0.05 | 0.49 | 0.49 | 0.01 | 0.28 | 0.28 | 1 | 64 | 64 | 2.0 | 40.0 | 40.0 | 2.0 | 3.5 | 3.5 | 20 | 223 | 223 | 5.0 | 7.0 | 7.0 | 1 | 1482 | 1482 | 1 | 64 | 64 |
| Versar | 2 | AP | 4/25/2017 | 1235 | 101 | O | S | | | | 60.36 | 6738 | 8.00 | 2 | 25 | 25 | 0.5 | 2.9 | 2.9 | 0.05 | 0.15 | 0.15 | 0.01 | 0.46 | 0.46 | 1 | 230 | 230 | 2.0 | 135.0 | 135.0 | 2.0 | 22.3 | 22.3 | 20 | 1150 | 1150 | 5.0 | 7.0 | 7.0 | 1 | 2292 | 2292 | 1 | 170 | 170 |
| Versar | 3 | AP | 4/25/2017 | 1435 | 101 | O | S | | | | 59.79 | 46059 | 7.81 | 2 | 3 | 3 | 0.5 | 0.8 | 0.8 | 0.05 | 0.29 | 0.29 | 0.01 | 0.12 | 0.12 | 1 | 10 | 10 | 2.0 | 11.8 | 11.8 | 2.0 | 0.0 | 2.0 | 20 | 70 | 70 | 5.0 | 0.0 | 5.0 | 1 | 2420 | 2420 | 1 | 27 | 27 |
| Event Mean Concentration: | | | | | | | | | | | 59.76 | | 7.83 | 2 | 6 | 6 | 0.5 | 1.1 | 1.1 | 0.05 | 0.29 | 0.29 | 0.01 | 0.17 | 0.17 | 1 | 40 | 40 | 2.0 | 28.3 | 28.3 | 2.0 | 2.9 | 4.5 | 20 | 209 | 209 | 5.0 | 1.3 | 5.4 | | ** | ** | 1 | 46 | 46 |
| Versar | 1 | AP | 5/12/2017 | 2230 | 101 | O | S | 0.98 | 21.0 | 0.05 | 56.87 | 8676 | 7.81 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.40 | 0.40 | 0.01 | 0.06 | 0.06 | 1 | 10 | 10 | 2.0 | 12.9 | 12.9 | 2.0 | 2.8 | 2.8 | 20 | 97 | 97 | 5.0 | 6.0 | 6.0 | 1 | 3637 | 3637 | 1 | 32 | 32 |
| Versar | 2 | AP | 5/13/2017 | 335 | 101 | O | S | | | | 54.09 | 42377 | 7.85 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.15 | 0.15 | 0.01 | 0.05 | 0.05 | 1 | 10 | 10 | 2.0 | 6.7 | 6.7 | 2.0 | 0.0 | 2.0 | 20 | 67 | 67 | 5.0 | 0.0 | 5.0 | 1 | 7600 | 7600 | 1 | 18 | 18 |
| Versar | 3 | AP | 5/13/2017 | 730 | 101 | O | S | | | | 52.93 | 50746 | 7.64 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.24 | 0.24 | 0.01 | 0.06 | 0.06 | 1 | 3 | 3 | 2.0 | 6.4 | 6.4 | 2.0 | 0.0 | 2.0 | 20 | 63 | 63 | 5.0 | 0.0 | 5.0 | 1 | 7811 | 7811 | 1 | 26 | 26 |
| Event Mean Concentration: | | | | | | | | | | | 53.75 | | 7.74 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.22 | 0.22 | 0.01 | 0.06 | 0.06 | 1 | 7 | 7 | 2.0 | 7.1 | 7.1 | 2.0 | 0.2 | 2.1 | 20 | 68 | 68 | 5.0 | 0.5 | 5.1 | 1 | 7367 | 7367 | 1 | 23 | 23 |
| Versar | 1 | AP | 6/19/2017 | 1625 | 101 | O | S | 0.36 | 4.0 | 0.09 | 81.37 | 125 | 7.39 | 2 | 10 | 10 | 0.5 | 2.4 | 2.4 | 0.05 | 0.65 | 0.65 | 0.01 | 0.31 | 0.31 | 1 | 530 | 530 | 2.0 | 200.0 | 200.0 | 2.0 | 55.0 | 55.0 | 20 | 1810 | 1810 | 5.0 | 9.0 | 9.0 | 1 | 1247 | 1247 | 1 | 400 | 400 |
| Versar | 2 | AP | 6/19/2017 | 1630 | 101 | O | S | | | | 78.62 | 2559 | 7.35 | 2 | 10 | 10 | 0.5 | 2.2 | 2.2 | 0.05 | 0.51 | 0.51 | 0.01 | 0.30 | 0.30 | 1 | 200 | 200 | 2.0 | 144.0 | 144.0 | 2.0 | 12.0 | 12.0 | 20 | 993 | 993 | 5.0 | 7.0 | 7.0 | 1 | 4553 | 4553 | 1 | 68 | 68 |
| Versar | 3 | AP | 6/19/2017 | 1735 | 101 | O | S | | | | 77.66 | 19896 | 7.40 | 2 | 3 | 3 | 0.5 | 0.8 | 0.8 | 0.05 | 0.53 | 0.53 | 0.01 | 0.11 | 0.11 | 1 | 22 | 22 | 2.0 | 23.0 | 23.0 | 2.0 | 0.0 | 2.0 | 20 | 163 | 163 | 5.0 | 5.0 | 5.0 | 1 | 65044 | 65044 | 1 | 28 | 28 |
| Event Mean Concentration: | | | | | | | | | | | 77.79 | | 7.40 | 2 | 4 | 4 | 0.5 | 1.0 | 1.0 | 0.05 | 0.53 | 0.53 | 0.01 | 0.13 | 0.13 | 1 | 45 | 45 | 2.0 | 37.7 | 37.7 | 2.0 | 1.7 | 3.4 | 20 | 266 | 266 | 5.0 | 5.2 | 5.2 | 1 | 57837 | 57837 | 1 | 35 | 35 |
| * No pH readings taken from CMP so volume-weighted pH composite could not be calculated. Values are for RCP only. | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| ** Samples were not diluted - the data has been discarded and not included in subsequent flow weighted EMC and loading calculations | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| *** Error in results - the data has been discarded and not included in subsequent flow weighted EMC and loading calculations | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |

**Anne Arundel County NPDES
Sampling and EMC Data – 2017 Reporting Year
Parole Plaza Station**

| | Temperature - field pF | pH - field pH | BOD (0) mg/L | BOD (dt) mg/L | Total Kjeldahl Nitrogen (0) mg/L | Total Kjeldahl Nitrogen (dt) mg/L | Nitrate+ Nitrite - N (0) mg/L | Nitrate+ Nitrite - N (dt) mg/L | Total Phosphorus (0) mg/L | Total Phosphorus (dt) mg/L | TSS (0) mg/L | TSS (dt) mg/L | Copper (0) µg/L | Copper (dt) µg/L | Lead (0) µg/L | Lead (dt) µg/L | Zinc (0) µg/L | Zinc (dt) µg/L | TPH (0) mg/L | TPH (dt) mg/L | E-COLI (0) MPN | E-COLI (dt) MPN | HARDNESS (0) mg/L | HARDNESS (dt) mg/L |
|---|---------------------------|------------------|-----------------|------------------|-------------------------------------|--------------------------------------|----------------------------------|-----------------------------------|------------------------------|-------------------------------|-----------------|------------------|--------------------|---------------------|------------------|-------------------|------------------|-------------------|-----------------|------------------|-------------------|--------------------|----------------------|-----------------------|
| Summer Quarter Flow-Weighted EMC (7/28/16, 9/19/16, 9/27/16): | 74.56 | 7.83 | 1.78 | 5.09 | 0.88 | 0.88 | 0.49 | 0.49 | 0.14 | 0.14 | 41.54 | 41.54 | 16.82 | 16.82 | 1.89 | 2.89 | 145.01 | 145.01 | 0.00 | 5.00 | 5841.28 | 5888.78 | 24.68 | 24.68 |
| Average: | | | 3.43 | mg/l | 0.88 | mg/l | 0.49 | mg/l | 0.14 | mg/l | 41.54 | mg/l | 16.82 | µg/l | 2.24 | µg/l | 145.01 | µg/l | 2.5 | mg/l | 5865.03 | MPN/100mL | 24.68 | mg/l |
| | | | 0.0002143 | lb/cf | 0.0000550 | lb/cf | 0.0000306 | lb/cf | 0.0000088 | lb/cf | 0.0025927 | lb/cf | 0.0000011 | lb/cf | 0.0000001 | lb/cf | 0.0000091 | lb/cf | 0.0001560 | lb/cf | | | 0.0015405 | lb/cf |
| Total Volume (Quarter Events): | | | 144,269 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 30.9 | lbs | 7.9 | lbs | 4.4 | lbs | 1.3 | lbs | 374.0 | lbs | 0.2 | lbs | 0.02 | lbs | 1.31 | lbs | 22.51 | lbs | | | 222.2 | lbs |
| Total Volume (Quarter): | | | 1,602,889 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 343.6 | lbs | 88.1 | lbs | 49.0 | lbs | 14.0 | lbs | 4,155.8 | lbs | 1.7 | lbs | 0.2 | lbs | 14.5 | lbs | 250.1 | lbs | | | 2,469.2 | lbs |
| Fall Quarter Flow-Weighted EMC (11/9/16, 11/29/16, 12/6/17): | 48.35 | 8.60 | 0.67 | 4.28 | 0.77 | 0.77 | 0.36 | 0.36 | 0.10 | 0.10 | 17.66 | 17.66 | 13.15 | 13.15 | 0.65 | 2.18 | 101.84 | 101.84 | 0.32 | 5.05 | 2503.94 | 2503.94 | 32.25 | 32.25 |
| Average: | | | 2.48 | mg/l | 0.77 | mg/l | 0.36 | mg/l | 0.10 | mg/l | 17.66 | mg/l | 13.15 | µg/l | 1.41 | µg/l | 101.84 | µg/l | 2.68 | mg/l | 2503.94 | MPN/100mL | 32.25 | mg/l |
| | | | 0.0001545 | lb/cf | 0.0000479 | lb/cf | 0.0000222 | lb/cf | 0.0000061 | lb/cf | 0.0011025 | lb/cf | 0.0000008 | lb/cf | 0.0000001 | lb/cf | 0.0000064 | lb/cf | 0.0001675 | lb/cf | | | 0.0020132 | lb/cf |
| Total Volume (Quarter Events): | | | 95,025 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 14.7 | lbs | 4.6 | lbs | 2.1 | lbs | 0.6 | lbs | 104.8 | lbs | 0.1 | lbs | 0.0 | lbs | 0.6 | lbs | 15.9 | lbs | | | 191.3 | lbs |
| Total Volume (Quarter): | | | 1,235,323 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 190.9 | lbs | 59.2 | lbs | 27.4 | lbs | 7.5 | lbs | 1,362.0 | lbs | 1.0 | lbs | 0.1 | lbs | 7.9 | lbs | 206.9 | lbs | | | 2,487.0 | lbs |
| Winter Quarter Flow-Weighted EMC (1/23/17, 2/28/17, 3/31/17): | 49.66 | 7.33 | 2.31 | 3.41 | 0.85 | 0.97 | 0.41 | 0.41 | 0.15 | 0.15 | 95.16 | 95.16 | 29.57 | 29.57 | 4.55 | 5.55 | 187.34 | 187.34 | 0.15 | 5.01 | 3971.33 | 3971.33 | 74.75 | 74.75 |
| Average: | | | 2.86 | mg/l | 0.91 | mg/l | 0.41 | mg/l | 0.15 | mg/l | 95.16 | mg/l | 29.57 | µg/l | 5.05 | µg/l | 187.34 | µg/l | 2.58 | mg/l | 3971.33 | MPN/100mL | 74.75 | mg/l |
| | | | 0.0001786 | lb/cf | 0.0000566 | lb/cf | 0.0000258 | lb/cf | 0.0000097 | lb/cf | 0.0059395 | lb/cf | 0.0000018 | lb/cf | 0.0000003 | lb/cf | 0.0000117 | lb/cf | 0.0001610 | lb/cf | | | 0.0046654 | lb/cf |
| Total Volume (Quarter Events): | | | 246,485 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 44.0 | lbs | 14.0 | lbs | 6.4 | lbs | 2.4 | lbs | 1,464.0 | lbs | 0.5 | lbs | 0.1 | lbs | 2.9 | lbs | 39.7 | lbs | | | 1,149.9 | lbs |
| Total Volume (Quarter): | | | 2,585,072 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 461.6 | lbs | 146.3 | lbs | 66.6 | lbs | 25.0 | lbs | 15,354.1 | lbs | 4.8 | lbs | 0.8 | lbs | 30.2 | lbs | 416.2 | lbs | | | 12,060.4 | lbs |
| Spring Quarter Flow-Weighted EMC (4/25/17, 5/12/17, 6/19/17): | 58.63 | 7.73 | 2.37 | 3.49 | 0.47 | 0.75 | 0.28 | 0.28 | 0.10 | 0.10 | 21.64 | 21.64 | 17.51 | 17.51 | 1.24 | 3.00 | 136.33 | 136.33 | 1.31 | 5.16 | 16529.53 | 16529.53 | 31.84 | 31.84 |
| Average: | | | 2.93 | mg/l | 0.61 | mg/l | 0.28 | mg/l | 0.10 | mg/l | 21.64 | mg/l | 17.51 | µg/l | 2.12 | µg/l | 136.33 | µg/l | 3.23 | mg/l | 16529.53 | MPN/100mL | 31.84 | mg/l |
| | | | 0.0001829 | lb/cf | 0.0000382 | lb/cf | 0.0000173 | lb/cf | 0.0000063 | lb/cf | 0.0013509 | lb/cf | 0.0000011 | lb/cf | 0.0000001 | lb/cf | 0.0000085 | lb/cf | 0.0002019 | lb/cf | | | 0.0019871 | lb/cf |
| Total Volume (Quarter Events): | | | 180,624 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 33.0 | lbs | 6.9 | lbs | 3.1 | lbs | 1.1 | lbs | 244.0 | lbs | 0.2 | lbs | 0.0 | lbs | 1.5 | lbs | 36.5 | lbs | | | 358.9 | lbs |
| Total Volume (Quarter): | | | 1,369,592 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 250.6 | lbs | 52.3 | lbs | 23.7 | lbs | 8.6 | lbs | 1,850.2 | lbs | 1.5 | lbs | 0.2 | lbs | 11.7 | lbs | 276.5 | lbs | | | 2,721.5 | lbs |
| AVERAGE ANNUAL EMCs: | 57.29 | 7.73 | 1.98 | mg/l | 0.74 | mg/l | 0.38 | mg/l | 0.13 | mg/l | 52.58 | mg/l | 21.20 | µg/l | 2.45 | µg/l | 152.16 | µg/l | 0.46 | mg/l | 7570.73 | mg/l | 46.22 | mg/l |
| TOTAL ANNUAL POLLUTANT LOAD (EVENTS): | | | 122.66 | lbs | 33.34 | lbs | 15.99 | lbs | 5.37 | lbs | 2,186.81 | lbs | 0.88 | lbs | 0.13 | lbs | 6.33 | lbs | 114.58 | lbs | | | 1,922.42 | lbs |
| Per Acre: | | | 2.03 | | 0.55 | | 0.26 | | 0.09 | | 36.20 | | 0.014 | | 0.002 | | 0.105 | | 1.90 | | | | 31.82 | |
| TOTAL 2017 POLLUTANT LOAD: | | | 1,246.59 | lbs | 345.98 | lbs | 166.67 | lbs | 55.22 | lbs | 22,722.01 | lbs | 8.97 | lbs | 1.33 | lbs | 64.24 | lbs | 1,149.74 | lbs | | | 19,738.13 | lbs |

Anne Arundel County NPDES Sampling and EMC Data – 2017 Reporting Year Church Creek Station

| Sampler | ID | Jurisdiction | Date | Time | Site | Outfall or Instream | Storm or Baseflow | Depth | Duration | In/Hr | Temperature - field | Flow | pH - field | dt for BOD | BOD | BOD | dt for Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | dt for Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | dt for Total Phosphorus | Total Phosphorus | Total Phosphorus | dt for TSS | TSS | TSS | dt for Copper | Copper | Copper | dt for Lead | Lead | Lead | dt for Zinc | Zinc | Zinc | dt for TPH | TPH | TPH | dt for E-COLI | E-COLI | E-COLI | dt for HARDNESS | HARDNESS | HARDNESS |
|---------------------------|----|--------------|------------|------|------|---------------------|-------------------|-------|----------|-------|---------------------|-----------|------------|------------|-----|-----|--------------------------------|-------------------------|-------------------------|-----------------------------|----------------------|----------------------|-------------------------|------------------|------------------|------------|-----|-----|---------------|--------|--------|-------------|------|------|-------------|------|------|------------|-----|-----|---------------|--------|--------|-----------------|----------|----------|
| Versar | 1 | AC | 7/28/2016 | 2130 | 102 | I | S | 0.34 | 3.0 | 0.11 | 78.98 | 10952 | 6.9 | 2 | 9 | 9 | 0.5 | 1.8 | 1.8 | 0.05 | 0.81 | 0.81 | 0.01 | 0.25 | 0.25 | 1 | 29 | 29 | 2.0 | 27.4 | 27.4 | 2.0 | 15.3 | 15.3 | 20 | 217 | 217 | 5.0 | 0.0 | 5.0 | 10 | 1396 | 1396 | 1 | 85 | 85 |
| Versar | 2 | AC | 7/28/2016 | 2155 | 102 | I | S | | | | 79.52 | 129017 | 6.9 | 2 | 20 | 20 | 0.5 | 1.1 | 1.1 | 0.05 | 0.67 | 0.67 | 0.01 | 0.19 | 0.19 | 1 | 72 | 72 | 2.0 | 17.2 | 17.2 | 2.0 | 13.0 | 13.0 | 20 | 122 | 122 | 5.0 | 0.0 | 5.0 | 10 | 73 | 73 | 1 | 36 | 36 |
| Versar | 3 | AC | 7/28/2016 | 2340 | 102 | I | S | | | | 80.42 | 225359 | 6.6 | 2 | 12 | 12 | 0.5 | 0.9 | 0.9 | 0.05 | 0.59 | 0.59 | 0.01 | 0.06 | 0.06 | 1 | 10 | 10 | 2.0 | 5.9 | 5.9 | 2.0 | 0.0 | 13.0 | 20 | 37 | 37 | 5.0 | 0.0 | 5.0 | 10 | 0 | 10 | 1 | 41 | 41 |
| Event Mean Concentration: | | | | | | | | | | | 80.06 | | 6.71 | 2 | 15 | 15 | 0.5 | 1.0 | 1.0 | 0.05 | 0.62 | 0.62 | 0.01 | 0.11 | 0.11 | 1 | 32 | 32 | 2.0 | 10.5 | 10.5 | 2.0 | 5.0 | 13.1 | 20 | 72 | 72 | 5.0 | 0.0 | 5.0 | 10 | 68 | 74 | 1 | 41 | 41 |
| Versar | 1 | AC | 9/19/2016 | 605 | 102 | I | S | 1.22 | 3.0 | 0.41 | 74.12 | 89836 | 6.7 | 4 | 9 | 9 | 0.5 | 1.8 | 1.8 | 0.05 | 0.24 | 0.24 | 0.01 | 0.33 | 0.33 | 1 | 190 | 190 | 2.0 | 8.2 | 8.2 | 2.0 | 7.3 | 7.3 | 20 | 74 | 74 | 5.0 | 0.0 | 5.0 | 100 | 15530 | 15530 | 1 | NA | NA |
| Versar | 2 | AC | 9/19/2016 | 615 | 102 | I | S | | | | 73.76 | 175729 | 6.4 | 4 | 0 | 4 | 0.5 | 0.9 | 0.9 | 0.05 | 0.21 | 0.21 | 0.01 | 0.20 | 0.20 | 1 | 90 | 90 | 2.0 | 9.7 | 9.7 | 2.0 | 7.9 | 7.9 | 20 | 77 | 77 | 5.0 | 0.0 | 5.0 | 100 | 11260 | 11260 | 1 | NA | NA |
| Versar | 3 | AC | 9/19/2016 | 705 | 102 | I | S | | | | 73.76 | 571547 | 6.6 | 4 | 0 | 4 | 0.5 | 0.6 | 0.6 | 0.05 | 0.35 | 0.35 | 0.01 | 0.15 | 0.15 | 1 | 42 | 42 | 2.0 | 8.1 | 8.1 | 2.0 | 8.1 | 8.1 | 20 | 72 | 72 | 5.0 | 0.0 | 5.0 | 100 | 12910 | 12910 | 1 | NA | NA |
| Event Mean Concentration: | | | | | | | | | | | 73.80 | | 6.57 | 4 | 1 | 5 | 0.5 | 0.8 | 0.8 | 0.05 | 0.31 | 0.31 | 0.01 | 0.18 | 0.18 | 1 | 68 | 68 | 2.0 | 8.5 | 8.5 | 2.0 | 8.0 | 8.0 | 20 | 73 | 73 | 5.0 | 0.0 | 5.0 | 100 | 12845 | 12845 | 1 | NA | NA |
| Versar | 1 | AC | 9/27/2016 | 355 | 102 | I | S | 1.79 | 4.5 | 0.40 | 66.02 | 8439 | 6.7 | 4 | 13 | 13 | 0.5 | 4.4 | 4.4 | 0.05 | 0.30 | 0.30 | 0.01 | 1.00 | 1.00 | 1 | 400 | 400 | 2.0 | 36.0 | 36.0 | 2.0 | 20.4 | 20.4 | 20 | 320 | 320 | 5.0 | 0.0 | 5.0 | 100 | 91000 | 91000 | 1 | 72 | 72 |
| Versar | 2 | AC | 9/27/2016 | 430 | 102 | I | S | | | | 67.28 | 55894 | 6.6 | 4 | 4 | 4 | 0.5 | 1.7 | 1.7 | 0.05 | 0.20 | 0.20 | 0.01 | 0.24 | 0.24 | 1 | 70 | 70 | 2.0 | 12.8 | 12.8 | 2.0 | 8.9 | 8.9 | 20 | 104 | 104 | 5.0 | 0.0 | 5.0 | 100 | 17850 | 17850 | 1 | 51 | 51 |
| Versar | 3 | AC | 9/27/2016 | 545 | 102 | I | S | | | | 67.64 | 72710 | 6.7 | 4 | 0 | 4 | 0.5 | 1.0 | 1.0 | 0.05 | 0.25 | 0.25 | 0.01 | 0.11 | 0.11 | 1 | 23 | 23 | 2.0 | 6.3 | 6.3 | 2.0 | 2.2 | 2.2 | 20 | 45 | 45 | 5.0 | 0.0 | 5.0 | 100 | 12110 | 12110 | 1 | 44 | 44 |
| Event Mean Concentration: | | | | | | | | | | | 67.39 | | 6.66 | 4 | 2 | 5 | 0.5 | 1.5 | 1.5 | 0.05 | 0.23 | 0.23 | 0.01 | 0.22 | 0.22 | 1 | 65 | 65 | 2.0 | 10.8 | 10.8 | 2.0 | 6.0 | 6.0 | 20 | 86 | 86 | 5.0 | 0.0 | 5.0 | 100 | 19309 | 19309 | 1 | 49 | 49 |
| Versar | 1 | AC | 11/9/2016 | 855 | 102 | I | S | 0.11 | 7.0 | 0.02 | 52.70 | 27598 | 6.3 | 4 | 7 | 7 | 0.5 | 3.6 | 3.6 | 0.05 | 1.70 | 1.70 | 0.01 | 0.89 | 0.89 | 1 | 52 | 52 | 2.0 | 4.6 | 4.6 | *** | *** | *** | 20 | 70 | 70 | 5.0 | 0.0 | 5.0 | 10 | 1081 | 1081 | 1 | 140 | 140 |
| Versar | 2 | AC | 11/9/2016 | 950 | 102 | I | S | | | | 53.60 | 16613 | 6.3 | 4 | 12 | 12 | 0.5 | 1.3 | 1.3 | 0.05 | 1.40 | 1.40 | 0.01 | 0.35 | 0.35 | 1 | 160 | 160 | 2.0 | 11.1 | 11.1 | *** | *** | *** | 20 | 154 | 154 | 5.0 | 0.0 | 5.0 | 10 | 10462 | 10462 | 1 | 150 | 150 |
| Versar | 3 | AC | 11/9/2016 | 1125 | 102 | I | S | | | | 55.58 | 27549 | 6.3 | 4 | 12 | 12 | 0.5 | 1.7 | 1.7 | 0.05 | 1.40 | 1.40 | 0.01 | 0.12 | 0.12 | 1 | 11 | 11 | 2.0 | 8.4 | 8.4 | *** | *** | *** | 20 | 70 | 70 | 5.0 | 0.0 | 5.0 | 10 | 209 | 209 | 1 | 100 | 100 |
| Event Mean Concentration: | | | | | | | | | | | 54.01 | | 6.30 | 4 | 10 | 10 | 0.5 | 2.3 | 2.3 | 0.05 | 1.52 | 1.52 | 0.01 | 0.47 | 0.47 | 1 | 61 | 61 | 2.0 | 7.6 | 7.6 | *** | *** | *** | 20 | 89 | 89 | 5.0 | 0.0 | 5.0 | 10 | 2918 | 2918 | 1 | 127 | 127 |
| Versar | 1 | AC | 11/29/2016 | 1115 | 102 | I | S | 0.08 | 8.0 | 0.01 | 48.74 | 58583 | 6.8 | 4 | 0 | 4 | 0.5 | 0.0 | 0.5 | 0.05 | 0.91 | 0.91 | 0.01 | 0.09 | 0.09 | 1 | 14 | 14 | 2.0 | 5.8 | 5.8 | 2.0 | 0.0 | 2.0 | 20 | 61 | 61 | 5.0 | 0.0 | 5.0 | 10 | 52 | 52 | 1 | 150 | 150 |
| Versar | 2 | AC | 11/29/2016 | 1205 | 102 | I | S | | | | 49.46 | 10486 | 6.8 | 4 | 4 | 4 | 0.5 | 0.9 | 0.9 | 0.05 | 0.96 | 0.96 | 0.01 | 2.60 | 2.60 | 1 | 24 | 24 | 2.0 | 3.5 | 3.5 | 2.0 | 0.0 | 2.0 | 20 | 52 | 52 | 5.0 | 0.0 | 5.0 | 10 | 146 | 146 | 1 | 140 | 140 |
| Versar | 3 | AC | 11/29/2016 | 1220 | 102 | I | S | | | | 49.46 | 3039 | 6.8 | 4 | 7 | 7 | 0.5 | 1.0 | 1.0 | 0.05 | 0.96 | 0.96 | 0.01 | 0.23 | 0.23 | 1 | 28 | 28 | 2.0 | 3.6 | 3.6 | 2.0 | 0.0 | 2.0 | 20 | 45 | 45 | 5.0 | 0.0 | 5.0 | 10 | 364 | 364 | 1 | 140 | 140 |
| Event Mean Concentration: | | | | | | | | | | | 48.88 | | 6.80 | 4 | 1 | 4 | 0.5 | 0.2 | 0.6 | 0.05 | 0.92 | 0.92 | 0.01 | 0.46 | 0.46 | 1 | 16 | 16 | 2.0 | 5.4 | 5.4 | 2.0 | 0.0 | 2.0 | 20 | 59 | 59 | 5.0 | 0.0 | 5.0 | 10 | 79 | 79 | 1 | 148 | 148 |
| Versar | 1 | AC | 12/6/2016 | 1355 | 102 | I | S | 0.78 | 12.0 | 0.07 | 46.40 | 22690 | 6.9 | 4 | 0 | 4 | 0.5 | 0.0 | 0.5 | 0.05 | 0.76 | 0.76 | 0.01 | 0.08 | 0.08 | 1 | 12 | 12 | 2.0 | 5.9 | 5.9 | 2.0 | 0.0 | 2.0 | 20 | 49 | 49 | 5.0 | 0.0 | 5.0 | 100 | 0 | 100 | 1 | 86 | 86 |
| Versar | 2 | AC | 12/6/2016 | 1605 | 102 | I | S | | | | 45.68 | 207310 | 7.0 | 4 | 8 | 8 | 0.5 | 0.6 | 0.6 | 0.05 | 0.26 | 0.26 | 0.01 | 0.14 | 0.14 | 1 | 34 | 34 | 2.0 | 10.1 | 10.1 | 2.0 | 3.9 | 3.9 | 20 | 65 | 65 | 5.0 | 0.0 | 5.0 | 100 | 4650 | 4650 | 1 | 33 | 33 |
| Versar | 3 | AC | 12/6/2016 | 2040 | 102 | I | S | | | | 45.68 | 525576 | 7.0 | 4 | 0 | 4 | 0.5 | 1.0 | 1.0 | 0.05 | 0.26 | 0.26 | 0.01 | 0.07 | 0.07 | 1 | 11 | 11 | 2.0 | 5.3 | 5.3 | 2.0 | 0.0 | 2.0 | 20 | 43 | 43 | 5.0 | 0.0 | 5.0 | 100 | 1560 | 1560 | 1 | 33 | 33 |
| Event Mean Concentration: | | | | | | | | | | | 45.70 | | 7.00 | 4 | 2 | 5 | 0.5 | 0.9 | 0.9 | 0.05 | 0.28 | 0.28 | 0.01 | 0.09 | 0.09 | 1 | 17 | 17 | 2.0 | 6.6 | 6.6 | 2.0 | 1.1 | 2.5 | 20 | 49 | 49 | 5.0 | 0.0 | 5.0 | 100 | 2361 | 2364 | 1 | 35 | 35 |
| Versar | 1 | AC | 1/23/2017 | 1155 | 102 | I | S | 0.49 | 12.0 | 0.04 | NA | 37702 | 7 | 2 | 3 | 3 | 0.5 | 1.3 | 1.3 | 0.05 | 0.56 | 0.56 | 0.01 | 0.12 | 0.12 | 1 | 34 | 34 | 2.0 | 13.0 | 13.0 | 2.0 | 2.9 | 2.9 | 20 | 104 | 104 | 5.0 | 0.0 | 5.0 | 10 | 3448 | 3448 | 1 | 94 | 94 |
| Versar | 2 | AC | 1/23/2017 | 1440 | 102 | I | S | | | | NA | 280072 | 7.1 | 2 | 3 | 3 | 0.5 | 0.9 | 0.9 | 0.05 | 0.18 | 0.18 | 0.01 | 0.28 | 0.28 | 1 | 88 | 88 | 2.0 | 58.4 | 58.4 | 2.0 | 27.8 | 27.8 | 20 | 361 | 361 | 5.0 | 0.0 | 5.0 | 10 | 4352 | 4352 | 1 | 100 | 100 |
| Versar | 3 | AC | 1/23/2017 | 1615 | 102 | I | S | | | | NA | 221801 | 7.1 | 2 | 0 | 2 | 0.5 | 0.9 | 0.9 | 0.05 | 0.20 | 0.20 | 0.01 | 0.12 | 0.12 | 1 | 33 | 33 | 2.0 | 7.4 | 7.4 | 2.0 | 2.5 | 13.0 | 20 | 63 | 63 | 5.0 | 0.0 | 5.0 | 10 | 5172 | 5172 | 1 | 48 | 48 |
| Event Mean Concentration: | | | | | | | | | | | NA | | 7.09 | 2 | 2 | 3 | 0.5 | 0.9 | 0.9 | 0.05 | 0.21 | 0.21 | 0.01 | 0.20 | 0.20 | 1 | 62 | 62 | 2.0 | 34.3 | 34.3 | 2.0 | 15.7 | 15.7 | 20 | 221 | 221 | 5.0 | 0.0 | 5.0 | 10 | 4626 | 4626 | 1 | 78 | 78 |
| Versar | 1 | AC | 2/28/2017 | 1915 | 102 | I | S | 0.33 | 9.0 | 0.04 | 54.5 | 10747.60 | 7.1 | 6 | 0 | 6 | 0.5 | 0.9 | 0.9 | 0.05 | 0.78 | 0.78 | 0.01 | 0.14 | 0.14 | 1 | 43 | 43 | 2.0 | 22.9 | 22.9 | 2.0 | 2.5 | 2.5 | 20 | 117 | 117 | 5.0 | 6.0 | 6.0 | 1 | 121 | 121 | 1 | 91 | 91 |
| Versar | 2 | AC | 2/28/2017 | 2030 | 102 | I | S | | | | 55.76 | 86078.26 | 7.2 | 2 | 5 | 5 | 0.5 | 1.0 | 1.0 | 0.05 | 0.44 | 0.44 | 0.01 | 0.14 | 0.14 | 1 | 41 | 41 | 2.0 | 16.8 | 16.8 | 2.0 | 3.8 | 3.8 | 20 | 91 | 91 | 5.0 | 0.0 | 5.0 | 1 | 3130 | 3130 | 1 | 47 | 47 |
| Versar | 3 | AC | 2/28/2017 | 2255 | 102 | I | S | | | | 55.58 | 151646.57 | 7.3 | 2 | 3 | 3 | 0.5 | 0.0 | 0.5 | 0.05 | 0.32 | 0.32 | 0.01 | 0.08 | 0.08 | 1 | 13 | 13 | 2.0 | 7.5 | 7.5 | 2.0 | 0.0 | 2.0 | 20 | 49 | 49 | 5.0 | 0.0 | 5.0 | 1 | 2160 | 2160 | 1 | 38 | 38 |
| Event Mean Concentration: | | | | | | | | | | | 55.60 | | 7.26 | 2 | 4 | 4 | 0.5 | 0.4 | 0.7 | 0.05 | 0.38 | 0.38 | 0.01 | 0.10 | 0.10 | 1 | 24 | 24 | 2.0 | 11.4 | 11.4 | 2.0 | 1.4 | 2.6 | 20 | 67 | 67 | 5.0 | 0.3 | 5.0 | 1 | 2408 | 2408 | 1 | 43 | 43 |
| Versar | 1 | AC | 3/31/2017 | 735 | 102 | I | S | 1.12 | 10.0 | 0.11 | 49.82 | 23443 | 7.2 | 2 | 23 | 23 | 0.5 | 2.1 | 2.1 | 0.05 | 1.60 | 1.60 | 0.01 | 0.20 | 0.20 | 1 | 23 | 23 | 2.0 | 20.3 | 20.3 | 2.0 | 0.0 | 2.0 | 20 | 82 | 82 | 5.0 | 0.0 | 5.0 | 1 | 2920 | 2920 | 1 | 98 | 98 |
| Versar | 2 | AC | 3/31/2017 | 1155 | 102 | I | S | | | | 50.00 | 471171 | 7.5 | 2 | 2 | 2 | 0.5 | 0.8 | 0.8 | 0.05 | 0.32 | 0.32 | 0.01 | 0.17 | 0.17 | 1 | 54 | 54 | 2.0 | 19.0 | 19.0 | 2.0 | 7.1 | 7.1 | 20 | 77 | 77 | 5.0 | | | | | | | | |

| Sampler | ID | Jurisdiction | Date | Time | Site | Outfall or Instream | Storm or Baseflow | Depth | Duration | Intensity | Temperature - field | Flow | pH - field | dt for BOD | BOD | BOD | dt for Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | Total Kjeldahl Nitrogen | dt for Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | Nitrate+ Nitrite - N | dt for Total Phosphorus | Total Phosphorus | Total Phosphorus | dt for TSS | TSS | TSS | dt for Copper | Copper | Copper | dt for Lead | Lead | Lead | dt for Zinc | Zinc | Zinc | dt for TPH | TPH | TPH | dt for E-COLI | E-COLI | E-COLI | dt for HARDNESS | HARDNESS | HARDNESS |
|---|----|--------------|-----------|------|------|---------------------|-------------------|-------|----------|-----------|---------------------|--------|-------------|------------|----------|----------|--------------------------------|-------------------------|-------------------------|-----------------------------|----------------------|----------------------|-------------------------|------------------|------------------|------------|-----------|-----------|---------------|-------------|-------------|-------------|-------------|-------------|-------------|------------|------------|------------|------------|------------|---------------|--------------|--------------|-----------------|-----------|-----------|
| Versar | 1 | AC | 4/25/2017 | 1240 | 102 | I | S | 0.39 | 5.0 | 0.08 | 58.46 | 8086 | 7 | 2 | 15 | 15 | 0.5 | 2.6 | 2.6 | 0.05 | 0.37 | 0.37 | 0.01 | 0.48 | 0.48 | 1 | 210 | 210 | 2.0 | 45.8 | 45.8 | 2.0 | 11.1 | 11.1 | 20 | 243 | 243 | 5.0 | 5.0 | 5.0 | 1 | 1553 | 1553 | 1 | 79 | 79 |
| Versar | 2 | AC | 4/25/2017 | 1320 | 102 | I | S | | | | 59.36 | 130027 | 7 | 2 | 5 | 5 | 0.5 | 1.3 | 1.3 | 0.05 | 0.19 | 0.19 | 0.01 | 0.24 | 0.24 | 1 | 88 | 88 | 2.0 | 21.0 | 21.0 | 2.0 | 6.4 | 6.4 | 20 | 84 | 84 | 5.0 | 0.0 | 5.0 | 1 | 2420 | 2420 | 1 | 35 | 35 |
| Versar | 3 | AC | 4/25/2017 | 1605 | 102 | I | S | | | | 60.08 | 276979 | 7.2 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.22 | 0.22 | 0.01 | 0.09 | 0.09 | 1 | 15 | 15 | 2.0 | 16.4 | 16.4 | 2.0 | 0.0 | 2.0 | 20 | 38 | 38 | 5.0 | 0.0 | 5.0 | 1 | 2420 | 2420 | 1 | 36 | 36 |
| Event Mean Concentration: | | | | | | | | | | | 59.82 | | 7.13 | 2 | 2 | 3 | 0.5 | 0.5 | 0.8 | 0.05 | 0.21 | 0.21 | 0.01 | 0.14 | 0.14 | 1 | 42 | 42 | 2.0 | 18.4 | 18.4 | 2.0 | 2.2 | 3.6 | 20 | 56 | 56 | 5.0 | 0.1 | 5.0 | | ** | ** | 1 | 37 | 37 |
| Versar | 1 | AC | 5/12/2017 | 2255 | 102 | I | S | 0.98 | 21.0 | 0.05 | 57.38 | 92661 | 7.1 | 2 | 0 | 2 | 0.5 | 0.7 | 0.7 | 0.05 | 0.29 | 0.29 | 0.01 | 0.05 | 0.05 | 1 | 11 | 11 | 2.0 | 7.0 | 7.0 | 2.0 | 0.0 | 2.0 | 20 | 40 | 40 | 5.0 | 7.0 | 7.0 | 1 | 909 | 909 | 1 | 24 | 24 |
| Versar | 2 | AC | 5/13/2017 | 415 | 102 | I | S | | | | 54.68 | 420994 | 7.2 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.14 | 0.14 | 0.01 | 0.09 | 0.09 | 1 | 24 | 24 | 2.0 | 6.8 | 6.8 | 2.0 | 2.6 | 2.6 | 20 | 44 | 44 | 5.0 | 0.0 | 5.0 | 1 | 2613 | 2613 | 1 | 42 | 42 |
| Versar | 3 | AC | 5/13/2017 | 1010 | 102 | I | S | | | | 55.40 | 585240 | 7.2 | 2 | 0 | 2 | 0.5 | 0.0 | 0.5 | 0.05 | 0.16 | 0.16 | 0.01 | 0.07 | 0.07 | 1 | 9 | 9 | 2.0 | 5.1 | 5.1 | 2.0 | 0.0 | 2.0 | 20 | 35 | 35 | 5.0 | 0.0 | 5.0 | 1 | 4352 | 4352 | 1 | 27 | 27 |
| Event Mean Concentration: | | | | | | | | | | | 55.29 | | 7.19 | 2 | 0 | 2 | 0.5 | 0.1 | 0.5 | 0.05 | 0.16 | 0.16 | 0.01 | 0.08 | 0.08 | 1 | 15 | 15 | 2.0 | 5.9 | 5.9 | 2.0 | 1.0 | 2.2 | 20 | 39 | 39 | 5.0 | 0.6 | 5.2 | 1 | 3395 | 3395 | 1 | 32 | 32 |
| Versar | 1 | AC | 6/19/2017 | 1635 | 102 | I | S | 0.36 | 4.0 | 0.09 | 79.16 | 7734 | 6.9 | 2 | 7 | 7 | 0.5 | 0.6 | 0.6 | 0.05 | 0.91 | 0.91 | 0.01 | 0.27 | 0.27 | 1 | 43 | 43 | 2.0 | 408.0 | 408.0 | 2.0 | 401.0 | 401.0 | 20 | 3600 | 3600 | 5.0 | 12.0 | 12.0 | 1 | 22470 | 22470 | 1 | 230 | 230 |
| Versar | 2 | AC | 6/19/2017 | 1740 | 102 | I | S | | | | 79.34 | 177693 | 7.1 | 2 | 12 | 12 | 0.5 | 1.3 | 1.3 | 0.05 | 0.59 | 0.59 | 0.01 | 0.22 | 0.22 | 1 | 94 | 94 | 2.0 | 23.0 | 23.0 | 2.0 | 9.0 | 9.0 | 20 | 125 | 125 | 5.0 | 0.0 | 5.0 | 1 | 26130 | 26130 | 1 | 42 | 42 |
| Versar | 3 | AC | 6/19/2017 | 1915 | 102 | I | S | | | | 79.16 | 176785 | 6.9 | 2 | 4 | 4 | 0.5 | 0.0 | 0.5 | 0.05 | 0.50 | 0.50 | 0.01 | 0.12 | 0.12 | 1 | 25 | 25 | 2.0 | 9.0 | 9.0 | 2.0 | 2.0 | 2.0 | 20 | 51 | 51 | 5.0 | 0.0 | 5.0 | 1 | 13540 | 13540 | 1 | 39 | 39 |
| Event Mean Concentration: | | | | | | | | | | | 79.25 | | 7.00 | 2 | 8 | 8 | 0.5 | 0.7 | 0.9 | 0.05 | 0.55 | 0.55 | 0.01 | 0.17 | 0.17 | 1 | 59 | 59 | 2.0 | 24.4 | 24.4 | 2.0 | 14.0 | 14.0 | 20 | 163 | 163 | 5.0 | 0.3 | 5.1 | 1 | 19907 | 19907 | 1 | 45 | 45 |
| ** Samples were not diluted - the data has been discarded and not included in subsequent flow weighted EMC and loading calculations | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |
| *** Error in results - the data has been discarded and not included in subsequent flow weighted EMC and loading calculations | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | | |

**Anne Arundel County NPDES
Sampling and EMC Data – 2017 Reporting Year
Church Creek Station**

| | Temperature - field pF | pH - field pH | BOD (0) mg/L | BOD (dt) mg/L | Total Kjeldahl Nitrogen (0) mg/L | Total Kjeldahl Nitrogen (dt) mg/L | Nitrate+ Nitrite - N (0) mg/L | Nitrate+ Nitrite - N (dt) mg/L | Total Phosphorus (0) mg/L | Total Phosphorus (dt) mg/L | TSS (0) mg/L | TSS (dt) mg/L | Copper (0) µg/L | Copper (dt) µg/L | Lead (0) µg/L | Lead (dt) µg/L | Zinc (0) µg/L | Zinc (dt) µg/L | TPH (0) mg/L | TPH (dt) mg/L | E-COLI (0) MPN | E-COLI (dt) MPN | HARDNESS (0) mg/L | HARDNESS (dt) mg/L |
|---|---------------------------|------------------|-----------------|------------------|-------------------------------------|--------------------------------------|----------------------------------|-----------------------------------|------------------------------|-------------------------------|-----------------|------------------|--------------------|---------------------|------------------|-------------------|------------------|-------------------|-----------------|------------------|-------------------|--------------------|----------------------|-----------------------|
| Summer Quarter Flow-Weighted EMC (7/28/16, 9/19/16, 9/27/16): | 74.85 | 6.62 | 4.87 | 7.32 | 0.92 | 0.92 | 0.39 | 0.39 | 0.17 | 0.17 | 58.02 | 58.02 | 9.27 | 9.27 | 7.00 | 9.19 | 74.15 | 74.15 | 0.00 | 5.00 | 10021.38 | 10023.06 | 42.74 | 42.74 |
| Average: | | | 6.10 | mg/l | 0.92 | mg/l | 0.39 | mg/l | 0.17 | mg/l | 58.02 | mg/l | 9.27 | µg/l | 8.09 | µg/l | 74.15 | µg/l | 2.50 | mg/l | 10022.22 | MPN/100mL | 42.74 | mg/l |
| | | | 0.0003805 | lb/cf | 0.0000574 | lb/cf | 0.0000242 | lb/cf | 0.0000103 | lb/cf | 0.0036211 | lb/cf | 0.0000006 | lb/cf | 0.0000005 | lb/cf | 0.0000046 | lb/cf | 0.0001560 | lb/cf | | | 0.0026678 | lb/cf |
| Total Volume (Quarter Events): | | | 1,339,485 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 509.6 | lbs | 76.9 | lbs | 32.4 | lbs | 13.8 | lbs | 4,850.4 | lbs | 0.8 | lbs | 0.7 | lbs | 6.2 | lbs | 209.0 | lbs | | | 3,573.5 | lbs |
| Total Volume (Quarter): | | | 28,590,186 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 10,877.5 | lbs | 1,641.4 | lbs | 691.0 | lbs | 294.6 | lbs | 103,527.3 | lbs | 16.5 | lbs | 14.4 | lbs | 132.3 | lbs | 4,461.2 | lbs | | | 76,273.6 | lbs |
| Fall Quarter Flow-Weighted EMC (11/9/16, 11/29/16, 12/6/17): | 46.62 | 6.93 | 2.72 | 5.42 | 0.92 | 0.97 | 0.43 | 0.43 | 0.15 | 0.15 | 20.74 | 20.74 | 6.60 | 6.60 | 0.97 | 2.47 | 53.37 | 53.37 | 0.00 | 5.00 | 2222.45 | 2224.97 | 51.06 | 51.06 |
| Average: | | | 4.07 | mg/l | 0.95 | mg/l | 0.43 | mg/l | 0.15 | mg/l | 20.74 | mg/l | 6.60 | µg/l | 1.72 | µg/l | 53.37 | µg/l | 2.50 | mg/l | 2223.71 | MPN/100mL | 51.06 | mg/l |
| | | | 0.0002539 | lb/cf | 0.0000590 | lb/cf | 0.0000266 | lb/cf | 0.0000093 | lb/cf | 0.0012946 | lb/cf | 0.0000004 | lb/cf | 0.0000001 | lb/cf | 0.0000033 | lb/cf | 0.0001560 | lb/cf | | | 0.0031871 | lb/cf |
| Total Volume (Quarter Events): | | | 899,444 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 228.3 | lbs | 53.1 | lbs | 23.9 | lbs | 8.4 | lbs | 1,164.4 | lbs | 0.4 | lbs | 0.1 | lbs | 3.0 | lbs | 140.3 | lbs | | | 2,866.6 | lbs |
| Total Volume (Quarter): | | | 19,349,299 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 4,912.4 | lbs | 1,142.0 | lbs | 514.0 | lbs | 180.7 | lbs | 25,048.8 | lbs | 8.0 | lbs | 2.1 | lbs | 64.5 | lbs | 3,019.3 | lbs | | | 61,668.6 | lbs |
| Winter Quarter Flow-Weighted EMC (1/23/17, 2/28/17, 3/31/17): | 37.38 | 7.28 | 2.39 | 2.65 | 0.83 | 0.87 | 0.32 | 0.32 | 0.15 | 0.15 | 39.66 | 39.66 | 19.63 | 19.63 | 6.97 | 8.34 | 118.27 | 118.27 | 0.03 | 5.01 | 2282.32 | 2282.32 | 45.29 | 45.29 |
| Average: | | | 2.52 | mg/l | 0.85 | mg/l | 0.32 | mg/l | 0.15 | mg/l | 39.66 | mg/l | 19.63 | µg/l | 7.66 | µg/l | 118.27 | µg/l | 2.52 | mg/l | 2282.32 | MPN/100mL | 45.29 | mg/l |
| | | | 0.0001572 | lb/cf | 0.0000532 | lb/cf | 0.0000198 | lb/cf | 0.0000092 | lb/cf | 0.0024757 | lb/cf | 0.0000012 | lb/cf | 0.0000005 | lb/cf | 0.0000074 | lb/cf | 0.0001572 | lb/cf | | | 0.0028267 | lb/cf |
| Total Volume (Quarter Events): | | | 1,943,537 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 305.6 | lbs | 103.3 | lbs | 38.5 | lbs | 17.9 | lbs | 4,811.6 | lbs | 2.4 | lbs | 0.9 | lbs | 14.3 | lbs | 305.6 | lbs | | | 5,493.8 | lbs |
| Total Volume (Quarter): | | | 22,568,474 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 3,548.5 | lbs | 1,199.7 | lbs | 447.2 | lbs | 207.4 | lbs | 55,872.1 | lbs | 27.7 | lbs | 10.8 | lbs | 166.6 | lbs | 3,548.8 | lbs | | | 63,794.1 | lbs |
| Spring Quarter Flow-Weighted EMC (4/25/17, 5/12/17, 6/19/17): | 60.92 | 7.14 | 1.95 | 3.42 | 0.26 | 0.65 | 0.25 | 0.25 | 0.11 | 0.11 | 29.39 | 29.39 | 12.23 | 12.23 | 3.76 | 4.78 | 66.70 | 66.70 | 0.42 | 5.13 | 7488.72 | 7488.72 | 35.71 | 35.71 |
| Average: | | | 2.69 | mg/l | 0.46 | mg/l | 0.25 | mg/l | 0.11 | mg/l | 29.39 | mg/l | 12.23 | µg/l | 4.27 | µg/l | 66.70 | µg/l | 2.77 | mg/l | 7488.72 | MPN/100mL | 35.71 | mg/l |
| | | | 0.0001677 | lb/cf | 0.0000285 | lb/cf | 0.0000156 | lb/cf | 0.0000068 | lb/cf | 0.0018344 | lb/cf | 0.0000008 | lb/cf | 0.0000003 | lb/cf | 0.0000042 | lb/cf | 0.0001730 | lb/cf | | | 0.0022291 | lb/cf |
| Total Volume (Quarter Events): | | | 1,876,198 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter Events): | | | 314.6 | lbs | 53.4 | lbs | 29.2 | lbs | 12.9 | lbs | 3,441.7 | lbs | 1.4 | lbs | 0.5 | lbs | 7.8 | lbs | 324.6 | lbs | | | 4,182.1 | lbs |
| Total Volume (Quarter): | | | 25,158,577.5 | cf | | | | | | | | | | | | | | | | | | | | |
| Pollutant Load (Quarter): | | | 4,219.0 | lbs | 716.1 | lbs | 392.0 | lbs | 172.3 | lbs | 46,150.4 | lbs | 19.2 | lbs | 6.7 | lbs | 104.7 | lbs | 4,353.1 | lbs | | | 56,079.8 | lbs |
| AVERAGE ANNUAL EMCs: | 54.33 | 7.04 | 2.85 | mg/l | 0.69 | mg/l | 0.33 | mg/l | 0.14 | mg/l | 37.73 | mg/l | 13.12 | µg/l | 5.09 | µg/l | 82.91 | µg/l | 0.14 | mg/l | 5596.70 | mg/l | 42.62 | mg/l |
| TOTAL ANNUAL POLLUTANT LOAD (EVENTS): | | | 1,358.19 | lbs | 286.71 | lbs | 124.01 | lbs | 52.91 | lbs | 14,267.97 | lbs | 4.96 | lbs | 2.20 | lbs | 31.35 | lbs | 979.61 | lbs | | | 16,116.07 | lbs |
| Per Acre: | | | 4.87 | | 1.03 | | 0.44 | | 0.19 | | 51.12 | | 0.02 | | 0.01 | | 0.11 | | 3.51 | | | | 57.75 | |
| TOTAL 2017 POLLUTANT LOAD: | | | 23,557.41 | lbs | 4,699.28 | lbs | 2,044.22 | lbs | 854.98 | lbs | 230,598.58 | lbs | 71.38 | lbs | 34.01 | lbs | 468.11 | lbs | 15,382.44 | lbs | | | 257,816.15 | lbs |

APPENDIX H
BMP CODES

MDE Approved BMP Classifications

| ESD BMPs | | |
|-------------------------------------|------|-------------------------------------|
| Category | Code | Code Description |
| Alternative Surfaces (A) | | |
| E | AGRE | Green Roof - Extensive |
| E | AGRI | Green Roof - Intensive |
| E | APRP | Permeable Pavements |
| E | ARTF | Reinforced Turf |
| Nonstructural Techniques (N) | | |
| E | NDRR | Disconnection of Rooftop Runoff |
| E | NDNR | Disconnection of Non-Rooftop Runoff |
| E | NSCA | Sheetflow to Conservation Areas |
| Micro-Scale Practices (M) | | |
| E | MRWH | Rainwater Harvesting |
| E | MSGW | Submerged Gravel Wetlands |
| E | MILS | Landscape Infiltration |
| E | MIBR | Infiltration Berms |
| E | MIDW | Dry Wells |
| E | MMBR | Micro-Bioretenention |
| E | MRNG | Rain Gardens |
| E | MSWG | Grass Swale |
| E | MSWW | Wet Swale |
| E | MSWB | Bio-Swale |
| E | MENF | Enhanced Filters |
| Structural BMPs | | |
| Ponds (P) | | |
| S | PWED | Extended Detention Structure, Wet |
| S | PWET | Retention Pond (Wet Pond) |
| S | PMPS | Mutliple Pond System |
| S | PPKT | Pocket Pond |
| S | PMED | Micropool Extended Detention Pond |
| Wetlands (W) | | |
| S | WSHW | Shallow Marsh |
| S | WEDW | ED - Wetland |
| S | WPWS | Wet Pond - Wetland |
| S | WPKT | Pocket Wetland |
| Infiltration (I) | | |
| S | IBAS | Infiltration Basin |
| S | ITRN | Infiltration Trench |

| Filtering Systems (F) | | |
|-----------------------|------|-----------------------------------|
| S | FBIO | Bioretention |
| S | FSND | Sand Filter |
| S | FUND | Underground Filter |
| S | FPER | Perimeter (Sand) Filter |
| S | FORG | Organic Filter (Peat Filter) |
| S | FBIO | Bioretention |
| Open Channels (O) | | |
| S | ODSW | Dry Swale |
| S | OWSW | Wet Swale |
| Other Practices (X) | | |
| S | XDPD | Detention Structure (Dry Pond) |
| S | XDED | Extended Detention Structure, Dry |
| S | XFLD | Flood Management Area |
| S | XOGS | Oil Grit Separator |
| S | XOTH | Other |

MDE Approved Alternative BMP Classifications

| Alt. BMPs (A) | Code | Code Description |
|---------------|------|---|
| A | MSS | Mechanical Street Sweeping |
| A | VSS | Regenerative/Vacuum Street Sweeping |
| A | IMPP | Impervious Surface Elimination (to pervious) |
| A | IMPF | Impervious Surface Elimination (to forest) |
| A | FPU | Planting Trees or Forestation on Pervious Urban |
| A | CBC | Catch Basin Cleaning |
| A | SDV | Storm Drain Vacuuming |
| A | STRE | Stream Restoration |
| A | OUT | Outfall Stabilization |
| A | SPSC | Regenerative Step Pool Storm Conveyance |
| A | SHST | Shoreline Management |
| A | SEPP | Septic Pumping |
| A | SEPD | Septic Denitrification |
| A | SEPC | Septic Connections to WWTP |

APPENDIX I
BMP ACRONYMS

MDE-APPROVED BMP TYPE ACRONYMS

BMP—Best Management Practice

APRP—Permeable Pavements

FBIO—Bioretention

FSND—Sand Filter

IBAS—Infiltration Basin

IMPP—Impervious Surface Elimination (to pervious)

ITRN—Infiltration Trench

MMBR—Micro-Bioretention

MSGW—Submerged Gravel Wetlands

MSWG—Grass Swale

MSWW—Wet Swale

NDRR—Disconnection of Rooftop Runoff

ODSW—Dry Swale

OWSW—Wet Swale

PWED—Extended Detention Structure, Wet

XDED—Extended Detention Structure, Dry

XDPD—Detention Structure (Dry Pond)

XOGS—Oil Grit Separator

